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Predicting Health Impacts of Wildfire Smoke in Amazonas basin, Brazil

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Conflict of Interest

The authors declare they have no conflicts of interest related to this work to disclose.

1 **Abstract**

2 Worldwide, smoke from forest fires has deleterious health effects. Even so, because of the
3 complexity of fire mechanics, public health authorities face challenges in forecasting and thus
4 mitigating population exposure to smoke. The population in the Amazon basin regularly suffers
5 from fire smoke tied to agriculture and land-use change. The people of Manaus, a city of two
6 million in the center of the basin, suffer the consequences. The study herein evaluates the time
7 lag between fire occurrence and hospital admission for cardiorespiratory illness. Understanding
8 the time lag is key to forecasting and mitigating the public health effects. The study approach is
9 sequential application of four increasingly complex methods of machine learning to examine the
10 relationships among black carbon concentrations, fire count, meteorology, and hospital
11 admissions. The mean absolute percentage error (MAPE) for predicting hospital admissions
12 ranged from 27% to 38%. Furthermore, a one-day lag was observed between the detection of
13 fires and the manifestations of respiratory health hazards. This finding suggests the potential for
14 developing an early warning system, which could enable public health officials to issue
15 advisories or implement preventive actions during the brief period before hospital admissions
16 begin to rise. The findings have applicability not only to the population exposed to fires in the
17 Amazon basin but also to populations where smoke is prevalent, notably increasingly in
18 Australia, southern Europe, the western USA, southern Canada, and southeast Asia.

19 **Keywords:** fire smoke, Amazon basin, cardiorespiratory illness, forecast and warning

20 **Synopsis:** A forewarning artificial neural network is developed for exposure to wildfire smoke at
21 the population level. Use of this approach can reduce respiratory and cardiovascular
22 hospitalizations.

23 **1. Introduction**

24 Forest fires are a global concern. Consequences include severe air pollution episodes,
25 human mortality, environmental damage, and substantial economic loss¹. Human activities and
26 climate change have led to heightened intensity, frequency, and duration of fire seasons.
27 Approximately 200,000 forest fires are reported annually.¹ Notably, the summer of 2022
28 witnessed record-breaking wildfire activity in the European Union and the United Kingdom,
29 surpassing the previous 15-year record.² Similarly, California experienced a record number of
30 wildfires in 2020, more than double the previous record.³ The severity of the 2019/2020 fire
31 season in Australia was unprecedented. Over 23% of the temperate forest in southeastern
32 Australia was affected.⁴

33 The Amazon forest also faced an alarming increase in deforestation fires in 2019. At that
34 time, the Brazilian government reversed commitments to control deforestation.⁵ The state of
35 Amazonas recording its highest fire count since 1998 in 2022.⁶ The impact on the Amazon
36 ecosystem, known for its biodiversity and vast freshwater, is severe. However, studies on the
37 effects of wildfire smoke in this region are limited due to the scarcity of air quality monitoring
38 stations in the northern Amazon.⁷ Forest fires contribute significantly to black carbon (BC)
39 emissions. Black carbon in turn is part of airborne particulate matter (PM_{2.5}).^{8,9}

40 Previous exposure assessment studies usually use PM concentration as a proxy for
41 wildfire smoke.¹⁰ And, there were few epidemiological studies of fire health effects prior to the
42 last decade because of fire occurrence far from populated areas where air pollution levels were
43 seldomly monitored.¹¹ More recently, Johnston et al.¹² report that the overall premature mortality
44 rate that can be attributed to wildfire smoke is 339,000 individuals globally each year. Nawaz
45 and Henze¹³ found that Brazilian biomass burning emissions (mostly in Amazonia) accounted for

46 a 74% increase in premature deaths. Alves et al.¹⁴ further demonstrated that biomass burning in
47 the Amazon region leads to DNA damage and cell death in human lung cells. Recently, Prist et
48 al.¹⁵ estimated that 500 cardiorespiratory infections per 100,000 inhabitants were related to forest
49 fires in the Amazon. Further studies in the Amazon region confirm a positive correlation
50 between wildfire smoke and an increase in the incidence of morbidity and mortality among
51 vulnerable populations, including children and the elderly^{7, 13-15}.

52 There are several methods widely used to assess exposure to wildfire smoke.¹⁶ The fires
53 and non-fire days can be compared, the concentration of PM can be monitored or modeled,
54 smoke indicators including counts and burned area from satellite observations can be used, and
55 combination methods can integrate one or more of these approaches. However, forewarning
56 predictions of health risks based on fire episodes are scarce. If such forewarning were possible,
57 the public, especially vulnerable populations, could take action to avoid smoke exposure.
58 Similarly, healthcare professionals, hospital systems, and health insurers could integrate potential
59 health impacts into day-to-day actionable planning.¹¹

60 The present study, focusing on the impact of wildfire smoke on the health of the general
61 population in the central Amazon, employs machine learning to examine relationships among
62 black carbon concentrations, fire count, meteorology, and hospital admissions for
63 cardiorespiratory illness. The use of artificial neural networks (ANNs) as machine-learning
64 forecasting models could provide elegant and robust solutions for non-linear relationships among
65 multiple variables and discontinuous datasets.¹⁷⁻²² The approach herein not only provides insights
66 into the dynamics in the central Amazon but also contributes significantly to the global discourse
67 on forest fires and their health implications.

68

70 2. Materials and Methods

71 2.1 *Sampling site*

72 Manaus is a metropolitan area located in the central Amazon with a population of 2.3
73 million in 2021.²³ In a subtropical monsoon climate, the average annual temperature is 27 °C,
74 and the average relative humidity is 80%.²⁴ The wet season lasts from November to May, and the
75 dry season takes place from June to October. There is intermittent intrusion of regional and
76 continental scale wildfire smoke, primarily during the dry season.²⁵ The severe, episodic
77 pollution strongly affects public health and hospital admissions. For this study, sampling
78 campaigns took place from 2011-2013 (3° 5'43.94"S, 59°59'25.56"W) and 2015-2016 (3°
79 6'12.5"S, 59°58'55.8"W) in a central area of Manaus (Figure 1). The obtained dataset of
80 particulate matter (PM_{2.5}) and black carbon (BC) had 785 samples collected over four years.
81 Corresponding meteorological data were obtained from the Brazilian National Institute of
82 Meteorology (INMET).

83 2.2 *PM_{2.5}, BC, and fire counts.*

84 PM_{2.5} was collected (24-hour sampling) from Oct 2011 to July 2013 and from Aug 2015
85 to Aug 2016 using a low-volume Harvard impactor and 37-mm polycarbonate filters. PM_{2.5} mass
86 concentrations were determined gravimetrically following the same procedure by Polezer et al.¹⁷,
87 positioning the impactor 2 m height and using blank filters to track and reduce errors due to filter
88 handling and transport. The BC fraction of the sampled PM_{2.5} was determined through
89 transmittance at an 880 nm wavelength (infrared) (Sootscan optical transmissometer, model OT
90 21, Magee Scientific Company). The BC concentration and the daily fire count were used as
91 proxy variables for wildfire smoke.

