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1 **Infant dietary exposure to dioxin-like polychlorinated biphenyls (dlPCBs),**
2 **polybrominated and mixed halogenated dibenzo-p-dioxins and furans (PBDD/Fs and**
3 **PXDD/Fs) in milk samples of lactating mothers in Accra, Ghana.**

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35

36 **Abstract**

37

38 In this study, polybrominated and mixed halogenated dibenzo-p-dioxins and furans (PBDD/Fs
39 and PXDD/Fs), and dioxin-like polychlorinated biphenyls (dlPCBs) were quantified in 24
40 human milk samples of first-time lactating mothers from Greater Accra region in Ghana. The
41 aims of the study were to determine the concentrations and toxic equivalent concentrations of
42 PBDD/F, PXDD/F and dlPCBs in human milk, and to estimate an infant's daily intake. The
43 samples were analysed for 12 dioxin-like PCBs, 7 congeners of 2,3,7,8-polybrominated
44 dibenzo-p-dioxins and furans (PBDD/Fs), and 7 congeners of 2,3,7,8-mixed halogenated
45 dioxins and furans (PXDD/Fs, where X= Br/Cl). The mean concentrations in human milk
46 ranged from 0.15-212.9 pg/g lipid for dlPCB congeners (mean TEQ: 1.67 pg WHO₂₀₀₅-TEQ/g
47 lipid). Lesser concentrations for 2,3,7,8-PXDD/Fs (and PBDD/Fs congeners) ranged between
48 <0.01-1.67 pg/g lipid, with a total mean tentative TEQ of 0.56 pg WHO₂₀₀₅-TEQ/g lipid. For
49 an infant of average weight 7 kg, consuming an estimated volume of 600 mL human milk, the
50 estimated average daily intake of dlPCBs in 21 human milk samples was 4.95 pg TEQ/kg
51 bw/day; contributions from dlPCBs, PXDD/Fs and PBDD/Fs resulted in an average estimated
52 daily intake of 6.56 pg TEQ/kg bw/day. The results obtained in this study, although lower than
53 infant dietary intake estimates in human milk from industrialized countries, exceeded the
54 recommended safety standards of 1 pg TEQ/kg bw/day and 1-4 pg TEQ/kg bw/day from the
55 Agency for Toxic Substances and Disease Registry (ATSDR) and the World Health
56 Organization (WHO), respectively.

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60

61 **1.0 Introduction**

62

63 Dioxins and dioxin-like compounds (DLCs) are toxic classes of environmental
64 contaminants that can be transferred from mother to foetus/baby via placenta/breastfeeding
65 (Schechter et al., 2006; van den Berg et al., 2017; van Leeuwen et al., 2000a). Oral intake of
66 foods by adults contribute over 90% of the total daily exposure to dioxins and DLCs. Similarly,
67 human milk, although noted for its significance in protecting an infants' health, is a major
68 source of toxic contaminants for a developing child (Djien Liem et al., 2000; Schechter et al.,
69 1996; Schechter et al., 1998b; Victora et al., 2016). The daily intake of dioxins and DLCs by
70 breastfed infants is linked to maternal body burdens. Due to bioaccumulation in lipid rich milk,
71 an infant's exposure is approximately 2 orders of magnitude higher than that of an average
72 adult (Djien Liem et al., 2000; Schechter et al., 1998b; Victora et al., 2016).

73

74 Dioxins and DLCs are produced from combustion and industrial processes of
75 brominated and/or chlorinated organic compounds (Djien Liem et al., 2000; van den Berg et
76 al., 2017; World Health Organization, 1996). The most toxic include 12 polychlorinated
77 biphenyls (dioxin-like PCBs- dlPCBs), and 17 congeners of 2,3,7,8-polychlorinated dibenzo-
78 p-dioxins and furans (PCDD/Fs) (Van den Berg et al., 2006). However, structurally related
79 analogues of 2,3,7,8- mixed halogenated dioxins and furans (PXDD/Fs, X=Br/Cl), and
80 polybrominated dibenzo-p-dioxins and furans (PBDD/Fs) exhibit similar toxicity profiles
81 (Birnbaum et al., 2003; Olsman et al., 2007; Van den Berg et al., 2006).

82

83 Globally, the vast majority of biomonitoring studies of DLCs have assessed risks of
84 infant/foetal exposure to intake of PCDD/Fs and dlPCBs in human milk/placental nutrients
85 (Dewailly et al., 1991; Koopman-Esseboom et al., 1994; Schechter et al., 1998a; Schechter et al.,
86 1998b; Schechter et al., 1990; Tanabe and Kunisue, 2007; van den Berg et al., 2017; Wittsiepe

87 et al., 2007). Relative to PCDD/Fs and dlPCBs, the risks of exposure of 2,3,7,8-PBDD/Fs and
88 PXDD/Fs for breastfed infants are understudied. These classes have been found to contribute
89 to the very limited literature on brominated and mixed halogenated dioxins, and significantly
90 to the total Toxic Equivalents (TEQs) in biomonitoring studies of vulnerable populations
91 (Bruce-Vanderpuije et al., 2019b; Jogsten et al., 2010; Ohta et al., 2004; Pratt et al., 2013).
92 Additionally, in occupationally exposed individuals (firefighters), PBDD/F TEQ concentration
93 in serum was approximately 20 times higher than for contributions from PCDD/Fs (Shaw et
94 al., 2013). Since the majority of studies performed suggest a linkage to neuro-developmental
95 and endocrine disrupting effects from exposure to dioxins and DLCs, investigation of human
96 biological matrices will allow for communication of biomonitoring data to population groups
97 most susceptible to exposure/high risks (Ames et al., 2018; Drover et al., 2019; Nakajima et
98 al., 2017; Sethi et al., 2019).

