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Lesser Kestrel Foraging Habitats in Special Protection Areas in Agro-ecosystems

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Abstract: The European Union (EU) Wild Birds Directive recognises that the most serious threats to wild birds' conservation in Europe are habitat loss and degradation, and hence, habitats of threatened and migratory species must be protected with the establishment of the network of the special protection areas (SPAs) for migratory and endangered bird species in the EU member states. The major European population of the lesser kestrel *Falco naumanni*, a migratory falcon listed in Annex I of the Birds Directive, occurs in low-input farming systems in the Mediterranean basin, including Greece. The aim of this study was to identify foraging habitats of lesser kestrels and relate them to the delimited SPAs in the agro-ecosystems of Greece, where the stronghold of the species population for Greece occurs. Foraging habitat preferences were assessed using Poisson regression models (PRMs). SPAs were examined on whether they can effectively protect foraging habitats for breeding lesser kestrels in the study area. Foraging lesser kestrel abundance was positively associated with grasslands and non-irrigated land (dry cereals), and negatively associated with irrigated land (wet cotton), scrubland and woodland. Electricity facilities were used as foraging perches by lesser kestrels. The current SPAs cover a small percentage of the species' foraging sites and cannot be considered coherent enough to support and protect the foraging habitats of lesser kestrels and other priority species in the agro-ecosystems of the study area. Proposals for effective conservation of low-input farming systems, supporting priority species, are also presented.

Key words: *Falco naumanni*, special protection areas (SPAs), Poisson regression models (PRMs), foraging habitats, species conservation, agro-ecosystems, Greece.

1. Introduction

The European Union (EU) Wild Birds Directive (Council Directive 2009/147/EEC) recognises that the most serious threats to the conservation of wild birds in Europe are habitat loss and degradation, focusing on the protection of the habitats of endangered and migratory species [1]. Maintenance of the ecological quality of the habitats of priority species, listed in Annex I of the Birds Directive, is a key conservation issue for the preservation of bird diversity in the EU. A network of special protection areas (SPAs) for priority species has been identified and delimited in most EU member states, comprising of wild birds'

most suitable breeding and foraging sites, aiming at improving species conservation status at European level [1, 2].

The major European population of a protected migratory bird species of lowland Europe, the lesser kestrel *Falco naumanni*, occurs in low-input farming systems in the Mediterranean basin [3, 4]. Although the lesser kestrel has been down-listed from "vulnerable" to "least concern" in the IUCN Red List of Threatened Species since 2011 [5], it is still an Annex I species of the EU Birds Directive due to its dramatic decline in recent years [1]. Many important lesser kestrel habitats have been designated as SPAs of the Natura 2000 network in EU member states where it breeds [4, 6]. Based on BirdLife International classification list on species of European conservation

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concern (SPEC), the lesser kestrel is a “SPEC 1” species [7] with 25%-49% of its world population breeding in Europe [8]. At European level, it is a priority species in steppic habitats (i.e., primary steppe, secondary steppe and pseudo-steppe of extensively farmed, mixed rotational systems of grassland, cereal, fodder crops and grazed fallow land) and in arable land and pastures (i.e., land regularly ploughed and/or cultivated for feed and non-feed crops); and its survival is linked with management practices in agro-ecosystems [9]. In the Mediterranean region, the species population had undergone a sharp decline, but it has now recovered in some countries, such as Portugal [4]. In Greece, the species was common in the 1960s, but it also suffered a dramatic decline and its distribution shrunk; it is legally protected and listed as “vulnerable” in the Red Data Book of Threatened Animals of Greece [10]. Nowadays, its population has recovered in mainland Greece and is estimated at 5,400-7,100 breeding pairs [11].

Lesser kestrel habitat selection studies reveal that its populations decline results from changes in agricultural practice and loss or deterioration of its foraging habitats on its breeding grounds [12-21]. However, the criteria for delimiting SPA boundaries for protected birds are not always appropriate, as they are often based solely on nesting distribution [22]. Conservation actions for priority birds should take into account both nesting and foraging requirements and only recently studies have focused on the effectiveness of SPAs for EU priority species [22-29]. Currently, the lesser kestrel is a priority species in 25 designated SPAs within the protected areas of the network of Natura 2000 sites of Greece, with more than 90% of its population occurring in the agro-ecosystems of central Greece [30, 31].