92 The variability of fire data around Manaus up to 500 kilometers is quite the same, as
93 shown in Figure S1. In this sense, we chose to use the number of fires within 200 kilometers
94 around Manaus, 10 times the city's average radius. The fire data was obtained from the National
95 Institute for Space Research²⁷. The data is from the reference satellite (AQUA, which uses a
96 MODIS sensor)²⁸.

97 *2.3 Hospital admissions*

98 The health impacts accompanying wildfire smoke were evaluated using the Manaus
99 hospital admission count for cardiorespiratory illness. Cardiorespiratory illness is widespread
100 following exposure to fire smoke.²⁹ Hospital admission data across Oct 2011 to Aug 2016 were
101 obtained from the Brazilian Unified Health System for respiratory diseases (RD) (International
102 Classification of Diseases – ICD-10, codes J00 to J99) and cardiovascular diseases (CVD) (ICD-
103 10, codes I00 to I99).³⁰

104 *2.4 Machine learning*

105 Machine learning was applied to examine the relationships among black carbon
106 concentrations, fire count, meteorology, and hospital admissions. Explanatory variables (i.e.,
107 input variables) included daily BC concentration, mean temperature, mean relative humidity,
108 precipitation, solar intensity, and the fire count. When dealing with air pollution epidemiological
109 studies, it is common to observe a relation between air pollution concentration some days ahead
110 of health outcomes, then it is crucial to consider a seven-day window when dealing with air
111 pollution health impacts, as suggested by ^{17, 18, 31-33}. Therefore, this study examined data from
112 zero (lag 0) to seven-day lag (lag 7) after exposure to forest fires and BC concentrations.

113 Machine learning by artificial neural networks (ANN) performed the analysis. ANNs are
114 nonlinear methodologies used to solve problems such as nonlinear mapping, forecasting,

115 classification, and clustering, among others.³⁴ The ANN neurons are organized into layers,
116 commonly named input, hidden (intermediate), and output layers.³⁵ Four artificial neural
117 networks architectures were sequentially applied, including so-called extreme learning machine
118 (ELM), echo state network (ESN), multilayer perceptron (MLP) with one and two hidden layers
119 (MLP-1 and MLP-2, respectively), and radial basis function network (RBF). ESN is a recurrent
120 neural network, and the others are feed-forward neural networks.^{34,36} MLP and RBF are fully
121 trained methodologies because all weights are adjusted. Two unorganized machines (UM) were
122 also considered (ELM and ESN). UM tunes only the output layer, which confers a simple
123 implementation and low computational cost.

124 The dry and wet season differed significantly from one another in terms of fire count, PM
125 concentrations, and hospital admissions (Table 1). Therefore, the analyses were also conducted
126 with and without a *Z*-score, which is a de-seasonalization technique. The goal was to evaluate if
127 transformation into stationary dataset without seasonal components improves model
128 performance.³⁵⁻³⁷ The *Z*-score consists of subtracting the value of each sample from the mean
129 and dividing the result by the standard deviation.

130 Statistics of the dataset used in the ANN analysis are listed in Table 1. For machine
131 learning, the dataset was divided into three parts, including a training dataset (used to adjust the
132 free parameters of the neural models; 585 samples), a validation dataset (to avoid overtraining;
133 100 samples), and a test dataset (used to evaluate the performance of the proposed models; 100
134 samples). For each ANN, sixteen analyses (lag days \times *Z*-score) were carried out including all
135 inputs, excluding forest fire count, and excluding BC concentration for respiratory and
136 cardiovascular diseases, totaling 96 analyses per ANN.

137 The performance of each neural network was evaluated based on the root mean square
138 error between predicted and observed admissions in the test dataset. The mean absolute
139 percentage error (MAPE) was also calculated. The cost function the ANNs minimize is the
140 RMSE and, in the case in which different error metrics indicate distinct models as the best, the
141 one with the lowest RMSE should be assumed as the best one.^{18, 35, 38} MAPE indicates the
142 absolute model performance relative to the observations of hospital admissions. It is important to
143 highlight that error metrics from different datasets are not comparable.

144 The Friedman test was applied to assess if the error values were statistically different
145 from each other, meaning one ANN performed better than another. Details about the ANN
146 designs and performances are in the Supplementary Material.

147

148 **3. Results and discussion**

149 Different approaches can be used to assess fire health risks: using monitored PM during
150 fire events, PM data from chemical transport modeling, satellite smoke data (counts and/or
151 burned area), comparison between smoky versus non-smoky days, self-questionnaire
152 information, satellite data plus chemical transport modeling, and others.^{16, 29, 39}

153 Urrutia-Pereira et al.⁷, in reviewing biomass burning and human health in the Amazon
154 rain forest, point out that studies related the effects of forest fire smoke in this region are limited
155 due to a lack of air quality measurements in the northern region of Brazil. This is in accordance
156 with Bowman et al.⁵, who point out that historical records of fire activity, even for simple
157 metrics like area burned, are limited. To that end, a decision was made that the best available
158 proxy for the studied region to compare the impact of fires on health in the Brazilian Amazon
159 Forest is active fire hotspots as it captures the dynamics of fires over time, a conclusion affirmed

160 by Sant'Anna and Rocha⁴⁰. Then, the fire count within 200 km around Manaus during the study
161 period was used.²⁷ The dataset is plotted in Figure 2. The year 2014 is missing because no BC
162 sampling campaigns were carried out during that year, which prevented its inclusion in the
163 analysis.

164 During the study period, the daily BC concentrations ranged from 0.06 to 5.56 $\mu\text{g m}^{-3}$,
165 with annual means varying from 1.29 to 2.28 $\mu\text{g m}^{-3}$ (Table 1). On days of severe episodic
166 smoke, the BC concentration reached levels up to 2-3 times higher than the average levels
167 observed during periods of urban pollution, with a maximum recorded concentration of 5.6 μg
168 m^{-3} .

169 Given the critical role of BC in our study, it is essential to acknowledge the complexities
170 and challenges associated with its measurement. The mass absorption cross-section of BC, often
171 employed to estimate BC mass from optical measurements, can vary significantly across
172 different environments and conditions, leading to substantial uncertainty in quantification. This
173 variability is further exacerbated by factors such as lensing effects and the mixing state of
174 particles, particularly in scenarios involving biomass burning, as discussed in the works of White
175 et al.⁴¹ (2016), Bond and Bergstrom⁴² (2006), and Zhang et al.⁴³ (2023). Furthermore, changes
176 in particle size and morphology post-sampling can influence light absorption measurements,
177 thereby impacting the accuracy of BC mass estimates. These challenges underscore the
178 importance of careful interpretation of BC data in epidemiological studies like ours, where
179 accurate exposure assessment is crucial for understanding health impacts.