99

100 In contrast to the numerous studies that have reported dioxins and DLCs in human milk
101 in industrialized countries, most developing African countries have participated in only one
102 global baseline survey conducted by the World Health Organization/United Nations
103 Environment Programme (WHO/UNEP) (van den Berg et al., 2017). Exposure risks from
104 DLCs in human milk have been found to be predominantly higher in industrialized countries
105 compared to developing countries (Van den Berg et al., 2006; van den Berg et al., 2017; van
106 Leeuwen et al., 2000a; World Health Organization, 1996). The majority of infant exposure
107 studies have reported concentrations that exceed United States Environmental Protection
108 Agency (USEPA) safe levels of 0.2 pg TEQ/g lipid [for PCDD/Fs + dlPCBs] in human milk
109 (Djien Liem et al., 2000; Focant et al., 2002; Schecter et al., 1996; van den Berg et al., 2017;
110 World Health Organization, 1996). Reported TEQ-concentrations of PCDD/Fs in human milk
111 from developing countries showed a lower mean value of ~10 pg/g International (I)-TEQ/g

112 milk fat whilst industrialized countries had a range between 10-35 pg/g I-TEQ/g milk fat
113 (World Health Organization, 1998). Specifically, for Ghana, a developing country situated on
114 the southernmost part of West Africa, exclusive breastfeeding is encouraged for a minimum of
115 6 months. TEQ concentrations obtained from studies completed by the WHO/UNEP survey,
116 and by Adu-Kumi et al. (2010), were approximately 5 pg TEQ/g lipid and 6.1 pg TEQ/g lipid
117 in 50 pooled and 42 individual human milk samples (for PCDD/Fs + dlPCBs), respectively
118 (Adu-Kumi et al., 2010; van den Berg et al., 2017). It is important to note that these studies
119 only focused on dlPCBs and 2,3,7,8-PCDD/Fs. As brominated and mixed halogenated
120 dioxins/furans have similar mechanisms of action, their toxicity is possibly additive to that of
121 other DLCs. Thus, exclusion of 2,3,7,8-PBDD/Fs and PXDD/Fs may lead to underestimation
122 of the total TEQ and potential health risks.

123

124 A limited number of investigations on POPs in environmental matrices in Ghana have
125 been undertaken, and these have focused on chlorinated pollutants- organochlorine pesticides
126 (OCPs), polychlorinated biphenyls (mostly indicator PCBs) in environmental and food
127 matrices, and non-dlPCBs in human biomonitoring studies (Bruce-Vanderpuije et al., 2019a).
128 Little is known about exposure of vulnerable and occupationally-exposed populations to
129 dioxins and DLCs, especially with respect to brominated and mixed halogenated analogues.
130 Currently, there are a few studies that have focused on Ghanaian human biomonitoring; these
131 include studies on dioxins and DLCs in sera of occupationally exposed e-waste workers at
132 Agbogbloshie (Dai et al., 2020; Wittsiepe et al., 2015a), and background exposure
133 concentrations in sera of primiparous Ghanaians (Bruce-Vanderpuije et al., 2019b).
134 Additionally, low concentrations of PCB 118 (amongst other indicator PCBs) in human milk
135 samples from lactating mothers residing in Accra, Kumasi and Tamale in 2010 ranged between
136 1.9 and 3.0 ng/g lw (Asante et al., 2011). Data obtained from a baseline study for sera of

137 primiparous Ghanaians indicated a background exposure of foeti to dlPCBs (1.25 pg WHO-
138 TEQ/g lw), PCDD/Fs (3.10 pg WHO-TEQ/g lw) and PXDD/Fs & PBDD/Fs (0.99 pg WHO-
139 TEQ/g lw) in comparison to other studies reported globally. The significant contributions of
140 brominated and mixed halogenated dioxins and furans (~ 20%) in sera of exposed populations
141 to the total TEQs indicate possibilities of varied exposure sources including combustion, and
142 dietary exposure.

143

144 A limited number of studies focusing on PXDD/Fs and PBDD/Fs have been published
145 globally in human milk (Croes et al., 2013; Kotz et al., 2005; Ohta et al., 2004; Pratt et al.,
146 2013; Tue et al., 2014; Wiberg et al., 1992). In a previous study on dioxins and DLCs in sera
147 of primiparous Ghanaians, PXDD/Fs (and PBDD/Fs) contributed about 20% of the total TEQ
148 (Bruce-Vanderpuije et al., 2019b). Based on the possibilities of transfer of dioxins and DLCs
149 in human milk from an exposed mother to her baby, and the lack of quantified dioxins and
150 DLC data in human biomonitoring studies in Ghana, the aims of this study were to determine
151 the relative risk to infants from the consumption of human milk containing DLCs. This was
152 achieved by the:

- 153 1) Determination of concentrations and TEQ concentrations of 12 dlPCBs, 7 analytes
154 of 2,3,7,8-PBDD/Fs and PXDD/Fs in 24 Ghanaian human milk samples from first-
155 time lactating mothers, with no known occupational or accidental exposure, and
- 156 2) Calculation of Ghanaian infant dietary intake from exclusive breastfeeding.

157

158 From the exposure risk calculations, the intake values of dioxin and DLCs in Ghanaian
159 human milk samples obtained in this study were compared with recommended safety standard
160 values and other studies reported globally, to determine possibilities of risk to breastfed infants.

161

162 **2.0 Materials and Methods**

163

164 **2.1 Participant Recruitment**

165

166 In 2017, 24 first-time lactating Ghanaians (primiparous) who reside in industrialized
167 areas of Accra were voluntarily recruited by research nurses from Ridge Regional hospital
168 where they receive prenatal care. Participants completed written informed consent forms and
169 exposure assessment questionnaires during regular prenatal visits, prior to birth and sample
170 collection. Data on age, occupation, diet, cigarette smoking, alcohol intake and bodyweight
171 were documented. Human milk samples (20 mL, n = 24) were collected in April, 2017, during
172 the first two weeks after delivery, into 50 mL corning tubes and stored at -20 °C prior to
173 extraction and analysis. The study was approved by the Ethics Review Committee of the Ghana
174 Health Services, and conducted in accordance with ethical principles for medical research
175 involving human subjects.

176

177 **2.2 Reagents and Chemicals**

178

179 Distilled in glass grade organic solvents- n-hexane, toluene, nonane, acetonitrile,
180 methanol and water were obtained from Caledon Laboratories Limited (Georgetown, Ontario,
181 Canada). Octadecyl non-encapped bonded silica C₁₈ cartridges (10 g/75 mL) were obtained
182 from Thermo Fisher Scientific. Captiva EMR-Lipid removal cartridges (600 mg/6 mL) were
183 obtained from Agilent Technologies. For the separation of dlPCBs, PBDD/Fs and PXDD/Fs
184 from non-planar compounds, ultra clean carbon mini-columns (2%) and re-usable glass column
185 reservoirs (20 cm in length, 0.5 cm in diameter) from Cape Technologies were used. ¹³C-
186 labelled isotope and native dlPCBs, PBDD/Fs and PXDD/Fs standards were obtained from
187 Wellington Laboratories Inc. (Guelph, Ontario, Canada). Chromatographic separation column:
188 DB5-MS (5% diphenyl 95% dimethyl polysiloxane, 60 m x 0.25 mm ID x 10 µm film

189 thickness, J&W Scientific, CA, USA) was obtained from Agilent. Preparation of calibration,
190 recovery and injection standards for PXDD/Fs, PBDD/Fs and dlPCBs have been described in
191 detail in a previously published paper on sera of primiparous Ghanaians (Bruce-Vanderpuije
192 et al., 2019b). PBDD/F, dlPCB, and PXDD/F analytes analysed in this study are listed in Table
193 S2. All standards were prepared in nonane, except for recovery spiking solutions which was
194 prepared in methanol.

