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Mapping Woody Cover in Semi-arid Savannahs using Multi-seasonal Composites from Landsat Data

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Abstract

Increasing attention is being directed at mapping the fractional woody cover of savannahs using Earth-observation data. In this study, we test the utility of Landsat TM/ETM-based spectral-temporal variability metrics for mapping regional-scale woody cover in the Limpopo Province of South Africa, for 2010. We employ a machine learning framework to compare the accuracies of Random Forest models derived using metrics calculated from different seasons. We compare these results to those from fused Landsat-PALSAR data to establish if seasonal metrics can compensate for structural information from the PALSAR signal. Furthermore, we test the applicability of a statistical variable selection method, the recursive feature elimination (RFE), in the automation of the model building process in order to reduce model complexity and processing time. All of our tests were repeated at four scales (30, 60, 90, and 120 m-pixels) to investigate the role of spatial resolution on modelled accuracies.

Our results show that multi-seasonal composites combining imagery from both the dry and wet seasons produced the highest accuracies ($R^2=0.77$, RMSE=9.4, at the 120 m scale). When using a single season of observations, dry season imagery performed best ($R^2=0.74$, RMSE=9.9, at the 120 m
resolution). Combining Landsat and radar imagery was only marginally beneficial, offering a mean relative improvement of 1% in accuracy at the 120 m scale. However, this improvement was concentrated in areas with lower densities of woody coverage (<30%), which are areas of concern for environmental monitoring. At finer spatial resolutions, the inclusion of SAR data actually reduced accuracies. Overall, the RFE was able to produce the most accurate model ($R^2=0.8$, RMSE=8.9, at the 120 m pixel scale). For mapping savannah woody cover at the 30 m pixel scale, we suggest that monitoring methodologies continue to exploit the Landsat archive, but should aim to use multi-seasonal derived information. When the coarser 120 m pixel scale is adequate, integration of Landsat and SAR data should be considered, especially in areas with lower woody cover densities. The use of multiple seasonal compositing periods offers promise for large-area mapping of savannahs, even in regions with a limited historical Landsat coverage.

Keywords:
- Landsat-metrics;
- Optical-radar fusion;
- woody cover mapping;
- Savannas;
- Large-area mapping

1 Introduction

Savannah ecosystems are characterised by a dynamic mosaic of tree, shrub and grass species. Variations in these components can result in widely divergent ecological functions (Sankaran et al., 2005). There is growing concern over the health and sustainability of savannahs across the world. Increases in shrub cover at the expense of grasslands (i.e. shrub encroachment) have been reported in semi-arid environments globally (Naito and Cairns 2011, Stevens et al., 2016, Tian et al., 2016, Skowno et al., 2017). In contrast, overexploitation of woody shrubs and trees for fuelwood may be depleting woody cover in other regions (Wessels et al., 2013, Brandt et al., 2017).
Monitoring savannahs is a challenging endeavour, and due to the discontinuous nature of land cover in such environments, categorical maps are of limited value. Alternatively, representing the 2-dimension horizontal woody cover component as a continuous fractional layer is more ecologically relevant, and recent advances in the field have focused their attention to this characteristic (Bucini et al., 2010, Armston 2009, Naidoo et al., 2016). However, the spatial heterogeneity of savannahs makes fractional cover modelling vulnerable to scale effects, as areas of very high or low coverage will be lost by aggregation to coarser scales (Guerschman et al., 2009). Therefore, it is necessary to consider analyses over a range of resolutions, enabling an optimum balance between model accuracy and spatial detail to be established (Urbazaev et al., 2015).