180 The $\text{PM}_{2.5}$ concentration ranged from 0.04 to 68.7 $\mu\text{g m}^{-3}$, with a daily average of 9.2 μg
181 m^{-3} and a standard deviation of 6.83 $\mu\text{g m}^{-3}$. For comparison, Table 2 lists values for studies in
182 the Amazon region from literature. For the 2000's, the literature observations agree with those of

183 this study, considering the standard deviation. However, Artaxo et al.¹¹ reported data for the
184 1990's, during which PM_{2.5} concentrations were 3 to 6 times higher. In Rondônia state, records
185 of 50-90 µg m⁻³ were observed during the dry season from 2002 to 2009⁴⁴. During fire episodes
186 from Aug to Sep 2020, average daily PM_{2.5} concentration were four to eleven times the National
187 Ambient Air Quality Standards (NAAQS) in major cities in California (USA), Washington
188 (USA), and Oregon (USA).⁴⁷ Notably, fire smoke is a global issue, and our finding will have
189 applicability not only to the population exposed to fires in Manaus, but also to populations where
190 smoke is prevalent, such as USA, southern Europe and others. Ahangar et al.⁴⁸ considered eight
191 cities along the South Coast Air Basin (USA) and analyzed the wildfire contribution to PM_{2.5} and
192 its carbon content for 2008 to 2016. The authors analyzed the reduction in PM_{2.5} and BC annual
193 averages when excluding the fire days. In the urban areas, the most significant difference was
194 7% for PM_{2.5} and BC in 2008. In remote regions, the differences were 4% for PM_{2.5} and 21% for
195 BC in 2016. For the present study, much larger reductions of 17% and 11% for PM_{2.5} and BC
196 were observed, respectively, in 2012.

197 Figure S2 shows the boxplot of BC percent of the total PM_{2.5} mass. The values ranged
198 from 0.8 to 95.4%, with an average value of 23.0% and a standard deviation of 12.4%. The BC
199 percent is above typical values of biomass burning in the Amazon rainforest⁷. Typically, 10 to
200 15% of PM_{2.5} composition from Amazonian wildfires is BC. The fire count on a single day
201 ranged from 0 to 253 throughout the 200 km region surrounding Manaus (Figure 2). The fire
202 count is at its highest during the dry season, with a record of 2,298 recorded in 2015.

203 As listed in Table 1, the fire count and BC concentrations do not correlate with
204 cardiorespiratory data. Figure S3 shows a dispersion diagram with Pearson correlation
205 coefficients. The relation between BC, forest fire, and cardiorespiratory diseases is not linear,

206 with Pearson correlation coefficients from -0.16 (between respiratory hospitalizations and the
207 fire count) to 0.12 (between cardiovascular hospitalizations and BC). For comparison, Andrade-
208 Filho et al.⁴⁴ reported Pearson correlation coefficients of -0.079 between respiratory
209 hospitalizations and fire count. The Spearman correlations are between -0.21 to 0.13 and Kendall
210 correlations are between -0.16 to 0.09 (cf. Tables S1 and S2). The low Pearson, Kendall, and
211 Spearman correlations indicate that any trends, if present, are a complex non-linear problem. For
212 this task, ANNs are suitable.

213 The main results of the ANN analyses are listed in Table 3. The complete set of results is
214 listed in Tables S3 to S8. Machine learning by multilayer perceptron was the most successful in
215 predicting hospitalization related to air pollution from forest fires. The MLP was also well
216 constrained by the explanatory variables, as confirmed using the Friedman test. The p values
217 were nearly zero, meaning a change in the model led to distinct results. The best performance by
218 MLP corroborates recent research about air pollution and health impacts^{33,39} that MLP tends to
219 outperform the “newer” ANN models such as RBF and the unorganized machines (ELM and
220 ESN). The fully learned structure of MLP allows for advantageous approximation of the
221 nonlinear mapping inputs.

222 The causality of fire count and black carbon concentrations as explanatory variables of
223 hospital admissions was evaluated by replicating the machine learning for three cases: (1)
224 including all input variables (daily fire count, black carbon, mean temperature, relative humidity,
225 precipitation, solar radiation, wind speed, and wind direction), (2) excluding the fire count, and
226 (3) excluding black carbon. As expected, the best performance used all input variables (Table 3).
227 For respiratory illness, there was a one-day lag between health effects and exposure. For
228 cardiovascular illness, there was no lag. Results changed drastically when excluding fire count or

229 black carbon as explanatory variables. Lag increased from 1 day to 2-3 days for respiratory
230 illness and from no lag to 6 days for cardiovascular illness. We may conclude that incorporating
231 both fire count and black carbon (BC) concentration is crucial for accurately predicting hospital
232 admissions because these variables capture distinct aspects of exposure to fire-related pollutants.
233 While BC provides a direct measure of particulate matter generated by fires, which has
234 immediate health impacts, the fire count reflects the broader scale and intensity of fire activity,
235 including additional fire-generated pollutants and physical effects that may not be directly
236 captured by BC measurements alone. The improved model performance when both variables are
237 included suggests that non-fire sources of BC and other pollutants resulting from fire activity
238 contribute significantly to health outcomes. These findings are consistent with literature⁴⁹
239 indicating that aging smoke, which contains various harmful compounds beyond BC, and
240 interactions with urban pollutants, may exacerbate health impacts.

241 Figure 3 plots the observed and predicted values of the test dataset using the best-
242 performing model. The figure shows the model has difficulty in predicting higher and lower
243 values, especially for discontinuous variables. The predictions for hospitalizations are related
244 only to the considered inputs, mainly BC and fires. By comparison, the output of hospital
245 admission depends on many other factors not included in the model, such as lifestyle, age,
246 economic factors, and so on. In this sense, errors around 27% and 38% (Table 3) are acceptable
247 and provide reasonable predictive power for discontinuous datasets. Some studies using ANN
248 and considering the same outputs (cardiorespiratory diseases)^{17-19,21,38} presented results with
249 errors of the same order of magnitude (ranging from 17 to 36%), which confirms that the ANNs
250 showed a good performance to our database.