195

196 **2.3 Sample Extraction**

197

198 Sample extraction used in this study was based on previously validated analytical
199 method (Focant and De Pauw, 2002) with minor modifications. The analytical method involved
200 spiking 10 mL human milk sample with 5 μ L of 2 pg/ μ L label recovery mix- 12 congeners of
201 $^{13}\text{C}_{12}$ -dlPCBs and 2,3,7,8-substituted congeners of $^{13}\text{C}_{12}$ -PXDD/Fs and $^{13}\text{C}_{12}$ -PBDD/F (3-B-
202 2,7,8-CDF, 2,3-B-7,8-CDF, 2,3-B-7,8-CDD, 4-B-2,3,7,8-CDF, 1,3-B-2,7,8-CDF, BDD-
203 2,3,7,8)- to determine extraction efficiency, matrix effects on recovery and enable quantitation
204 by isotope dilution mass spectrometry. Extraction of lipophilic dioxins and DLCs was
205 performed on human milk samples using C_{18} SPE (10 g/75 mL) with acetonitrile and water,
206 after addition of sodium oxalate (20 mg/1 g milk) to disrupt the fat globules. Ten millilitres of
207 acetonitrile and 10 mL water were added. C_{18} cartridges were conditioned gravimetrically using
208 two cartridge volumes of acetonitrile and water prior to loading human milk mixture. Sample
209 was eluted at a flow rate of 5 mL/min. Human milk tubes were rinsed with 2 x 10 mL H_2O and
210 transferred onto C_{18} cartridge barrels. Cartridges were dried under vacuum pump suction for 1
211 hour to remove water. Analytes were eluted from C_{18} cartridge using 2 x 10 mL hexane and
212 collected in clear EPA vials, at a flow rate of 5 mL/min. All extracts were evaporated, and the
213 extracted lipids were determined gravimetrically after solvent evaporation. Of the 24

214 participants that were sampled, 3 human milk samples were omitted because of technical
215 difficulties during sample preparation.

216

217 **2.4 Lipid removal Clean-up and Fractionation**

218

219 Extracts were evaporated to 1 mL, and 3 mL acetonitrile was added. Extracts were
220 loaded onto Captiva-EMR lipid removal cartridge and allowed to flow under gravity. Vials
221 were rinsed with 5 mL acetonitrile. Eluate was evaporated under N₂ to 1 mL, and solvent was
222 exchanged for hexane. A 5 g acidified silica column, connected to 2% ultra clean carbon mini
223 column was activated with 20 mL hexane. The human milk extract was loaded, and cartridge
224 was rinsed with 30 mL hexane. The silica cartridge was replaced with reusable glass column
225 reservoir. The carbon column was inverted, and eluted with 30 mL toluene. Eluate was
226 collected in 40 mL EPA vial, evaporated to 350 µL under low N₂, transferred into inset vials,
227 evaporated to dryness, and reconstituted with 10 µL of 1 pg/µL injection standard.

228

229 **2.5 Instrumental Analysis**

230

231 The instrumental method used for the analyses of the human milk samples has been
232 previously reported in an earlier study developed for sera of pregnant Ghanaian women (Bruce-
233 Vanderpuije et al., 2019b). Briefly, sample analyses were performed using a GC-APCI-MS/MS
234 (Waters - Xevo TQ-XS). One microlitre of sample was injected in splitless mode on a DB5-
235 MS column (60 m x 0.25 mm x 10 µm), using an oven program optimized for separation of
236 dlPCBs, PBDD/Fs and PXDD/Fs. Instrumental parameters and operating conditions are
237 summarized in Table S1. The mass spectrometer was operated in positive ion mode, using
238 multiple reaction monitoring. Four transitions (2 quantifiers and 2 qualifiers for native and
239 ¹³C₁₂ label components) were monitored for dlPCBs, PBDD/Fs and PXDD/Fs. The transitions,

240 collision energies and isotope ratios are summarized in Table S2. Samples were analysed using
241 capillary gas chromatography with atmospheric pressure chemical ionization (APCI) and triple
242 quadrupole tandem mass spectrometry (GC-APCI-MS/MS, Xevo TQ-XS) from Waters
243 Corporation, Manchester, UK.

244

245 **2.5.1 Quality assurance/ Quality Control**

246

247 Analytes were quantified by isotope dilution using the ions specified in Table S2. For
248 linearity, the response obtained for a native, relative to its corresponding ^{13}C labelled standard
249 was linear for the range of calibration standards analyzed. Calculated coefficient of
250 determination for dlPCBs and PXDD/Fs (and PBDD/Fs) were $R^2 \geq 0.998$ and $R^2 \geq 0.995$
251 respectively. The % RSDs obtained for dlPCBs and PXDD/Fs and PBDD/Fs) ranged between
252 1.2 and 13.2%; this is in agreement with the acceptable 15% value (Centre for Diseases Control
253 and Prevention, 2006). Recoveries for isotopically labelled standards, spiked into human milk
254 prior to extraction/clean-up, ranged between 50-80% (recoveries obtained fell within the
255 acceptable EPA ranges). Where concentrations were below the LOD, $\frac{1}{2}$ LODs were assigned
256 and used in TEQ calculations. The instrument limit of quantitation was set at 5 fg/ μL which
257 represents the lowest detectable concentration in a calibration standard. The lipid adjusted
258 human milk concentrations of dioxins and DLCs were reported as pg/g lipid. The Toxic
259 Equivalent (TEQ) dlPCBs was calculated by multiplying the concentration (pg/g lipid) of each
260 congener by its Toxic Equivalency Factor (TEF) value. For PXDD/Fs and PBDD/Fs, tentative
261 TEQs were calculated using the assigned TEF values for polychlorinated dibenzo-p-
262 dioxins/furans (Van den Berg et al., 2006).

263

264 **2.5.2 Statistical Analyses**

265

266 Data for dioxins and DLCs in human milk were statistically analysed using the
267 Minitab 18 software package (Minitab, 2010). Descriptive statistics were performed on 21
268 human milk samples. Exploratory data analysis on histograms demonstrated that the levels of
269 all congeners of dlPCBs, PXDD/Fs and PBDD/Fs follow a lognormal distribution, thus the
270 geometric means were reported rather than the mean.

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287 **3.0 Results**

288

289 **3.1 Food consumption**

290

291 Questionnaire responses on food consumption included major dietary intake of seafood,
292 fish, meat and meat products, and dairy products. The majority of participants reported a similar
293 consumption pattern. Seafood consumed included shrimp, clams, mussels, snails, squid,
294 oysters, and lobsters. Fish types included: salmon, mackerel, tilapia, tuna, and dried herring.
295 Dairy products frequently consumed included eggs and milk.