The aim of this study was to identify important foraging habitats of breeding lesser kestrels and relate those to delimited SPAs in the agro-ecosystems of central Greece, where the stronghold of its breeding

population for Greece occurs. Questions on: (1) the lesser kestrel foraging habitat preferences, (2) the information obtained on environmental variables at different spatial extents and incorporated in foraging habitat analysis for conservation purposes and (3) the coherence of SPAs on whether they can effectively protect foraging habitats along with nesting sites for breeders in the study area are assessed and proposals for effective management are presented.

2. Materials and Methods

2.1 Study Area

The study area is located in the largest agricultural plain of the country, covering approximately 4,000 km², in the region of Thessaly, central Greece (Fig. 1). Nearly half of the plain is dominated by cultivations of intensive irrigated cotton and non-intensive dry cereal, while pastures are on hilly slopes close to urban areas [32]. The climate is typical continental Mediterranean, characterised by wet, cold winters and dry, hot summers [33]. The elevation in the study area ranges in 0-2,005 m. Five SPAs established in the region of Thessaly include the lesser kestrel as a priority species [31, 32]. Three SPAs “Periochi Thessalikou Kampou” (GR1420011), “Periochi Farsalon” (GR1420012) and “Oros Ossa” (GR1420007) are in the study area (Fig. 1). Two SPAs “Oros Mavrovouni” (GR1420006) and “Periochi Tyrnavou” (GR1420013) are not included in study area, located at its periphery (Fig. 1). The SPAs include urban areas (towns and villages) with lesser kestrel colonies, the presence of which in the study area was mapped in the years 2006 and 2007 [34] (Fig. 1). Other priority species in the SPAs of the study area include the short-toed eagle *Circaetus gallicus*, the long-legged buzzard *Buteo rufinus*, the lanner falcon *Falco biarmicus*, the calandra lark *Melanocorypha calandra* and the stone curlew *Birhonus oedinemus*.