Passive optical Earth observation (EO) data, such as Landsat, have commonly been employed to map savannah vegetation in the past (Prince and Astle 1986). Such data discriminate vegetation type by exploiting the full spectral range of reflected solar radiation. Passive optical data also allow for vegetation indices, such as the Normalized Difference Vegetation Index (NDVI), to be used as proxies of various biogeophysical parameters, such as net primary productivity (NPP), fraction of photosynthetically active radiation (fPAR), and leaf area index (LAI) (Carlson and Ripley 1997, Higginbottom and Symeonakis 2014, Zhu et al., 2013). Yet single date optical imagery can be inappropriate for discriminating woody and grass coverage, as photosynthetic activity is detected indiscriminately (Olsson, Leeuwen, and Marsh 2011). In savannahs, the woody cover component decreases temporal variation within the NDVI signal, as bushes and shrubs maintain leaves throughout the dry season (Bucini et al., 2010, Naidoo et al., 2016). Information derived from a pixel-level time series can therefore contain valuable information for land cover mapping. If sufficient observations are available, phenological metrics detailing the start and end points of seasons can be calculated (Brandt et al., 2016). Alternatively, spectral-temporal variability metrics from single spectral bands or indices (e.g. minimum, maximum, mean, median, etc.) can quantify variability even in regions with lower observation densities (Müller et al., 2015, Zhong, Gong, and Biging 2014).
Irrespective of processing method, optical data possesses fundamental limitations for mapping woody environments, because it does not directly correlate to surface structure (Naidoo et al., 2016). Active EO sensors such as Synthetic Aperture Radar (SAR) provide information on the 3-dimensional structure of the land surface, by emitting microwaves and measuring the intensity of energy reflected back to the sensor after interactions with ground objects i.e. the backscatter ($\sigma^0$) of the signal. The use of SAR data in fractional woody cover mapping, particularly L-band, operating with wavelengths of 0-15 cm, has been well demonstrated (Bucini et al., 2010, Mathieu et al., 2013, Naidoo et al., 2015, 2016). Mitchard et al., (2009) identified a consistent relationship between cross-polarised L-band backscatter and aboveground biomass (AGB) across four pan-African tropical savannahs, regardless of vegetation composition. Advanced Land Observing Satellite (ALOS) Phased Array type L-band Synthetic Aperture Radar (PALSAR) imagery has been highlighted as the most reliable satellite-based indicator of both AGB and canopy coverage for woody cover in semi-arid savannahs (Naidoo et al., 2015, 2016). However, the use of L-band imagery for mapping long-term land cover change is affected by a number of data continuity issues, sensor failures (JERS-1, ALOS PALSAR), high data costs, and the short lifespan of radar systems, resulting in a limited temporal archive compared to Landsat. There are less limitations when using C-band radar, such as Radarsat or Sentinel-1, due to more consistent coverage (Reiche et al., 2016). However, C-band radar is not as sensitive to woody cover, compared to L-band (Mathieu et al., 2013).

More recently, the fusion of optical and radar imagery has been shown to provide an improvement upon single-sensor fractional cover accuracies (Bucini et al., 2010, Naidoo et al., 2016). Bucini et al., (2010) and Naidoo et al., (2016) combined L-band radar data with Landsat to map woody canopy coverage in the Kruger National Park, South Africa: the fusion approach improved the accuracy over single sensor predictions, particularly when combining SAR with multi-season imagery. Lucas et al., (2006) used PALSAR thresholds in conjunction with Landsat-derived Foliage Projected Cover maps to successfully discriminate regrowth stages in open Eucalyptus forests. Merging various SAR wavebands, such as C, X, or L, have also been shown to provide benefits for woody cover mapping,
although these improvements were found to be smaller (~3%) when compared with L-band alone (Naidoo et al., 2015). Choosing the appropriate sensor, or combination of sensors, for woody cover mapping, is therefore an increasingly complex decision with further study required.