296

297 **3.2 Characteristics of lactating mothers**

298

299 The subjects were lactating mothers aged between 21-32 years, at the time they gave
300 birth; the mean age was 24 years. The average body mass index (BMI), calculated from the
301 height and weight, was approximately 27.5 kg/m². The lipid concentrations measured in the
302 human milk samples ranged between 3.2-4.2% (with a mean concentration of 3.7% ± 0.27:
303 Table 3).

304

305 **3.3 Congener specific dioxins and DLCs identified**

306

307 A summary of results of the congener-specific concentrations of dlPCBs, 2,3,7,8-
308 PXDD/Fs and 2,3,7,8-PBDD/Fs in human milk samples, and TEQ concentrations are shown
309 in Tables 1, 2 and 3. Dioxin-like PCBs were detected in all 21 samples; all congeners detected
310 were above the limit of detection (LOD). Figure 1 summarizes TEQ contributions of dlPCBs,
311 PXDD/Fs and PBDD/Fs from each participant's human milk sample. An evaluation of the
312 relative percentages of each congener to the total concentration showed the highest percentage
313 mean contribution from PCB-156 (27%), and the highest percentage TEQ contribution from
314 PCB-126. PCBs 81 and 169 contributed the least (3.6%) to the sum of average dlPCBs; PCBs

315 105, 156 and 167 contributed the most (22-26%) towards the overall sum of dlPCB congeners.
316 Comparatively, in European countries, in addition to PCBs- 118, 167, and 105, PCB-156 was
317 reported as one of the congeners with the highest concentration in human milk (Chovancová et
318 al., 2011; Ulaszewska et al., 2011). These observations are in agreement with the
319 concentrations detected in the current study.

320

321 **3.4 Estimate of daily exposure of infants to TEQs from the consumption of human milk** 322 **containing DLCs.**

323

324 The concentrations of dioxins and DLCs in participants were within two orders of
325 magnitude; concentrations were in order of dlPCBs >> PBDD/Fs > PXDD/Fs in human milk.
326 The average total concentrations for 21 human milk samples ranged between 0.76 and 261
327 pg/g lw for all congeners studied (Tables 1-3). Percentage mean contributions from PXDD/Fs
328 showed 2,3-B-78-CDD amounted to the lowest contribution of 1.36%; 1,3-B-2,7,8-CDF
329 (30.9%) contributed the highest amount to the total mean PXDD/Fs. Percentage contributions
330 of furan concentrations to the total concentrations were dominated by 1,2,3,4,6,7,8-BDF and
331 2,3,7,8-BDF. The mean TEQ concentration calculated for dlPCBs was 1.67 pg WHO₂₀₀₅-
332 TEQ/g lipid (Table 1); tentative TEQs calculated for 2,3,7,8-PBDD/Fs and 2,3,7,8-PXDD/Fs
333 were 0.29 pg WHO₂₀₀₅-TEQ/g lipid and 0.26 pg WHO₂₀₀₅-TEQ/g lipid (Table 2),
334 respectively. The mean TEQ for the sum of dlPCBs, 2,3,7,8-PXDD/Fs and 2,3,7,8-PBDD/Fs
335 in 21 Ghanaian human milk samples was 2.23 pg TEQ/g lipid.

336

337 **3.5 Assessment of estimated daily intake of TEQs by infants from consumption of human** 338 **milk containing dlPCBs, PXDD/Fs and PBDD/Fs (2,3,7,8).**

339

340 A risk assessment for breastfed infants was determined using the baseline data (for
341 PXDD/Fs and PBDD/Fs) and dlPCBs presented in Table 3. Considering exclusive
342 breastfeeding (for a minimum of 6 months) as the only exposure source to infants, an evaluation

343 of the daily intake and body burdens of dlPCBs, PXDD/Fs and PBDD/Fs was conducted based
344 on the following assumptions:

345

346 1) An approximate/average volume of human milk consumed by an infant per day until
347 its first year was estimated to be 600 mL milk per day, assuming an approximate human
348 milk density of 1.03 g/mL (Focant et al., 2002; Schechter et al., 1998b).

349 2) An average body weight of an infant to be 7 kg (Focant et al., 2002).

350

351 To estimate the daily TEQ intake of dioxins and DLCs (pg TEQ/day) in human milk, the
352 equations below were used.

353

354 Daily TEQ intake (pg TEQ/day) = TEQ concentrations measured in human milk (pg TEQ/g) x

355
$$\frac{\text{Mass of breastmilk consumed (g)}}{\text{day}} \times \text{Lipid (\%)} \quad \text{Equation 2}$$

356
$$\text{Daily TEQ intake (pg TEQ/day) per kg} = \frac{\text{Daily TEQ intake (pg TEQ/day)}}{\text{Estimated average body weight of infant}} \quad \text{Equation 3}$$

357

358 **3.6 Discussions**

359 Maternal diet impacts concentrations of dioxins and furans in human milk. Responses to dietary
360 consumption on the questionnaire showed similarities in dietary pattern for all lactating
361 mothers, for this study; thus, a correlation between age, food type, congener type could not be
362 assessed. The prevalent food types (meat, fish, shellfish, dairy products sampled from
363 industrialized areas) consumed have been reported to show PXDD/F contributions from fish
364 (60%), shellfish (98%) and most edible terrestrial food products (Fernandes et al., 2014).
365 Reporting PXDD/F concentrations in food consumed, their congener profiles, and correlations
366 with specific food type by lactating mothers is beyond the scope of this study.