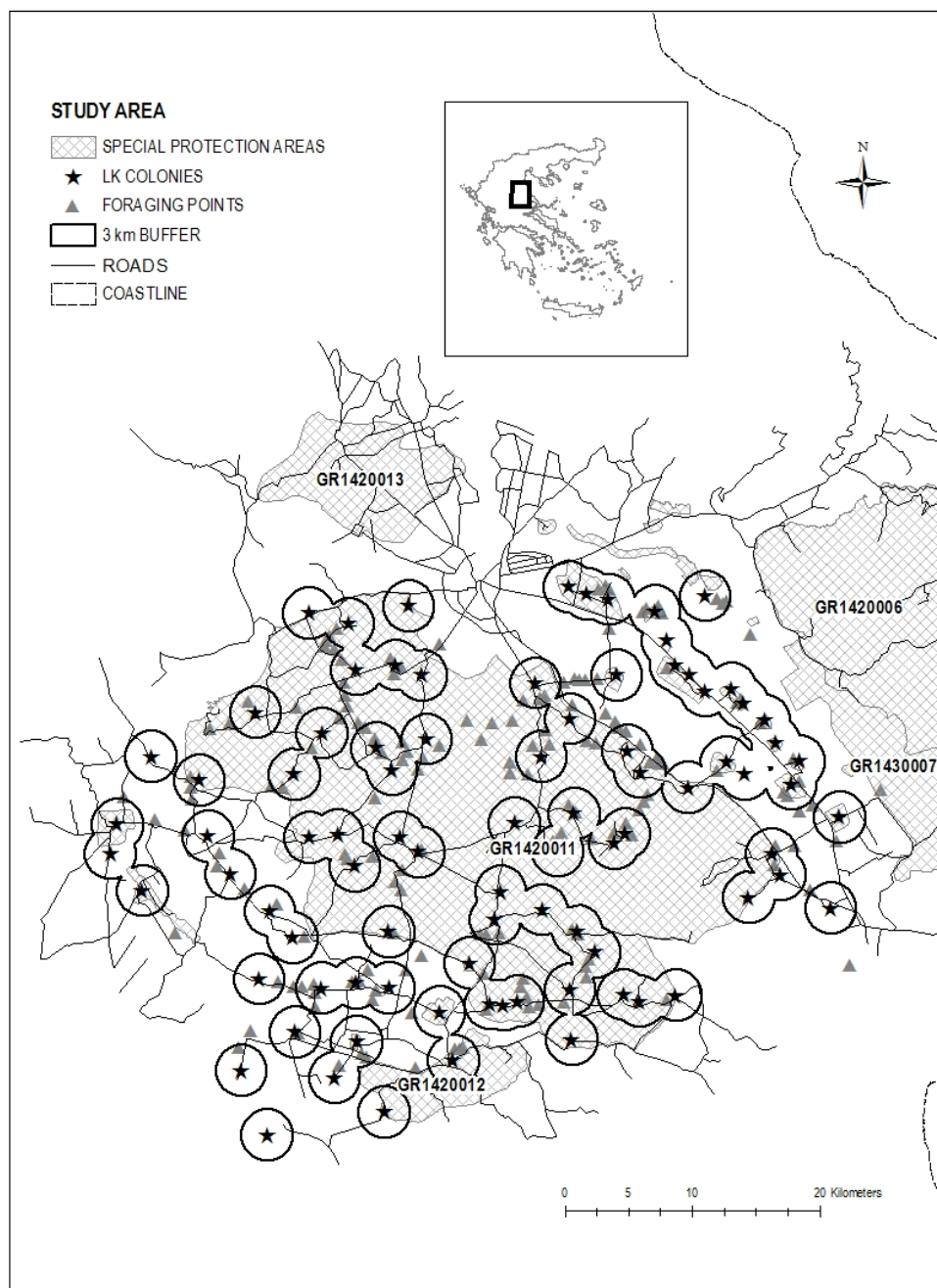


Fig. 1 Map of the study area and the SPAs (cross-hatched areas).

The SPAs GR1420011 (comprised of a larger and many smaller areas), GR1420012 and GR1420007 lie within the study area, while the SPAs GR1420006 and GR1420013 are located at its periphery.

Black stars: the lesser kestrel colonies in the study area; grey triangles: the foraging observation points.

Thick black lines: buffer zones of 3 km radius around each colony (dissolved between them for overlapping colonies); thin black lines: the network of main roads.

2.2 Bird Data

Data on foraging birds were collected along road transects during June and July of 2007. Road counts

can be used for large and obvious species, such as raptors, particularly those that hover when searching for food in open habitats [35, 36]. Road transects can be limited by road availability [37], but the study area

is open and has a smooth terrain with less tall vegetation, allowing visibility at long distances; counts were made on both sides of the road. Highways were avoided due to speed limitations. Transects involved slowly driving of approximately 30 km/h in days of good weather conditions for the detection of birds and frequent stops both when birds were seen but also to search for birds from vantage points [14]. Twenty road transects, on accessible routes of the main and the secondary roads and of the farm tracks in the study area, were randomly selected, covering a length of approximately 2,000 km, and were driven once. The network of main roads in the study area is presented in Fig. 1. Surveys were conducted from sunrise until sunset, considering the fact that lesser kestrels can be active throughout the day [12, 38]. Only those birds considered as “foraging individuals” i.e., hovering over a habitat or landing to catch a prey from the ground or hunting from a perch were recorded as a “foraging observation point” and the position was marked with a global positioning system (GPS) receiver, from the position of the researcher, as close as possible to bird location. The habitat type was identified in the field and each foraging bird was attributed to the habitat type where it was first observed [39].

2.3 Environmental Data

To obtain information for lesser kestrel foraging habitat analysis at different scales, sampling plots at two spatial extents around each “foraging observation point” were selected for data retrieval: (1) at a circular buffer of 100 m radius and (2) at a circular buffer of 500 m radius. The selection of the extents was based on the determination of the main components of spatial scale which are the “grain size” and “extent size”; according to Legendre, P. and Legendre, L. [40], the former is “the size of the elementary sampling unit” and the latter is the overall area from where observations are made, i.e., extent size is the whole study area [41]. In this study, the grain size was

selected at two extents, the sampling plots of 100 m and 500 m plots. This information could be used for conservation purposes, for example, for the establishment of effective agri-environmental measures (AEM) [42] and the determination of ecological focus areas (EFA) that should be identified and maintained in large size arable cultivations under the greening of the current EU Common Agricultural Policy (CAP) [43].