251 The emissions from the Amazon fires affect the short- and long-term health of the 25
252 million people living throughout the Amazon biome, as reflected in increased hospital
253 admissions for respiratory and cardiovascular diseases during the burning season. During 2021
254 and 2022, a trend of worsening conditions in Amazonas state continued²⁷, suggesting even more
255 for the future. The development of effective early-warning mechanisms could prevent severe
256 health impacts, but the precise mix of conditions to activate an alert has been uncertain in this
257 data-poor region. In this context, the study herein used the best available health records in the
258 central Amazon, corresponding to the several million people living in Manaus, and likewise the
259 best-available datasets of air quality. Simple multilinear regressions failed to establish Pearson,
260 Kendall, or Spearman coefficients above a noise threshold between environmental measurements
261 and health effects. Analysis by artificial neural networks did establish relationships, however.
262 Specifically, the ANNs successfully modeled the time lag between fire incidence and hospital
263 admissions based on input factors of daily averages of black carbon concentrations, temperature,
264 relative humidity, precipitation, solar radiation, and regional fire count. The prediction of the
265 time lag between environmental observations and health effects can be used for health-oriented
266 decision-making. Timely information can be provided in advance to the health care sector to
267 properly allocate resources during periods when admissions are expected to increase. The ANN
268 approach developed herein can be applied in diverse scenarios worldwide to forecast health
269 hazards resulting from fires. In the long term, the study results highlight the need for effective
270 measures to reduce fire occurrence and thereby mitigate the adverse impacts of regional air
271 pollution on human health.

272 The Amazon region has one of the highest deforestation rates worldwide⁵⁰, related to
273 socioeconomic factors, governance effectiveness, and climate change. The wildfire smoke

274 negatively impacts the human health of local communities, ecosystems, and climate change. For
275 Manaus, Paralovo et al.⁵¹ report on the decline in air quality from forest wildfires and controlled
276 burning after deforestation. Throughout the last few years, this issue is growing more severe. In
277 Amazonia, deforestation, maintenance of cleared areas, and forest fires account for 8%, 39%,
278 and 53% of fire outbreaks, respectively, with distinct social and environmental impacts.⁷ At
279 times in the dry season, a smoke layer envelops the larger part of the entire Amazon basin and
280 much of central South America. Marlier et al.⁵² highlighted that the duration of the dry season is
281 lengthening, which may increase the incidence of fire in Amazonia.

282 In the Amazon, wildfire smoke, often driven by deforestation and land-use changes,
283 poses severe public health challenges, leading to increased hospital admissions for
284 cardiorespiratory illnesses⁷. Similarly, California and Australia experience intense wildfires,
285 exacerbated by climate change, resulting in episodes of air pollution and corresponding health
286 impacts. However, while California and Australia have more developed monitoring and response
287 measures, the Amazon struggles with a scarcity of air quality monitoring stations, limiting the
288 ability to forecast and mitigate these impacts effectively. In Southeast Asia⁹, particularly during
289 agricultural burning seasons, air pollution reaches critical levels, causing adverse health effects
290 akin to those observed in the Amazon. This global analysis highlights the urgent need for early
291 warning systems and targeted mitigation policies tailored to the unique socioeconomic and
292 environmental contexts of each region, with the aim of protecting populations exposed to
293 wildfire smoke. The results show the potential of ANN as a tool capable of predicting forest fire
294 health risks, suggesting their potential to support the creation of an early warning system,
295 although further research is needed to fully elucidate the underlying mechanisms and optimize
296 the timing of interventions. Some limitations that need to be addressed in future research are the

297 use of different proxies for fires, more complete databases, emergency visits as output, and to
298 include socioeconomic and other individual characteristics.

299 Relating our findings to the United Nations' Sustainable Development Goals (SDGs), our
300 study aligns significantly with SDG 3 (Good Health and Well-being), SDG 13 (Climate Action),
301 and SDG 15 (Life on Land). SDG 3 aims to ensure healthy lives and promote well-being for all
302 ages, and our research underscores the critical need for early warning systems to reduce the
303 adverse health impacts of wildfire smoke, particularly in vulnerable populations. The observed
304 increase in cardiorespiratory illnesses due to black carbon exposure from wildfires directly
305 relates to this goal by highlighting the importance of mitigating environmental health risks. SDG
306 13 focuses on combating climate change and its impacts, and our study provides evidence of the
307 exacerbated wildfire activity driven by climate change, emphasizing the urgent need for climate
308 action to reduce the frequency and intensity of these fires. Lastly, SDG 15 seeks to protect,
309 restore, and promote sustainable use of terrestrial ecosystems. Our findings on the detrimental
310 health effects of wildfires on Manaus' population illustrate the critical need for sustainable land
311 management practices. By addressing these SDGs, our research contributes to a broader
312 understanding of the intersection between environmental degradation and public health,
313 advocating for integrated policies that promote a healthier and more sustainable future.

314

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References

- (1) Wu, Z.; Li, M.; Wang, B.; Quan, Y.; Liu, J. Using Artificial Intelligence to Estimate the Probability of Forest Fires in Heilongjiang, Northeast China. *Remote Sensing*. **2021**, 13(9), 1813. DOI: 10.3390/rs13091813.
- (2) The Copernicus Programme. <https://atmosphere.copernicus.eu/europes-summer-wildfire-emissions-highest-15-years#:~:text=According%20to%20data%20from%20the%20CAMS%20Global%20Fire%20Assimilation%20System,the%20highest%20level%20since%202007> (accessed 2022-10-3).
- (3) Safford, H. D.; Paulson, A. K.; Steel, Z. L.; Young, D. J. N.; Wayman, R. B. The 2020 California Fire Season: A Year Like no Other, a Return to the Past or a Harbinger of the Future? *Global Ecology and Biogeography*. **2020**. DOI: 10.1111/geb.13498.