The increasing number and variety of EO systems in operation, coupled with open-data policies, presents a wide range of pathways for land cover mapping. Compared to earlier investigations, it is now routine for studies to use high-dimensional data. However, this approach comes with statistical limitations. Predictive models trained using high-dimensional data are prone to overfitting, thus transferring poorly to unseen validation data. This issue is important, potentially incurring a high degree of variance into classifications, whilst reducing bias (i.e. the bias-variance dilemma) (James et al., 2013, Kuhn and Johnson 2013). A number of techniques exist to process high-dimensional data and extract the most relevant variables, aiming to reduce model complexity whilst retaining predictive accuracy (Guyon et al., 2002, Guyon and Elisseeff 2003). To date the implementation of these methods in remote sensing analyses has been limited (Meyer et al., 2016), but may be increasingly beneficial in the near future as the number of data sources continues to increase.

Within this context, the overarching aim of this study is to develop a framework for accurately mapping the fractional woody cover of semi-arid savannahs at large spatial scales, using freely and widely available data sources. We address this overarching aim by carrying out a multi-scale comparative exercise that provides answers to the following questions:

1. Can annual time series of Landsat metrics be used to accurately map fractional woody cover, and to what extent does seasonality of the compositing period influence results?

2. How do Landsat-based estimates compare to multi-sensor fusion approaches combining L-band SAR data?

3. Can automated variable selection methods, such as Recursive Feature Elimination, assist in reducing the number of variables used without compromising accuracy?


2 Study Area

Our study area is the Limpopo Province, South Africa (Fig 1). The province is ~125,000 km² and intersects 14 Landsat WRS-2 scenes. This region is predominantly open deciduous savannah and grassland, with discontinuous woody cover ranging from 0-60% coverage (Mucina and Rutherford 2006). The climate is mainly semi-arid with small humid subtropical areas (Kottek et al., 2006). Mean annual temperatures range from 21-23°C and winters are mild and frost-free (Scholes et al., 2001). Rainfall increases along a north-south gradient, with mean annual precipitation of 450 mm/year in the north, rising to 700 mm/year in the south (Scholes et al., 2001). The majority of rainfall occurs in the winter months (October to March; Fig 2).

Figure 1: Location of the study area, the Limpopo Province of South Africa
Figure 2: Annual seasonality of NDVI for a dense shrub and grassland pixel, and regional rainfall for the Limpopo Province. NDVI is the mean value from 15 year of MODIS-MCD43A4 16 day observations, rainfall data is the mean and standard deviations from FEWS-NET (https://www.fews.net/). The vertical lines indicate the start and the end of the dry season.

There are pronounced contrasts in land use intensity across the region. In the east, the Kruger National Park is the largest protected area in South Africa featuring minimal human usage beyond fire experiments and animal enclosures. This contrasts with the communally governed areas originating from apartheid-era homelands (Worden 2012). These areas generally feature very high population densities ranging from 200-300 people per km$^2$, resulting from forced resettlement in the 1960-1990 apartheid period (Pollard et al., 2003). Consequently, overgrazing and unsustainable wood harvesting are widespread with many areas classified as degraded (Wessels et al., 2013).
3 Data

3.1 Training and validation data

We aimed to develop a transferable method for woody cover mapping. Accordingly, we used training and validation data from aerial imagery, so that our methodological framework would be applicable in study areas where such imagery is available but other data may not be or are costly, e.g. field surveys, Lidar. In South Africa, the National Geospatial Information (NGI) agency of the Department of Rural Development and Land Reform have been providing 0.5 m colour aerial photography since 2008, with an orthorectification accuracy of ±3 m (NGI 2017). Six 5×5 km images were selected according to a stratified approach based on mean annual precipitation, with acquisition dates between the 19th April and 7th August of the years 2008 and 2009 (Appendix 1).