Habitat data—used as explanatory variables in the model development—were extracted from a thematic map, produced by a supervised image classification analysis of a Landsat satellite image of 30 m pixel size resolution [34], obtained from the U.S. Geological Survey for the year 2006 [44]. Seven land cover types (non-irrigated farmland, irrigated farmland, grassland, urban, woodland, scrubland and water) were determined, based on the land cover classes of the EU Programme Corine 2000 [45]. Elevation variables (min, max, mean and standard deviation describing topographic complexity) were derived from a digital elevation model (DEM) of 90 m pixel size resolution [46] and analyses were performed in ArcGIS [47]. Other researchers have also included elevation variables in lesser kestrel foraging habitat analyses [15].

Other information recorded at foraging observation points includes presence of electricity facilities and field margins when used by birds and were incorporated in the models as categorical explanatory variables. These attributes have been identified as important elements in the vicinity of nesting colonies in the European Species Action Plan [4]. A variable on the “nearest distance-to-colony” (from each foraging observation point) was also calculated using Hawth’s tool (v. 3×) [48]. In total, 13 explanatory variables were used for model development (Table 1).

2.4 GIS-Based Analysis

GIS-based analysis was used to: (1) examine overlapping of the SPAs with the breeding colonies and

Table 1 Explanatory variables used for the development of foraging habitat models.

Variable	Description and units
Urban	Urban and other built-up areas and artificial surfaces, such as roads, airports, etc. (%) [*]
Irrigated	Irrigated agricultural land, dominated by cotton fields and other industrial plants (maize, tobacco) (%) [*]
Non-irrigated	Non-irrigated agricultural land, dominated by dry cereals (mainly wheat) (%) [*]
Grassland	Grasslands, pastures and fallow land (%) [*]
Scrubland	Sclerophyllus vegetation (garrigue and short maquis) (%) [*]
Woodland	Forest, tall maquis and areas of woody crop plantations and tree groves (%) [*]
Elevation min	Minimum value of elevation within each polygon (m) [*]
Elevation max	Maximum value of elevation within each polygon (m) [*]
Elevation mean	Mean value of elevation within each polygon (m) [*]
Elevation SD	Standard deviation of elevation; a measure of topographic complexity (m) [*]
Distance_colony	Distance of each foraging point to the nearest colony (m)
Dummy_fmargins	Dummy of field margins; Coded as 1/0; 1 corresponds to field margins as foraging habitat and 0 to other foraging habitats
Dummy_elecwires	Dummy of electricity facilities, such as wires and poles; Coded as 1/0; 1 corresponds to “presence” and 0 to “absence” of electricity wires and poles that birds use as foraging perches

^{*} Measured at two different spatial extents within each sampling plot.

the lesser kestrel foraging sites and (2) obtain information on the habitat types in the SPAs. Spatial data extraction: (1) the extent of the current SPAs and (2) buffer zones of 3 km radius around each breeding colony (buffers dissolved for overlapping colonies) were respectively overlaid with: (1) the lesser kestrel foraging observation points, (2) the species breeding colonies and (3) the habitat thematic map. The 3 km radius was selected based on suggestion of the European Species Action Plan, according to that “all nesting locations must provide access (within range 1-3 km) to open areas for hunting, usually in steppe-like habitats, natural or managed grasslands and non-intensively cultivated land” [4]. In this study, the maximum distance of 3 km radius was selected.

2.5 Statistical Analysis

Descriptive statistics were calculated for all explanatory variables in the sampling plots at both spatial extents (100 m and 500 m). Dummy variables were used to express the categorical variables for the presence of electricity facilities and field margins (Table 1); for explanatory variables with c categories, $c-1$ dummy variables are needed and a reference

category is set in which all dummy variables equal zero [49]. The dummy on electricity facilities was coded as “1” for birds using them as foraging perches and as “0” if not used. The dummy on field margins was coded as “1” when birds were recorded to forage there and “0” when birds foraged elsewhere. A principal components analysis (PCA) was applied to remove problems from correlated variables [50]. The principal components (PC_s) were used as explanatory variables in the model building. Components were obtained with the varimax rotation [50] and only factors with eigenvalues greater than one were included in the analysis.