367

368 Table 4 summarizes the mean TEQ concentrations and infant estimated dietary intake
369 in human milk from different countries reported globally on dlPCBs, PCDD/Fs and PXDD/Fs
370 (Adu-Kumi et al., 2010; Deng et al., 2012; Focant et al., 2013; Focant et al., 2002; Harrison
371 et al., 1998; Ohta et al., 2004; Papke, 1999; Pratt et al., 2013; Schechter et al., 1994; van den
372 Berg et al., 2017; van Leeuwen et al., 2000b; Yang et al., 2002). Calculations on estimated
373 dietary intake for nursing infants from industrialized countries such as Belgium, United
374 States, United Kingdom, Germany, Sweden, China, France and Korea show that the
375 estimated dietary intake of dioxins and DLCs in human milk ranged between 24-145 pg
376 TEQ/kg bw/day, and was higher than the daily intake observed in Ghana. The estimated
377 mean dietary intake of dlPCBs, PBDD/Fs and PXDD/Fs in 600 mL human milk (per day) for
378 an infant of average weight, 7 kg, in this study is 6.5 pg TEQ/kg body weight (bw)/day; the
379 range of intake varied from 2.4 to 29.2 pg TEQ/kg bw/day. The TEQ concentration estimated
380 for infants in this study (6.5 pg TEQ/kg bw/day) does not include TEQ contributions from
381 PCDD/Fs. Using the upper end WHO TDI safety standard assigned value of 4 pg TEQ/kg
382 bw/day, the Agency for Toxic Substances and Disease Registry (ATSDR) reference dose
383 (RfD) standard of 1 pg TEQ/kg bw/day, and the United States Environmental Protection
384 Agency (0.7 pg TEQ/kg bw/day), 71% of infants' mean daily intake [dlPCBs + PXDD/Fs +
385 PBDD/Fs] calculated in this study exceeded the recommended standard values. From the
386 current study, a breakdown of contribution from dlPCBs to the daily estimated dietary intake
387 indicated an approximate range between 1-28 pg TEQ/kg bw/day. Sixty-seven percent of
388 infants were below the WHO estimated dietary intake of 4 pg TEQ/kg bw/day (upper end);
389 all infants exceeded the recommended estimated daily intake of 1 pg TEQ/kg bw/day for
390 dlPCBs. This indicates that although concentrations and TEQ concentrations detected from
391 infant dietary intake in Ghana are much lower than that reported globally, the risks of

392 Ghanaian infants to toxic dioxins and DLCs from human milk intake are still high during
393 periods of nursing.

394

395 The current concentrations, and TEQ concentrations of dlPCBs detected in this study
396 from first time lactating mothers are lower than the approximate baseline dlPCB mean
397 concentrations detected in 2008 in studies of Adu-Kumi et al. (2010): 3 pg TEQ/g lw for
398 individual human milk samples, and the calculated value of 2 pg TEQ/g lw for pooled human
399 milk in the WHO-UNEP global survey monitoring study in Ghana (Adu-Kumi et al., 2010;
400 van den Berg et al., 2017). Additionally, the TEQ concentration estimated for infants in this
401 study are for PBDD/Fs and PXDD/Fs; the overall TEQ concentration does not include TEQ
402 contributions from PCDD/Fs, which appear to be the major contributor to the total TEQs in
403 human milk in Ghana in studies of Adu-Kumi et al. (2010). A decrease in concentration of
404 dlPCBs over a period of 9 years can be noted, although increases of dioxins and DLCs
405 pollution have been reported in e-waste recycling areas in Accra (Tue et al., 2016; Wittsiepe
406 et al., 2015b).

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408 An assessment of tetra, penta and hexa PBDD/F and PXDD/F concentrations reported
409 in literature showed that a number of studies did not find measurable amounts to quantify in
410 human milk (Wiberg et al., 1992). Additionally, for studies which reported concentrations,
411 relatively low contributions to the total TEQ were observed (Croes et al., 2013; Pratt et al.,
412 2013). Typical concentrations detected (pg/g lipid) in pooled human milk samples in the
413 above studies ranged between non-detect to 0.7 pg/g lipid for 2,3,7,8-PBDD/Fs in Flanders in
414 studies of Croes et al. (2013). The mean concentrations of 2,3,7,8-PBDD/Fs and 2,3,7,8-
415 PXDD/Fs reported in studies of Pratt et al. (2013) in pooled Irish human milk ranged from
416 0.05 to 1.36 pg/g lipid, and 0.02-0.54 pg/g lipid, respectively (Croes et al., 2013; Pratt et al.,

417 2013). The sum of mean concentrations of 2,3,7,8- PXDD/Fs and 2,3,7,8-PBDD/Fs identified
418 in the current study were 0.76 and 2.1 pg/g lw, respectively; these are in agreement with the
419 relatively low concentrations reported in literature. However, in another study on 2,3,7,8-
420 PBDD/Fs and 2,3,7,8-PXDD/Fs detected in Japanese human milk, high concentrations of
421 2,3,7,8-PBDD/Fs were reported with an average concentration of 269 pg/g lipid; mean
422 PXDD/Fs reported ranged between 0.68 to 12 pg/g lipid (Ohta et al., 2004).

423

424 Contributions from PCDD/Fs were not determined in this study; however, an estimated
425 dietary intake was calculated using previously reported TEQ values for PCDD/F in human
426 milk samples from Ghana from studies of Adu-Kumi et al. (2010) and the WHO global
427 survey result on Ghanaian infants. Thus, assuming a 3.5% lipid content, and a daily intake of
428 600 mL human milk/day, an estimated dietary intake of 9 pg TEQ/kg bw/day [PCDD/Fs] is
429 expected for human milk consumed by nursing infants in 2008. In an absence of known
430 accidental or occupational exposure, the major source of dioxins and DLCs, and their
431 respective TEQ concentrations in human milk of lactating mothers are likely to be attributed
432 to consumption of dairy products, meat and meat products, fish and seafood. This is further
433 corroborated by the relatively low concentrations when compared to other global values. The
434 results from this study were compared against concentrations detected in serum samples
435 obtained from primiparous Ghanaians (Bruce-Vanderpuije et al., 2019b). The lipid corrected
436 total concentration of dlPCBs in human milk samples (261 pg/g lipid) was greater than for
437 concentrations reported in serum (77.7 pg/g lipid). The calculated TEQ concentration of
438 dlPCBs in human milk was slightly higher (1.24 pg/TEQ/g lipid: sera and 1.67 pg/TEQ/g
439 lipid: human milk). The difference in total concentrations were attributed to higher
440 concentrations of the more chlorinated PCBs in milk samples: namely 189, 157, 167 and 105.
441 Lipid corrected total concentration of PXDD/F in human milk samples (0.75 pg/g lipid) and

442 sera were comparable (0.90 pg/g lipid); however, the calculated TEQ concentration of
443 PXDD/F was greater in sera (0.26 pg/TEQ/g lipid: human milk and 0.50 pg/TEQ/g lipid:
444 sera). This difference was due to higher concentrations of the highly potent 2,3-B 7,8-CDD
445 detected in sera. Lipid corrected total concentrations of PBDD/F in human milk and sera
446 samples (2.05 pg/g lipid: human milk, sera: 2.39 pg/g lipid) were comparable. However, the
447 calculated TEQ concentration of PBDD/F was greater in sera (0.28 pg/TEQ/g lipid: human
448 milk and 0.49 pg/TEQ/g lipid: sera). This difference was due to higher concentrations of the
449 highly potent 2,3,7,8-BDD and 12,3,7,8-BDD detected in the sera. The data indicates the
450 potential for preferential accumulation of certain DLCs in human milk when compared to
451 sera. This is interesting considering that previous studies on PCB profiles in 15 different
452 tissue types from the same organisms have been found to be largely consistent (Megson et al.,
453 2018). However, it should be noted that samples from the milk and sera were obtained from
454 different participants at different times; therefore, further research is required to verify any
455 differences observed between the two matrices.