- (4) Abram, N. J.; Henley, B. J.; Gupta, A. S.; Lippmann, T. J. R.; Clarke, H.; Dowdy, A. J.; Sharples, J. J.; Nolan, R. H.; Zhang, T.; Wooster, M. J.; Wurtzel, J. B.; Meissner, K. J.; Pitman, A. J.; Ukkola, A. M.; Murphy, B. P.; Tapper, N. J.; Boer, M. M. Connections of Climate Change and Variability to Large and Extreme Forest Fires in Southeast Australia. *Communications Earth & Environment*. **2021**, 2 (8). DOI: 10.1038/s43247-020-00065-8.
- (5) Bowman, D. M. J. S.; Kolden, C. A.; Abatzoglou, J. T.; Johnston, F. H.; van der Werf, G. R.; Flannigan, M. Vegetation Fires in the Anthropocene. *Nature Reviews Earth & Environment*. **2020**, DOI: 10.1038/s43017-020-0085-3.
- (6) INPE, National Institute for Space Research. https://queimadas.dgi.inpe.br/queimadas/portal-static/estatisticas_estados/ (accessed 2023-02-15).
- (7) Urrutia-Pereira, M.; Rizzo, L. V.; Chong-Neto, H. J.; Solé, D. Impact of Exposure to Smoke from Biomass Burning in the Amazon Rain Forest on Human Health. *J. Bras. Pneumol.* **2021**, 47(5). DOI: 10.36416/1806-3756/e20210219.
- (8) Liakakou, E.; Stavroulas, I.; Kaskaoutis, D. G.; Grivas, G.; Paraskevopoulou, D.; Dumka, U. C.; Tsagkaraki, M.; Bougiatioti, A.; Oikonomou, K.; Sciare, J.; Gerasopoulos, E.; Mihalopoulos, N. Long-Term Variability, Source Apportionment and Spectral Properties of Black Carbon at an Urban Background Site in Athens, Greece. *Atmos. Environ.* **2020**, 222, 117137. DOI: 10.1016/j.atmosenv.2019.117137.
- (9) Pani, S. K.; Wang, S-H.; Lin, N-H.; Chantara, S.; Lee, C-T.; Thepnuan, D. Black Carbon Over an Urban Atmosphere in Northern Peninsular Southeast Asia: Characteristics, Source

Apportionment, and Associated Health Risks. *Environ. Pollut.* **2020**, 259, 113871. DOI: 10.1016/j.envpol.2019.113871.

(10) Black, C.; Tesfaigzi, Y.; Bassein, J. A.; Miller, L. A. Wildfire Smoke Exposure and Human Health: Significant Gaps in Research for a Growing Public Health Issue. *Environmental Toxicology and Pharmacology*. **2017**, 55. DOI: 10.1016/j.etap.2017.08.022.

(11) Cascio, W. E. Wildland Fire Smoke and Human Health. *Sci. Total Environ.* **2018**, 624, 586–595. DOI: 10.1016/j.scitotenv.2017.12.086.

(12) Johnston, F. H.; Henderson, S. B.; Chen, Y.; Randerson, J. T.; Marlier, M.; Defries, R. S.; Kinney, P.; Bowman, D. M. J. S.; Brauer, M. Estimated Global Mortality Attributable to Smoke from Landscape Fires. *Environ. Health Perspect.* **2012**, 120, 695-701. DOI: 10.1289/ehp.1104422.

(13) Nawaz, M. O.; Henze, D.K. Premature Deaths in Brazil Associated with Long-Term Exposure to PM_{2.5} from Amazon Fires Between 2016 and 2019. *GeoHealth*. **2020**, 4(8). DOI: 0.1029/2020GH000268.

(14) Alves, N. O.; Vessoni, A. T.; Quinet, A.; Fortunato, R. S.; Kajitani, G. S.; Peixoto, M. S.; Hacon, S. S.; Artaxo, P.; Salvida, P.; Menck, C. F. M.; Medeiros, S. R. B. Biomass Burning in the Amazon Region Causes DNA Damage and Cell Death in Human Lung Cells. *Scientific Reports*. **2017**, 7, 10937 (2017). DOI: 10.1038/s41598-017-11024-3.

(15) Prist, P. R.; Sangermano, F.; Bailey, A.; Bugni, V.; Villalobos-Segura, M.C.; Pimiento-Quiroga, N.; Daszak, P.; Zambrana-Torrel, C. Protecting Brazilian Amazon Indigenous Territories Reduces Atmospheric Particulates and Avoids Associated Health Impacts and Costs.

Communications earth & environment. Nature. **2023**, 4 (34). DOI: 10.1038/s43247-023-00704-w.

(16) Reid, C. E.; Brauer, M.; Johnston, F. H.; Jerrett, M.; Balmes, J. R.; Elliott, C. T. Critical Review of Health Impacts of Wildfire Smoke Exposure. *Environ. Health Perspect.* **2016**, 124 (9), 1334-1343. DOI: 10.1289/ehp.1409277.

(17) Polezer, G.; Tadano, Y.S.; Siqueira, H.V.; Godoi, A. F. L.; Yamamoto, C. I.; André, P. A.; Pauliquevis, T.; Andrade, M. F.; Oliviera, A.; Saldiva, P. H. N.; Taylor, P. E.; Godoi, R. H. M. Assessing the Impact of PM_{2.5} on Respiratory Disease Using Artificial Neural Networks. *Environ. Pollut.* **2018**, 235, 394–403. DOI: 10.1016/j.envpol.2017.12.111.

(18) Araujo, L. N.; Belotti, J. T.; Antonini Alves, T.; Tadano, Y. S.; Siqueira, H. Ensemble Method Based on Artificial Neural Networks to Estimate Air Pollution Health Risks. *Environ. Model. Softw.* **2020**, 123. DOI: 10.1016/j.envsoft.2019.104567.

(19) Kachba, Y.; Chiroli, D. M. G.; Belotti, J. T.; Antonini Alves, T.; Tadano, Y.S.; Siqueira, H. Artificial Neural Networks to Estimate the Influence of Vehicular Emission Variables on Morbidity and Mortality in the Largest Metropolis in South America. *Sustainability.* **2020**, 12, 2621. DOI: 10.3390/su12072621.

(20) Kassomenos, P.; Petrakis, M.; Sarigiannis, D.; Gotti, A.; Karakitsios, S. Identifying the Contribution of Physical and Chemical Stressors to the Daily Number of Hospital Admissions Implementing an Artificial Neural Network Model. *Air Qual. Atmos. Health.* **2011**, 4 (3-4), 263-272. DOI: 10.1007/s11869-011-0139-2.

- (21) Tadano, Y. S.; Siqueira, H.; Antonini Alves, T. Unorganized Machines to Predict Hospital Admissions for Respiratory Diseases. In: *Computational Intelligence (LA-CCI)*. IEEE Latin American Conference on, 1-6. **2016**. DOI: 10.1109/LA-CCI.2016.7885699.
- (22) Tadano, Y. S.; Potgieter-Vermaak, S.; Kachba, Y. R.; Chiroli, D. M. G.; Casacio, L.; Santos-Silva, J. C.; Moreira, C. A. B.; Machado, V.; Antonini Alves, T.; Siqueira, H.; Godoi, R. H. M. Dynamic Model to Predict the Association Between Air Quality, COVID-19 Cases, and Level of Lockdown. *Environmental Pollution*. **2021a**, 268, 115920. DOI: 10.1016/j.envpol.2020.115920.
- (23) IBGE: Brazilian Institute of Geography and Statistics. Cidades. <https://cidades.ibge.gov.br/brasil/am/manaus/panorama>. (accessed: 2023-05-14).
- (24) INMET - National Institute of Meteorology. Climatological Charts (1931-1960 and 1961-1990). <http://www.inmet.gov.br/portal/index.php?r=clima/graficosclimaticos> (accessed: 2018-11-10).
- (25) Wu, L.; Li, X.; Kim, H.; Geng, H.; Godoi, R. H. M.; Barbosa, C. G. G.; Godoi, A. F. L.; Yamamoto, C. I.; Souza, R. A.F.; Pöhlker, C.; Andreae, M. O.; Ro, C-U. Single-Particle Characterization of Aerosols Collected at a Remote Site in the Amazonian Rainforest and an Urban Site in Manaus, Brazil. *Atmos. Chem. Phys.* **2019**, 19, 1221-1240. DOI: 10.5194/acp-19-1221-2019.
- (26) Google Maps. Map data©2019 Google; <https://www.google.com/maps/place/Brasil>, and Google Earth Pro (Map data©2019 Google; <https://www.google.com/maps/@-10,-55.00001,12646636m/data=!3m1!1e3>). The maps were edited with PowerPoint (version 16.28-

19081202).