For the assessment of lesser kestrel foraging habitat selection, Poisson regression models (PRMs), i.e., generalized linear models for count data [51] were developed with bird count data from the foraging observation points as the response variable and a set of explanatory variables. Residual deviance was first checked for overdispersion [52]. Because overdispersion was detected, Quasi-Poisson models (QPMs) were employed, where the variance was given as function of the mean using a dispersion parameter ρ (i.e., variance = $\rho \times \mu$) [53]. All analyses were performed using R (v.2.8.1) [54].

3. Results

3.1 Bird Data—Univariate Analysis

In total, 760 lesser kestrels were recorded along the road transects, at 166 foraging observation points (Fig. 1). Higher aggregations of foraging birds were observed in cereal fields accounting for 77.7% of all individuals (Table 2). About 15% of birds were recorded to feed in fallow land and grassland and only a few in cotton fields (3.2%) and in field margins (2.9%) (Table 2). Mean coverage of irrigated land was 20.4% at the 100 m and 32.8% at the 500 m extent plots, respectively, while non-irrigated land occupied 37.7% of the area at the 100 m and 29.0% of the area at the 500 m plots, respectively (Table 3). Urban areas covered large parts of plots at both extents (Table 3). Both irrigated land and woodland occupied larger areas at the 500 m compared to the 100 m plots and topographic complexity (standard deviation of elevation) was also higher in the 500 m plots than the 100 m plots (Table 3). The mean distance of foraging observation points to the colonies was approximately 1,645 m (Table 3).

3.2 Principal Components Analysis

At the 100 m extent analysis, four principal components (PC1a-PC4a) were extracted accounting for 75.5% of the total explained variance within the original variables. PC1a was an “elevation” component (Table 4). PC2a was a “habitat and distance” component, referring to scrubland and non-irrigated

land; the latter located away from colonies (Table 4). PC3a was a “habitat” component (grasslands located away from irrigated land and woodland), while PC4a was an “urban and habitat” component with urban areas located away from grasslands (Table 4). At the 500 m extent analysis, four components (PC1b-PC4b) were also extracted accounting for 80.7% of the original variance. PC1b (Table 5) was an “elevation and habitat” component, including elevation and non-irrigated and irrigated farmland, hilly areas covered with cereals, away from cotton fields. PC2b was also a “habitat” component, describing areas of woodland and irrigated land, away from urban sites (Table 5). PC3b was a mixed “habitat and distance” component, referring to areas covered with scrubland situated away from colonies (Table 5), while PC4b was a “habitat” component, positively related to grasslands (Table 5).

3.3 Quasi-Poisson Models

The present study aimed at obtaining information from two different scales on the environmental variables; no comparison of the models performance was conducted. At the 100 m extent analysis, in QPM1, significant variables were those components related with agricultural and natural habitats and the nearest distance to the colonies; the PC2a was positively and the PC3a was negatively associated with lesser kestrel abundance (Table 6). Also, at the 100 m extent, the elevation component (PC1a) was significant. At the 500 m extent analysis, in QPM2, significant components were the elevation and

Table 2 Number of foraging observation points (FOP) recorded at the habitat types as identified in the field and number of birds (%) recorded at these points.

Habitat type (identified in the field) of FOP	Number of FOP	Number of birds recorded at FOP	Percentage of birds at each habitat type (%)
Alfalfa field	3	7	0.9
Cereal field	105	591	77.7
Cotton field	19	24	3.2
Fallow land	26	83	10.9
Field margin	8	22	2.9
Grassland	4	31	4.1
Tomato field	1	2	0.3
All	166	760	100.0

Table 3 Descriptive statistics with mean values and standard deviations (SD) at the 100 m and 500 m extents for the explanatory variables used in the models.

Predictor (units)	100 m extent		500 m extent	
	Mean	SD	Mean	SD
Urban (%)	14.7	15.5	14.4	8.9
Irrigated (%)	20.4	27.7	32.8	22.8
Non-irrigated (%)	37.7	33.9	29.0	22.8
Grassland (%)	13.4	20.1	10.7	12.2
Scrubland (%)	12.8	15.5	11.6	8.2
Woodland (%)	0.8	2.2	1.1	1.2
Elevation min (m)	133.1	65.4	122.7	56.1
Elevation max (m)	138.7	69.4	152.2	81.2
Elevation mean (m)	135.8	67.3	135.4	66.2
Elevation SD (m)	2.4	2.4	6.9	7.8
Distance_colony (m)	1,644.9	990.6	1,644.9	990.6

Table 4 Loadings of the PCs, extracted from the PCA, and their percentage of variance, at the 100 m extent.