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468 **4.0 Conclusion**

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470 This first cumulative study on brominated and mixed halogenated dioxins and furans highlights
471 Ghanaian infants to be at risk from dietary intake of dlPCBs, PBDD/Fs and PXDD/Fs from
472 human milk; an average daily intake of DLCs in excess of the WHO recommended safety was
473 estimated. The sum of average concentrations observed in 21 Ghanaian human milk samples
474 were 261 pg/g lipid (dlPCBs), 2.05 pg/g lipid (PBDD/Fs), and 0.76 pg/g lipid (PXDD/Fs). The
475 mean TEQ concentration was 2.23 pg TEQ/g, with contributions of 75.6% from dlPCBs, 13.3%
476 from PBDD/Fs and 11.1% from PXDD/Fs. The calculated mean infant daily intake for dlPCBs
477 was 4.95 pg TEQ/kg bw/day; a total average daily intake of 18.30 pg TEQ/kg bw/day was
478 estimated when also considering existing PCDD/F and dlPCB data from previous Ghanaian
479 studies. The major contributors to the background concentrations of dioxins and DLCs and
480 their respective TEQ concentrations in human milk of Ghanaian lactating mothers can likely
481 be attributed to consumption of dairy products, meat and meat products, fish and seafood. The
482 exposure assessment performed indicates that all infants consuming human milk exceeded the
483 recommended standard intake of 1 pg TEQ/kg bw/day (as set by the ATSDR & WHO).
484 Provided an individual continually consumes dioxin and DLC-contaminated food from infancy
485 to adulthood, the amount of daily intake of dioxins and DLCs would highly exceed the safety
486 standards assigned by the ATSDR, the USEPA and the WHO. Given that a foetus/infant's
487 developmental growth is affected by nutrients consumed transplacentally and via human milk,
488 we suggest strategies by Ghanaian governing bodies be directed towards minimizing/reducing
489 PCB and dioxin intake through the food chain. Within this work consideration should also be
490 given towards brominated and mixed halogenated DLCs as well as the more traditional
491 chlorinated analogues.

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497

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503

504 **Financial Interest Declaration**

505 The authors declare no financial interests.

506

507 **Ethical Approval**

508 This study received ethical approval from the Ghana Health Service Ethics Review
509 Committee (ref: GHS-ERC 04/08/16) on February 16th, 2017.

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References

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522 Adu-Kumi, S., Malisch, R., Kotz, A., Kypke, K., Asante, K., Takahashi, S., Tanabe, S.,
523 Takasuga, T., Clarke, E., Weber, R., 2010. Levels of persistent organic pollutants (POPs) in
524 human breast milk samples from Ghana. *Organohalogen Compd.* 72, 1046-1049.

525 Ames, J., Warner, M., Brambilla, P., Mocarelli, P., Satariano, W.A., Eskenazi, B., 2018.
526 Neurocognitive and physical functioning in the Seveso Women's Health Study. *Environ. Res.*
527 162, 55-62.

528 Asante, K.A., Adu-Kumi, S., Nakahiro, K., Takahashi, S., Isobe, T., Sudaryanto, A.,
529 Devanathan, G., Clarke, E., Ansa-Asare, O.D., Dapaah-Siakwan, S., Tanabe, S., 2011. Human
530 exposure to PCBs, PBDEs and HBCDs in Ghana: temporal variation, sources of exposure and
531 estimation of daily intakes by infants. *Environ. Int* 37, 921-928.

532 Birnbaum, L.S., Staskal, D.F., Diliberto, J.J., 2003. Health effects of polybrominated
533 dibenzo-p-dioxins (PBDDs) and dibenzofurans (PBDFs). *Environ. Int* 29, 855-860.

534 Bruce-Vanderpuije, P., Megson, D., Jobst, K., Jones, G.R., Reiner, E., Sandau, C.D.,
535 Clarke, E., Adu-Kumi, S., Gardella Jr, J.A.J.S.o.T.T.E., 2019b. Background levels of dioxin-
536 like polychlorinated biphenyls (dlPCBs), polychlorinated, polybrominated and mixed
537 halogenated dibenzo-p-dioxins and dibenzofurans (PCDD/Fs, PBDD/Fs & PXDD/Fs) in sera
538 of pregnant women in Accra, Ghana. *Sci. Total Environ.* 673, 631-642.

539 Bruce-Vanderpuije, P., Megson, D., Reiner, E.J., Bradley, L., Adu-Kumi, S., Gardella,
540 J.A., 2019a. The state of POPs in Ghana- A review on persistent organic pollutants:
541 Environmental and human exposure. *Environ. Pollut.* 245, 331-342.

542 Centre for Diseases Control and Prevention, 2006. Laboratory procedure manual:
543 PCDDs, PCDFs, cPCBs and ortho-substituted PCBs, in: Health, E. (Ed.).

544 Chovancová, J., Čonka, K., Kočan, A., Sejáková, Z.S., 2011. PCDD, PCDF, PCB and
545 PBDE concentrations in breast milk of mothers residing in selected areas of Slovakia.
546 *Chemosphere* 83, 1383-1390.

547 Croes, K., Colles, A., Koppen, G., De Galan, S., Vandermarken, T., Govarts, E.,
548 Bruckers, L., Nelen, V., Schoeters, G., Van Larebeke, N., 2013. Determination of PCDD/Fs,
549 PBDD/Fs and dioxin-like PCBs in human milk from mothers residing in the rural areas in
550 Flanders, using the CALUX bioassay and GC-HRMS. *Talanta* 113, 99-105.

551 Dai, Q., Xu, X., Eskenazi, B., Asante, K.A., Chen, A., Fobil, J., Bergman, Å., Brennan,
552 L., Sly, P.D., Nnorom, I.C., Pascale, A., Wang, Q., Zeng, E.Y., Zeng, Z., Landrigan, P.J., Bruné
553 Drisse, M.-N., Huo, X., 2020. Severe dioxin-like compound (DLC) contamination in e-waste
554 recycling areas: An under-recognized threat to local health. *Environ. Int* 139, 105731.

555 Deng, B., Zhang, J., Zhang, L., Jiang, Y., Zhou, J., Fang, D., Zhang, H., Huang, H., 2012.
556 Levels and profiles of PCDD/Fs, PCBs in mothers' milk in Shenzhen of China: Estimation of
557 breast-fed infants' intakes. *Environ. Int* 42, 47-52.

558 Dewailly, É., Weber, J.-P., Gingras, S., Laliberté, C., 1991. Coplanar PCBs in human
559 milk in the province of Quebec, Canada: are they more toxic than dioxin for breast fed infants?
560 Bull. Environ. Contam. Toxicol. 47, 491-498.