(27) INPE – National Institute for Space Research. Fires database.

<http://www.inpe.br/queimadas/bdqueimadas>. (accessed: 2023-10-02).

(28) INPE – National Institute for Space Research. What is reference satellite? <https://queimadas.dgi.inpe.br/queimadas/portal/informacoes/perguntas-frequentes#p7>. (accessed: 2022-11-11).

(29) Youssouf, H.; Liousse, C.; Roblou, L.; Assamoi, E. M.; Salonen, R. O.; Maesano, C.; Banerjee, S.; Annesi-Maesano, I. Quantifying Wildfires Exposure for Investigating Health-Related Effects. *Atmos. Environ.* **2014a**, 97, 239–251. DOI: 10.1016/j.atmosenv.2014.07.041.

(30) DATASUS, 2019. SIHSUS (Sistema de Informações Hospitalares do Sistema Único de Saúde - *Hospital Information System of Public Health*). Brazilian Health Ministry. <https://datasus.saude.gov.br/aceso-a-informacao>. (accessed: 2019-04-15).

(31) Belotti, J. T., Castanho D. S., Araujo, L. N., Da Silva, L. V., Alves, T. A., Tadano, Y. S., Stevan, S. L., Corrêa, F. C. Siqueira, H. V. Air Pollution Epidemiology: A Simplified Generalized Linear Model Approach Optimized by Bio-Inspired Metaheuristics. *Environmental Research.* **2020a**, 191, 110106.

(32) Belotti, J., Siqueira, H., Araujo, L., Stevan Jr, S. L., de Mattos Neto, P. S., Marinho, M. H., de Oliveira, J. F. L., Usberti, F., de Almeida Leone Filho, M., Converti, A., Sarubbo, L. A. Neural-based ensembles and unorganized machines to predict streamflow series from hydroelectric plants. *Energies*, **2020b**, 13(18), 4769.

- (33) Tadano, Y. S., Ugaya, C. M. L., Franco, A. T. Methodology to Assess Air Pollution Impact on Human Health Using the Generalized Linear Model with Poisson Regression. In: Kare, M. (Ed.), *Air Pollution-Monitoring, Modelling and Health*. InTech, **2012**, 281-304. DOI: 10.5772/33385.
- (34) Haykin, S. *Neural Networks and Learning Machines*, 3rd ed., Pearson Upper Saddle River, NJ, USA, **2009**.
- (35) Siqueira, H., Boccato, L.; Attux, R.; Lyra, C. Unorganized Machines for Seasonal Streamflow Series Forecasting, *International Journal of Neural Systems*, **2014**, 24, 1430009. DOI: 10.1142/S0129065714300095.
- (36) Siqueira, H., Boccato, L.; Luna, I.; Attux, R.; Lyra, C. Performance Analysis of Unorganized Machines in Streamflow Forecasting of Brazilian Plants. *Applied Soft Computing*, **2018**, 68, 494-506. DOI: 10.1016/j.asoc.2018.04.007.
- (37) Siqueira, H.; Bacalhau, E. T.; Casacio, L.; Puchta, E.; Antonini Alves, T.; Tadano, Y. S. Hybrid unorganized machines to estimate the number of hospital admissions caused by PM10 concentration. *Environmental Science and Pollution Research*, **2023**, 30, 113175-113192. DOI: 10.1007/s11356-023-30180-w.
- (38) Tadano, Y.S.; Bacalhau, E. T.; Casacio, L.; Puchta, E.; Pereira, T. S.; Antonini Alves, T.; Ugaya, C. M. L.; Siqueira, H. Unorganized Machines to Estimate the Number of Hospital Admissions Due to Respiratory Diseases Caused by PM₁₀ Concentration. *Atmosphere*. **2021b**, 12 (10), 1345. DOI: 10.3390/atmos12101345.

- (39) Youssouf, H.; Liousse, C.; Roblou, L.; Assamoi, E-M.; Salonen, R. O.; Maesano, C.; Banerjee, S.; Annesi-Maesano, I. Non-Accidental Health Impacts of Wildfire Smoke. *Int. J. Environ. Res. Public Health*. **2014b** 11, 11772-11804. DOI: 10.3390/ijerph111111772.
- (40) Sant'Anna, A. A.; Rocha, R. Health Impacts of Deforestation-Related Fires in the Brazilian Amazon. Technical Report. Instituto de Estudos para Políticas de Saúde. https://www.hrw.org/sites/default/files/media_2020/08/Health%20Impacts%20of%20Deforestation-Related%20Fires%20in%20the%20Amazon_EN_0.pdf. (accessed: 2023-04-27).
- (41) White, W. H.; Trzepla, K.; Hyslop, N. P.; Schichtel, B. A. A Critical Review of Filter Transmittance Measurements for Aerosol Light Absorption, and de Novo Calibration for a Decade of Monitoring on PTFE Membranes. *Aerosol Science and Technology*. 2016, 50 (9). DOI: 10.1080/02786826.2016.1211615.
- (42) Bond, T. C.; Bergstrom, R. W. Light Absorption by Carbonaceous Particles: An Investigative Review. *Aerosol Science and Technology*. 2006, 40 (1). DOI: 10.1080/02786820500421521.
- (43) Zhang, Z.; Cheng, Y.; Liang, L.; Liu, J. The Measurement of Atmospheric Black Carbon: A Review. *Toxics*. 2023, 11 (12). DOI: 10.3390/toxics11120975.
- (44) Andrade-Filho, V. S.; Artaxo, P.; Hacon, S.; Carmo, C. N.; Cirino, G. Aerosols From Biomass Burning and Respiratory Diseases in Children, Manaus, Northern Brazil. *Revista Saúde Pública*. **2013**, 47 (2). DOI: 10.1590/S0034-8910.2013047004011.