Predictor (% of variance)	PC1a* (35.1)	PC2a* (16.3)	PC3a* (13.1)	PC4a* (10.9)
Urban			+	0.873
Grassland		+	0.585	-0.555
Irrigated	-		-0.671	-
Woodland		+	-0.655	
Non-irrigated	+	-0.719		
Scrubland		0.770		
Distance_colony		-0.587		
Elevation min	0.974			
Elevation max	0.979			
Elevation mean	0.977			
Elevation SD	0.760		+	

* Only loadings larger than 0.5 are shown; for loadings with values between 0.2 and 0.5 only the sign is shown.

PC1a is an elevation component, PC2a a habitat component, PC3a a habitat-distance component, PC4a a habitat component.

Table 5 Loadings of the components extracted from the PCA, and its percentage of variance, at the 500 m extent.

Predictor (% of variance)	PC1b* (40.0)	PC2b* (14.7)	PC3b* (13.4)	PC4b* (12.5)
Woodland		0.732		
Urban	-	-0.702	+	
Irrigated	-0.621	0.570		-
Non-irrigated	0.640	-	-	-
Grassland				0.917
Elevation min	0.957			
Elevation max	0.976			
Elevation mean	0.974			
Elevation SD	0.837			+
Scrubland		+	0.669	+
Distance_colony			-0.794	

* Only loadings larger than 0.5 are shown; for loadings with values between 0.2 and 0.5 only the sign is shown.

PC1b is a habitat-elevation component, PC2b a habitat component, PC3b a habitat-distance component, PC4b a habitat component.

Table 6 Model coefficients and their significance for QPM1 at 100 m extent analysis.

Predictor	Coefficient	Standard error	<i>t</i> value	Pr (> <i>t</i>)
Intercept	1.404	0.129	10.885	0.000***
Dummy_fmargins	0.013	0.292	0.047	0.962
Dummy_elecwires	-0.307	0.148	-2.061	0.040*
PC1a	0.120	0.064	1.865	0.064.
PC2a	-0.145	0.061	-2.365	0.019*
PC3a	0.222	0.061	3.614	0.000***
PC4a	-0.001	0.060	-0.018	0.985

Significance level: 0 '***', 0.001 '**', 0.01 '*', 0.05 '.', 0.1 ' ' 1.

Table 7 Model coefficients and their significance for QPM2 at 500 m extent analysis.

Predictor	Coefficient	Standard error	<i>t</i> value	Pr (> <i>t</i>)
Intercept	1.452	0.132	11.001	0.000***
Dummy_fmargins	-0.051	0.300	-0.173	0.863
Dummy_elecwires	-0.366	0.152	-2.410	0.017*
PC1b	0.122	0.066	1.835	0.068.
PC2b	-0.116	0.063	-1.841	0.067.
PC3b	-0.102	0.062	-1.633	0.104
PC4b	0.093	0.062	1.501	0.135

Significance level: 0 '***', 0.001 '**', 0.01 '*', 0.05 '.', 0.1 ' ' 1.

agricultural and natural habitat components; PC1b was positively and PC2b was negatively associated with lesser kestrel abundance, respectively (Table 7). Presence of electricity facilities was a significant explanatory variable for foraging lesser kestrel abundance at both extents, while presence of field margins was not a significant predictor in any of the models (Table 7).

3.4 Nesting and Foraging Lesser Kestrels in the SPAs

Based on the GIS-based analysis of the SPAs under their current extent, the 25% of the colonies and approximately 40% of foraging points are outside the SPAs (Fig. 1). The SPA GR1420011 covers very small areas around bird colonies in eastern and western sites. Colonies situated at the western boundaries of the study area are not included in the SPAs at all (Fig. 1). When the boundaries of the SPAs are extended, based on the 3 km radius buffer zone around each breeding site, they include (except for all colonies) 86% of the foraging observation points (Fig. 1). Concerning habitat type coverage, the current

SPAs include 14% of irrigated land, 27% of non-irrigated land and 11% of grasslands and fallow land. When the SPAs boundaries are extended by the 3 km buffer zone around colonies, they include 35% of irrigated, 26% of non-irrigated land and 11% of grasslands and fallow land.