561 Djien Liem, A., Furst, P., Rappe, C., 2000. Exposure of populations to dioxins and related
562 compounds. Food Addit. Contam. 17, 241-259.

563 Drover, S.S., Villanger, G.D., Aase, H., Skogheim, T.S., Longnecker, M.P., Zoeller,
564 R.T., Reichborn-Kjennerud, T., Knudsen, G.P., Zeiner, P., Engel, S.M., 2019. Maternal thyroid
565 function during pregnancy or neonatal thyroid function and attention deficit hyperactivity
566 disorder: A systematic review. Epidemiology 30, 130-144.

567 Fernandes, A.R., Mortimer, D., Wall, R.J., Bell, D.R., Rose, M., Carr, M., Panton, S.,
568 Smith, F., 2014. Mixed halogenated dioxins/furans (PXDD/Fs) and biphenyls (PXBs) in food:
569 Occurrence and toxic equivalent exposure using specific relative potencies. Environ. Int 73,
570 104-110.

571 Focant, J.-F., De Pauw, E., 2002. Fast automated extraction and clean-up of biological
572 fluids for polychlorinated dibenzo-p-dioxins, dibenzofurans and coplanar polychlorinated
573 biphenyls analysis. J. Chromatogr. B 776, 199-212.

574 Focant, J.-F., Fréry, N., Bidondo, M.-L., Eppe, G., Scholl, G., Saoudi, A., Oleko, A.,
575 Vandentorren, S., 2013. Levels of polychlorinated dibenzo-p-dioxins, polychlorinated
576 dibenzofurans and polychlorinated biphenyls in human milk from different regions of France.
577 Sci. Total Environ. 452, 155-162.

578 Focant, J.F., Pirard, C., Thielen, C., De Pauw, E., 2002. Levels and profiles of PCDDs,
579 PCDFs and cPCBs in Belgian breast milk.: Estimation of infant intake. Chemosphere 48, 763-
580 770.

581 Harrison, N., Wearne, S., Gem, M.G.d.M., Gleadle, A., Starting, J., Thorpe, S., Wright,
582 C., Kelly, M., Robinson, C., White, S., Hardy, D., Edinburgh, V., 1998. Time trends in human
583 dietary exposure to PCDDs, PCDDs and PCBS in the UK. Chemosphere 37, 1657-1670.

584 Jogsten, I.E., Hagberg, J., Lindström, G., van Bavel, B., 2010. Analysis of POPs in
585 human samples reveal a contribution of brominated dioxin of up to 15% of the total dioxin
586 TEQ. Chemosphere 78, 113-120.

587 Koopman-Esseboom, C., Morse, D.C., Weisglas-Kuperus, N., Lutkeschipholt, I.J., Van
588 Der Paauw, C.G., Tuinstra, L.G., Brouwer, A., Sauer, P.J., 1994. Effects of dioxins and
589 polychlorinated biphenyls on thyroid hormone status of pregnant women and their infants.
590 Pediatr. Res. 36, 468.

591 Kotz, A., Malisch, R., Kypke, K., Oehme, M., 2005. PBDE, PBDD/F and mixed
592 chlorinated-brominated PXDD/F in pooled human milk samples from different countries.
593 Organohalogen Compd. 67, 1540-1544.

594 Megson, D., Brown, T.A., O'Sullivan, G., Robson, M., Ortiz, X., Worsfold, P.J., Comber,
595 S., Lohan, M.C., Reiner, E.J., 2018. Changes to polychlorinated biphenyl (PCB) signatures and
596 enantiomer fractions across different tissue types in Guillemots. Mar. Pollut. Bull. 131, 174-
597 179.

598 Minitab, 2010. Statistical Software [Computer Software]. State College, PA: Minitab Inc.

599 Nakajima, S., Saijo, Y., Miyashita, C., Ikeno, T., Sasaki, S., Kajiwara, J., Kishi, R., 2017.
600 Sex-specific differences in effect of prenatal exposure to dioxin-like compounds on
601 neurodevelopment in Japanese children: Sapporo cohort study. *Environ. Res.* 159, 222-231.

602 Ohta, S., Okumura, T., Nishimura, H., Nakao, T., Shimizu, Y., Ochiai, F., Aozasa, O.,
603 Miyata, H., 2004. Levels of PBDEs, TBBPA, TBPs, PCDDs/DFs, PXDDs/DFs and
604 PBDDs/DFs in human milk of nursing women and dairy milk products in Japan.
605 *Organohalogen Compd.* 66, 2857-2862.

606 Olsman, H., Engwall, M., Kammann, U., Klempt, M., Otte, J., Van Bavel, B., Hollert,
607 H., 2007. Relative differences in aryl hydrocarbon receptor-mediated response for 18
608 polybrominated and mixed halogenated dibenzo-p-dioxins and-furans in cell lines from four
609 different species. *Environ. Toxicol. Chem.* 26, 2448-2454.

610 Papke, O., 1999. Background contamination of humans with dioxins and dioxin-like
611 PCBs. *Organohalogen Compd.* 44, 5-8.

612 Pratt, I., Anderson, W., Crowley, D., Daly, S., Evans, R., Fernandes, A., Fitzgerald, M.,
613 Geary, M., Keane, D., Morrison, J.J., 2013. Brominated and fluorinated organic pollutants in
614 the breast milk of first-time Irish mothers: is there a relationship to levels in food? *Food Addit.*
615 *Contam. A* 30, 1788-1798.

616 Schechter, A., Birnbaum, L., Ryan, J.J., Constable, J.D., 2006. Dioxins: An overview.
617 *Environ. Res.* 101, 419-428.

618 Schechter, A., Kassis, I., Pöpke, O., 1998a. Partitioning of dioxins, dibenzofurans, and
619 coplanar PCBS in blood, milk, adipose tissue, placenta and cord blood from five American
620 women. *Chemosphere* 37, 1817-1823.

621 Schechter, A., Pöpke, O., Lis, A., Ball, M., Ryan, J., Olson, J., Li, L., Kessler, H., 1996.
622 Decrease in milk and blood dioxin levels over two years in a mother nursing twins: estimates
623 of decreased maternal and increased infant dioxin body burden from nursing. *Chemosphere* 32,
624 543-549.

625 Schechter, A., Ryan, J.J., Pöpke, O., 1998b. Decrease in levels and body burden of dioxins,
626 dibenzofurans, PCBS, DDE, and HCB in blood and milk in a mother nursing twins over a
627 thirty-eight month period. *Chemosphere* 37, 1807-1816.