- (45) Fernandes, K. S.; Machado C. M. D. WSOC and Its Relationship with BC, Levoglucosan and Transition Metals in the PM_{2.5} of an Urban Area in the Amazon. *J. Braz. Chem. Soc.* **2021**, 33 (6). DOI: 10.21577/0103-5053.20220011.
- (46) Jacobson, L. V.; Hacon, S.; Ignotti, E.; Hermano, C.; Artaxo, P.; de Leon, A. P. Effects of Air Pollution from Biomass Burning in Amazon: A Panel Study of Schoolchildren. *Epidemiology*. **2009**, 20, S90. DOI: 10.1097/01.ede.0000362981.13814.a2.
- (47) Filonchyk, M.; Peterson, M. P.; Sun, D. Deterioration of Air Quality Associated with the 2020 US Wildfires. *Science of the Total Environment*. **2022**, 826. DOI: 10.1016/j.scitotenv.2022.154103.
- (48) Ahangar, F. E.; Pakbin, P.; Hasheminassab, S.; Epstein, S. A.; Li, X.; Polidori, A.; Low, J. Long-Term Trends of PM_{2.5} and Its Carbon Content in the South Coast Air Basin: A Focus on the Impact of Wildfires. *Atmospheric Environment*. **2021**, 255(15), 118431. DOI: 10.1016/j.atmosenv.2021.118431.
- (49) Shrivastava, M.; Andreae, M. O.; Artaxo, P.; Barbosa, H. M. J.; Berg, L. K.; Brito J.; Ching, J.; Easter, R. C.; Fan, J.; Fast, J. D.; Feng, Z.; Fuentes, J. D.; Glasius, M.; Goldstein, A. H.; Alves, E. G.; Gomes, H.; Gu, D.; Guenther, A.; Jathar, S. H.; Kim, S.; Liu, Y.; Lou, S.; Martin, S. T.; McNeill, F.; Medeiros, A.; Sá, S. S.; Shilling, J. E.; Springston, S. R.; Souza, R. A. F.; Thornton, J. A.; Isaacman-VanWertz, G.; Yee, L. D.; Ynoue, R.; Zaveri, R. A.; Zelenyuk, A.; Zhao, C. Urban Pollution Greatly Enhances Formation of Natural Aerosols Over the Amazon Rainforest. *Nature Communications*. **2019**, 10. DOI: 10.1038/s41467-019-08909-4.

(50) Ribeiro, I. O.; Andreoli, R. V.; Kayano, M. T.; Sousa, T. R.; Medeiros, A. S.; Guimarães, P. C.; Barbosa, C. G. G.; Godoi, R. H. M.; Martin, S. T.; Souza, R. A. F. Impact of the Biomass Burning on Methane Variability During Dry Years in the Amazon Measured from an Aircraft and the AIRS Sensor. *Science of The Total Environment*. **2018b**, 624, 509-516. DOI: 10.1016/j.scitotenv.2017.12.147.

(51) Paralovo, S. L.; Barbosa, C. G. G.; Carneiro, I. P. S.; Kurzlop, P.; Borillo, G. C.; Schiochet, M. F. C.; Godoi, A. F. L.; Yamamoto, C. I.; Souza, R. A. F.; Andreoli, R. V.; Ribeiro, I. O.; Manzi, A. O.; Kourtchev, I.; Bustillos, J. O. V.; Martin, S. T.; Godoi, R. H. M. Observations of Particulate Matter, NO₂, SO₂, O₃, H₂S and Selected VOCs at a Semi-Urban Environment in the Amazon Region. *Sci. Total Environ.* **2019**, 650, 996–1006. DOI: 10.1016/j.scitotenv.2018.09.073.

(52) Marlier, M. E.; Bonilla, E. X.; Mickley, L. J. How do Brazilian Fires Affect Air Pollution and Public Health? *GeoHealth*. **2020**, 4:e2020GH000331. DOI: 10.1029/2020GH000331.

Table 1. Statistics of observations. *Source: J00 - J99 from ICD-10. **Source: I00 to I99 from ICD-10.

Variable	Season	Statistic	2011	2012	2013	2015	2016
Number of days in analysis	wet		45	115	114	34	135
	dry		13	123	48	72	86
PM _{2.5} (µg m ⁻³)	wet	max	19.4	30.0	68.7	24.9	16.2
		min	4.3	1.7	0.8	4.8	1.9
		avg	9.7	8.9	9.5	10.5	5.6
		median	8.5	6.8	6.9	8.4	5.3
	dry	max	16.3	58.0	31.9	29.6	22.1
		min	4.5	0.0	2.1	3.5	2.6
		avg	10.8	11.6	10.8	11.5	7.7
		median	8.9	10.0	9.0	8.5	7.0
BC (µg m ⁻³)	wet	max	4.7	5.6	4.7	3.4	3.7
		min	0.4	0.4	0.4	0.7	0.6
		avg	2.0	2.0	2.0	1.3	1.4
		median	1.9	1.7	1.9	0.9	1.3
	dry	max	5.2	5.2	4.0	3.2	5.1
		min	0.1	0.2	0.2	0.3	0.6
		avg	2.2	2.3	1.8	1.3	2.0
		median	1.8	1.9	1.7	1.2	1.7
Daily hospitalization count: respiratory disease *	wet	max	45	85	64	34	58
		min	7	12	15	9	10
		avg	26.36	31.67	34.51	22.18	27.78
		median	25	27	34	23	26

	dry	max	34	49	51	37	63
		min	19	14	18	8	9
		avg	27.54	29.24	34.15	20.92	29.28
		median	28	28	33	20	27
Daily hospitalization count: cardiovascular disease **	wet	max	37	40	37	30	32
		min	9	5	8	10	5
		avg	19.51	19.57	19.65	18.71	19.50
		median	19	19	19	18	19
	dry	max	27	48	29	31	37
		min	9	4	4	10	8
		avg	18.08	20.58	17.52	18.65	19.15
		median	18	20	18	19	19
Fire count	wet		153	375	30	295	166
	dry		100	1089	167	2298	377

Table 2. Comparison between PM_{2.5} and BC concentrations of this study to those in literature.

*Estimated based on satellite image.