4. Discussion

4.1 Foraging Habitat Selection of Lesser Kestrels

The foraging habitat analysis demonstrated that abundance of foraging lesser kestrel was positively associated with non-irrigated agricultural land (mostly cereals) and natural habitats (grasslands) in the study area. Based on outcomes of foraging observation points, cereal fields support more birds than any other habitat type. Similar studies in the Western Mediterranean found that cereals are a highly preferred foraging habitat by lesser kestrels [12-14, 16, 18, 19, 21, 38, 55]. Grasslands (including pastures and fallow land) are also considered as an optimal foraging habitat for breeding lesser kestrels in the Iberian Peninsula [12, 14, 19, 56]. This is because

they are rich in prey availability, mainly insects, including Orthoptera, Coleoptera, etc. [12, 57]. In diet analyses studies, Orthoptera were identified as lesser kestrel's most favourable prey items [16, 17].

Foraging lesser kestrel abundance was negatively associated with irrigated land (mainly cotton), a habitat type that covers large areas in the study area [34]. Other studies highlight that birds do not forage on irrigated crops [18, 58], while Tella et al. [13] remarked that birds had very large home ranges in areas with irrigated fields compared to those smaller ones, in cereal cultivations. Cotton fields are poor habitats in prey because of the high agro-chemical inputs that make them toxic to invertebrate fauna [59]. Ursua et al. [60] found that field margins are important feeding habitats for breeding lesser kestrels in areas surrounded by irrigated farmland. However, selection of field margins as preferred foraging habitats for lesser kestrels was not documented in this study, probably due to the time of the year field work was conducted (see below). Yet, field margins are widely recognised as a significant foraging habitat for the species, supporting their preferred prey [12, 13, 19, 57, 60].

Foraging lesser kestrel abundance was negatively associated with scrubland and woodland. Researcher argue that in habitats with dense vegetation cover (i.e., scrubland, woodland, tree plantations and olive orchards), prey is likely to be less accessible or scarce [12] and birds avoid them [13, 14, 17, 20, 56, 60]. Besides, foraging range has been found to be larger in areas with inappropriate habitats, such as irrigated fields, forest and scrubland, compared to areas with non-intensive cultivations and birds make longer foraging trips to search for food [13, 18, 21]. On the contrary, birds are known to forage close to colonies (2-3 km) in areas with good quality habitats and during the chick rearing period [13, 14, 21]. The mean distance of foraging observation points to the colonies in the study area was relatively small (1,645 m), indicating good foraging habitats.

This study was conducted during chick rearing period, which could explain why cereals were used for foraging. It has been found that in summer, during chick rearing period, cereal stubble are rich in Orthoptera and are highly preferred by birds [13, 38, 55]. When cereals are harvested, birds follow the harvest machines and capture insects that can be easily seen and accessed [14, 34]. Thus, selection of certain habitats for feeding by the lesser kestrel in cultivated areas can be determined by the time the study is conducted, depending on vegetation structure that affects accessibility and abundance of their prey [12, 17, 38]. For example, in spring, cereal fields are dense, plants are tall and hunting of insects is difficult and this habitat is avoided by lesser kestrels and birds may use field margins to feed [56, 60]. Further investigation would be needed for identifying the importance of field margins in the study area at different periods during the breeding season.

Abundance of foraging lesser kestrels was also associated with the presence of electricity facilities. During field surveys at foraging observation points, electricity facilities were reported to be present at the approximately 75% of the feeding areas in the study area [34]. In areas lacking wires, birds were recorded to stand on bales of hay, the only prominent locations that could be used to search for prey. Presence of electricity wires and trees near colonies seems favourable for birds, particularly in the post-fledging and pre-migratory periods for roosting and resting [19, 61]. Zank and Kemp [62] found that perch-hunting in lesser kestrels was more successful than hover-hunting at non-breeding grounds, in South Africa, possibly because birds could make more accurate strikes and spend more time on assessing prey.