628 Schechter, A., Startin, J., Rose, M., Wright, C., Parker, I., Woods, D., Hansen, H., 1990.
629 Chlorinated dioxin and dibenzofuran levels in human milk from Africa, Pakistan, southern
630 Vietnam, the southern US and England. *Chemosphere* 20, 919-925.

631 Schechter, A., Startin, J., Wright, C., Kelly, M., Pöpke, O., Lis, A., Ball, M., Olson, J.,
632 1994. Dioxins in US food and estimated daily intake. *Chemosphere* 29, 2261-2265.

633 Sethi, S., Morgan, R.K., Feng, W., Lin, Y., Li, X., Luna, C., Koch, M., Bansal, R., Duffel,
634 M.W., Puschner, B., Zoeller, R.T., Lehmler, H.-J., Pessah, I.N., Lein, P.J., 2019. Comparative
635 analyses of the 12 most abundant PCB congeners detected in human maternal serum for activity
636 at the thyroid hormone receptor and ryanodine receptor. *Environ. Sci. Technol.* 53, 3948-3958.

637 Shaw, S.D., Berger, M.L., Harris, J.H., Yun, S.H., Wu, Q., Liao, C., Blum, A., Stefani,
638 A., Kannan, K., 2013. Persistent organic pollutants including polychlorinated and
639 polybrominated dibenzo-p-dioxins and dibenzofurans in firefighters from Northern California.
640 *Chemosphere* 91, 1386-1394.

641 Tanabe, S., Kunisue, T., 2007. Persistent organic pollutants in human breast milk from
642 Asian countries. *Environ. Pollut.* 146, 400-413.

643 Tue, N.M., Goto, A., Takahashi, S., Itai, T., Asante, K.A., Kunisue, T., Tanabe, S., 2016.
644 Release of chlorinated, brominated and mixed halogenated dioxin-related compounds to soils
645 from open burning of e-waste in Agbogbloshie (Accra, Ghana). *J. Hazard. Mater.* 302, 151-
646 157.

647 Tue, N.M., Katsura, K., Suzuki, G., Takasuga, T., Takahashi, S., Viet, P.H., Tanabe, S.,
648 2014. Dioxin-related compounds in breast milk of women from Vietnamese e-waste recycling
649 sites: Levels, toxic equivalents and relevance of non-dietary exposure. *Ecotoxicol. Environ.*
650 *Saf.* 106, 220-225.

651 Ulaszewska, M.M., Zuccato, E., Capri, E., Iovine, R., Colombo, A., Rotella, G.,
652 Generoso, C., Grassi, P., Melis, M., Fanelli, R., 2011. The effect of waste combustion on the
653 occurrence of polychlorinated dibenzo-p-dioxins (PCDDs), polychlorinated dibenzofurans
654 (PCDFs) and polychlorinated biphenyls (PCBs) in breast milk in Italy. *Chemosphere* 82, 1-8.

655 Van den Berg, M., Birnbaum, L.S., Denison, M., De Vito, M., Farland, W., Feeley, M.,
656 Fiedler, H., Hakansson, H., Hanberg, A., Haws, L., Rose, M., Safe, S., Schrenk, D., Tohyama,
657 C., Tritscher, A., Tuomisto, J., Tysklind, M., Walker, N., Peterson, R.E., 2006. The 2005 World
658 Health Organization re-evaluation of human and mammalian toxic equivalency factors for
659 dioxins and dioxin-like compounds. *Toxicol. Sci.* 93, 223-241.

660 van den Berg, M., Kypke, K., Kotz, A., Tritscher, A., Lee, S.Y., Magulova, K., Fiedler,
661 H., Malisch, R., 2017. WHO/UNEP global surveys of PCDDs, PCDFs, PCBs and DDTs in
662 human milk and benefit–risk evaluation of breastfeeding. *Arch. Toxicol.* 91, 83-96.

663 van Leeuwen, F.R., Feeley, M., Schrenk, D., Larsen, J.C., Farland, W., Younes, M.,
664 2000a. Dioxins: WHO’s tolerable daily intake (TDI) revisited. *Chemosphere* 40, 1095-1101.

665 van Leeuwen, F.X.R., Feeley, M., Schrenk, D., Larsen, J.C., Farland, W., Younes, M.,
666 2000b. Dioxins: WHO’s tolerable daily intake (TDI) revisited. *Chemosphere* 40, 1095-1101.

667 Victora, C.G., Bahl, R., Barros, A.J., França, G.V., Horton, S., Krasevec, J., Murch, S.,
668 Sankar, M.J., Walker, N., Rollins, N.C., 2016. Breastfeeding in the 21st century: epidemiology,
669 mechanisms, and lifelong effect. *The Lancet* 387, 475-490.

670 Wiberg, K., Rappe, C., Haglund, P., 1992. Analysis of bromo-, chloro- and mixed
671 bromochloro-dibenzo-p-dioxins and dibenzofurans in salmon, osprey and human milk.
672 *Chemosphere* 24, 1431-1439.

673 Wittsiepe, J., Fobil, J.N., Till, H., Burchard, G.-D., Wilhelm, M., Feldt, T., 2015a. Levels
674 of polychlorinated dibenzo-p-dioxins, dibenzofurans (PCDD/Fs) and biphenyls (PCBs) in
675 blood of informal e-waste recycling workers from Agbogbloshie, Ghana, and controls.
676 *Environ. Int.* 79, 65-73.

677 Wittsiepe, J., Fobil, J.N., Till, H., Burchard, G.-D., Wilhelm, M., Feldt, T., 2015b. Levels
678 of polychlorinated dibenzo-p-dioxins, dibenzofurans (PCDD/Fs) and biphenyls (PCBs) in
679 blood of informal e-waste recycling workers from Agbogbloshie, Ghana, and controls. *Environ*
680 *Int* 79, 65-73.

681 Wittsiepe, J., Fürst, P., Schrey, P., Lemm, F., Kraft, M., Eberwein, G., Winneke, G.,
682 Wilhelm, M., 2007. PCDD/F and dioxin-like PCB in human blood and milk from German
683 mothers. *Chemosphere* 67, S286-S294.

684 World Health Organization, 1996. Levels of PCBs, PCDDs, and PCDFs in human milk
685 (Environmental Health in Europe No. 3). WHO European Centre for Environment and Health,
686 Bilthoven.

687 World Health Organization, 1998. Assessment of the Health Risk of Dioxins:
688 Reevaluation of the Tolerable Daily Intake TDI, May, Geneva, Switzerland, pp. 25-29.

689 Yang, J., Shin, D., Park, S., Chang, Y., Kim, D., Ikonomou, M.G., 2002. PCDDs, PCDFs,
690 and PCBs concentrations in breast milk from two areas in Korea: body burden of mothers and
691 implications for feeding infants. *Chemosphere* 46, 419-428.

692