Reference	Region	Period	PM_{2.5} conc ($\mu\text{g m}^{-3}$)	BC conc ($\mu\text{g m}^{-3}$)
this study	Manaus	2011-2013, 2015- 2016	9.20 ± 6.83	1.83 ± 0.99
Andrade Filho et al. ⁴⁴	Manaus	2002-2009	15*	-
Fernandes et al. ⁴⁵	Manaus	Aug to Sep 2017	14.7	3.0
Jacobson et al. ⁴⁶	Mato Grosso, Brazil	2008	19.6 ± 11.9	1.00 ± 0.48

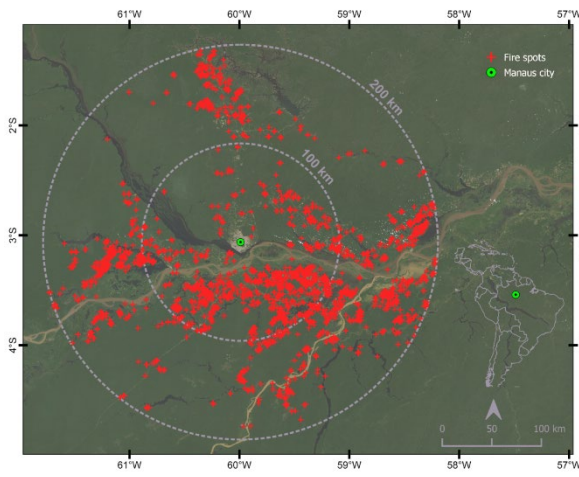
Table 3. Best-performing neural networks for predictions of hospital admissions. Results are listed for respiratory and cardiovascular diseases. Abbreviations include artificial neural network (ANN), neuron count (NC) of each hidden layer, standard score (Z), root mean square error (RMSE), mean absolute percent error (MAPE), and multilayer perceptron (MLP) with 1 or 2 hidden layers.

Inputs	ANN	NC	Z	lag (day)	RMSE	MAPE (%)
<i>Respiratory disease</i>						
All	MLP-2	60 / 40	without	1	10.4	38
Excluding forest fires	MLP-2	40 / 50	with	3	11.8	36
Excluding black carbon	MLP-2	40 / 100	without	2	11.7	35
<i>Cardiovascular disease</i>						
All	MLP-2	40 / 80	without	0	4.9	27
Excluding forest fires	MLP-2	70 / 40	with	6	5.2	25
Excluding black carbon	MLP-1	80	without	6	5.2	24



Figure 1. Sampling locations for 2011-2013 ($3^{\circ}5'43.94''\text{S}$, $59^{\circ}59'25.56''\text{W}$) and 2015-2016 ($3^{\circ}6'12.5''\text{S}$, $59^{\circ}58'55.8''\text{W}$). In counterclockwise direction, the four panels show progressively smaller scales from South America at the largest scale (top left), to Amazonas, to Manaus environs (corresponding to red box), to localized urban view of Manaus at the smallest scale (top right).²⁶

A.



B.

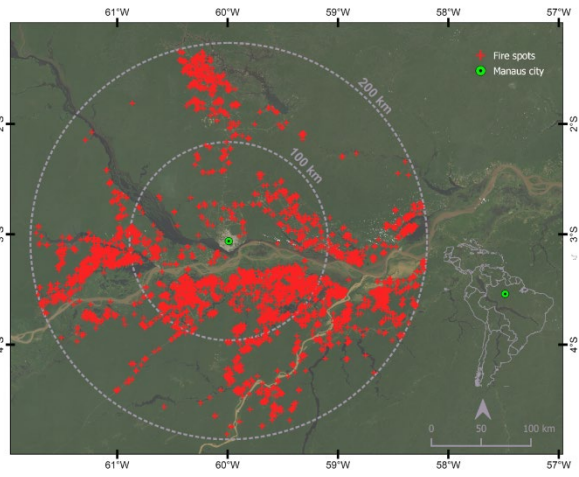
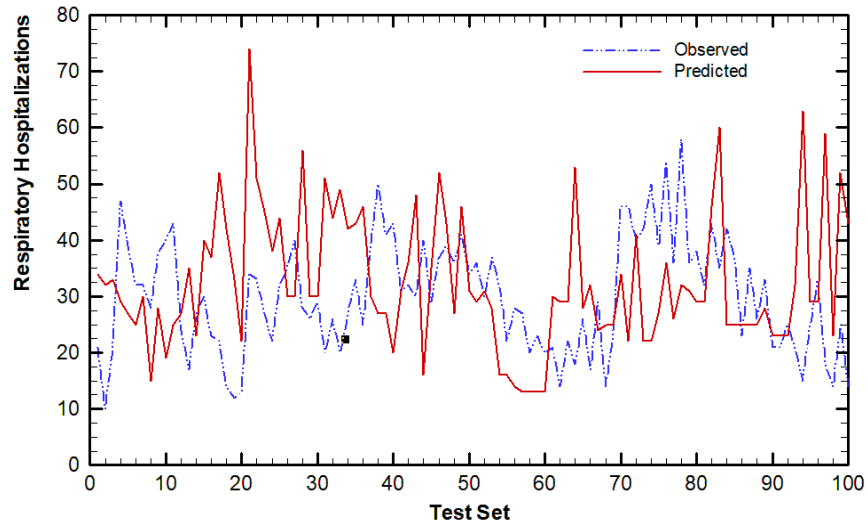


Figure 2. Forest fire count in a 200-km radius around Manaus for the two BC campaign periods
(a) 5 Oct 2011 to 26 Jul 2013 and (b) 15 Aug 2015 to 30 Aug 2016.⁴⁴

A.



B.

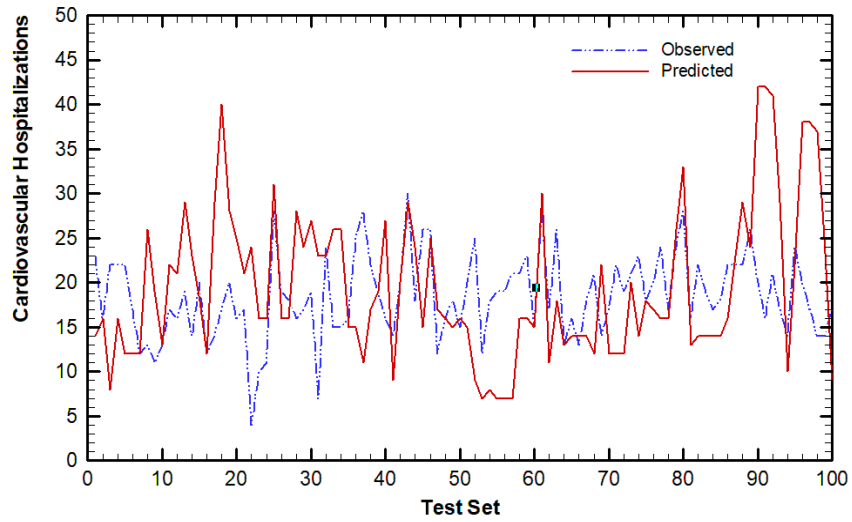


Figure 3. Comparison between observed and predicted hospitalizations across test dataset for (A) respiratory disease using MLP-2 with a one-day lag and (B) cardiovascular disease using MLP-2 and no lag.