4.2 Designation of SPAs and Foraging Lesser Kestrel Habitats

The designated SPAs in the agro-ecosystems of the study area include the bulk of breeding lesser kestrel

population in Greece, covering colonies surrounded partially by non-irrigated farmland (mainly cereals), the extent of which has dramatically shrunk in Central Greece over the last decades [34]. This study indicated that this land cover type remained of a preferred foraging habitat for the species. The significance of non-intensive cultivations also referred as high nature value (HNV) farmland (i.e., agro-ecosystems rich in biodiversity such as cereal cropping and semi-natural grasslands that sustain species of European and/or national conservation concern [63, 64]) is illustrated in the European Action Plan of the lesser kestrel [4]. Schemes for maintaining and enhancing HNV farmland, both within and outside the network of protected areas would be needed for the preservation not only of priority species occurring in these habitats but for the overall farmland biodiversity, contributing to the target 2 (maintain and restore ecosystems) and the target 3 (achieving more sustainable agriculture and forestry) of the EU Biodiversity Strategy to 2020 [65].

The current SPAs do not cover all nesting sites and cannot be considered coherent enough to safeguard preferred foraging habitats for lesser kestrels, in order to improve their favourable conservation status as required by the European Species Action Plan [4]. First, the SPAs have small extent around lesser kestrel colonies in the eastern range of the species population, while all colonies in the western range of the population are left without protection. Moreover, the SPAs do not include good quality feeding sites, as they are comprised by the one third of lesser kestrel preferred feeding areas. As the colonies, mostly in recent years, are surrounded by inappropriate foraging habitats, such as intensively irrigated cultivations [34], a potential expansion of the current SPAs would cover all colonies but, in terms of foraging habitats, it would only increase the percentage of cotton within them. Coverage of preferred feeding habitats would not increase and other conservation actions would therefore be needed, such as the establishment of that

hold elements, such as field margins, hedges and trees, fallow land and buffer strips under the EU CAP greening policy [43]. Rodriguez and Wiengand [66] recommend that restoration of field margins and sowing in-field strips can enhance biodiversity and improve the quality of foraging habitats in arable land.

Studies in the Iberian Peninsula also highlight inconsistency of the Natura 2000 network related to agro-ecosystems, showing that lesser kestrels are not well protected in farmlands that are under-represented in the network of protected areas in Spain [27]. While Traba et al. [25] showed that the SPAs network, in Spain, does not adequately cover the most important areas for steppe species (including lesser kestrels). Moreover, Santana et al. [28], who studied the effectiveness of conservation funding in Natura 2000 sites in Portugal for steppe birds (including lesser kestrels), found that only the specialised species were favoured, while the wider biodiversity remained under-protected. Guixé and Arroyo [22] who studied Montagu's harriers *Circus pygargus*, a species that also uses agricultural land to forage, suggest that conservation management should include larger radius around colonies to protect not only breeding but also feeding sites, as SPAs might include more inappropriate than appropriate foraging habitats. Underestimation of SPAs in agro-ecosystems is, thus, a more broad issue and should be addressed in a wider framework of protection for breeding lesser kestrels and other farmland birds in these ecosystems.

5. Conclusions

This study showed that cereals, a non-intensive cultivation and grasslands were highly preferred foraging habitats by lesser kestrels in the SPAs in agro-ecosystems of Central Greece. However, the current extent of the SPAs covers a small percentage of the species' foraging sites and cannot be considered coherent enough to support and protect the foraging habitats of lesser kestrels and other priority species in the study area. Proposals for efficient lesser kestrel

and other priority species conservation in the agro-ecosystems of Greece should include actions, such as: (1) reconsideration of existing SPAs extent and better designation of new ones to encompass good quality foraging habitat and incorporate small or isolated populations in marginal areas of the species breeding ranges and (2) preservation of low-input farming systems and HNV farmland areas. Local stakeholders could benefit through the establishment of specific agri-environmental measures and effective EFA in arable land within the framework of the Greek Rural Development Programme under the CAP that would provide farmers with extra income when involved in the preservation of biodiversity in Natura 2000 sites or with compensations for potential losses due to the enforcement of protection measures in agro-ecosystems.

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