3.2 Satellite imagery

3.2.1 Landsat

Spectral-temporal variability metrics are a method of capturing information on the temporal evolution of spectral values within a pixel (Muller et al., 2016). We hypothesised that metrics capturing this variability would be effective for woody cover monitoring. To generate metrics, all available Landsat 5 and 7 images that intersected the Limpopo Province for 2009-2010 were used, for the wet season additional images from the two neighbouring hydrological years were also used (scene footprints shown in Fig 1). Top-of-atmosphere (TOA) reflectance was calculated using standard bias-gain equations. Pixels affected by clouds or cloud shadow were removed based on the F-mask algorithm (Zhu and Woodcock 2012), no correction was applied for missing Scan Line Corrector (SLC-off) pixels. For each pixel, all co-located observations were used to calculate the following statistics: mean, median, minimum, maximum, and standard deviation. These metrics were calculated over three time-periods: annual, dry season and wet season (Fig 2), resulting in a total of 90 Landsat-
derived layers. The number of images used within each observation period is given in Table 1. Due to persistent high cloud cover, wet season metrics were calculated over three southern hemisphere hydrological years. Processing was undertaken in the Google Earth Engine cloud computing environment (Gorelick et al., 2017, Moore and Hansen, 2011).

<table>
<thead>
<tr>
<th>Period</th>
<th>Start Date</th>
<th>End Date</th>
<th>Landsat 5 Images</th>
<th>Landsat 7 Images</th>
<th>Total Images</th>
</tr>
</thead>
<tbody>
<tr>
<td>Annual Cycle</td>
<td>1st January</td>
<td>31st December</td>
<td>86</td>
<td>259</td>
<td>345</td>
</tr>
<tr>
<td>Dry Season</td>
<td>1st November</td>
<td>30th April</td>
<td>52</td>
<td>186</td>
<td>238</td>
</tr>
<tr>
<td>Wet Season</td>
<td>1st May</td>
<td>1st October</td>
<td>27</td>
<td>102</td>
<td>129</td>
</tr>
<tr>
<td>Total Unique Images</td>
<td></td>
<td></td>
<td>88</td>
<td>324</td>
<td>412</td>
</tr>
</tbody>
</table>

Table 1: Number of Landsat images used in each period for variability metric calculations. Wet season metrics are calculated over three hydrological years: 2009-2010, 2010-2011, and 2011-2012. Total Unique does not equal the sum of rows as images can be included in both a single season and the annual period.

3.2.2 ALOS-PALSAR

ALOS PALSAR, and its successor ALOS-2 PALSAR-2, are fully polarimetric L-band Synthetic Aperture Radar systems (Rosenqvist et al., 2007). These sensors operate at a wavelength of 23.6 cm. We used the 2010 data from the ALOS PALSAR global mosaic, a science-ready product generated annually for 2007 to 2010 (ALOS), and 2015 (ALOS-2). The images for this mosaic were from the dry season, with acquisition dates between 1st July - 3rd October and two images from 2009. Dual polarization HH (horizontal-horizontal) and HV (horizontal-vertical) backscatter data were used. Pre-processing of the input raw imagery includes orthorectification using the Shuttle Radar Topography Mission (SRTM).
Digital Elevation Model (DEM), calibration, speckle reduction, and a destriping procedure (Shimada and Ohtaki 2010, Shimada et al., 2014). The raw digital number format was converted to backscatter ($\sigma^0$) using the calibration equation:

$$\sigma^0 = 10 \times \log_{10}(DN + 0.001)^2 + CF$$  \hspace{1cm} (1)$$

where DN is the raw digital number and CF is a calibration factor (= -83). The 25 m mosaic was resampled to match the Landsat resolution using bilinear resampling.

4 Methods

The methodological framework is shown in Fig 3. To establish the optimum approach for fractional woody cover mapping, we ran a series of random forest regressions to compare the accuracies achieved from single season Landsat metrics, multi-season data, or multi-sensor combining Landsat and SAR data. These models were repeated at four resolutions: 30, 60, 90 and 120 m, to ascertain the ideal scale for large-area monitoring. Processing was undertaken in the R Statistical Software Environment, using the raster, caret, and randomForest packages (Hijmans et al., 2015, Kuhn 2015, Liaw and Wiener 2002, R Core Team 2015). Fractional cover sampling code was adapted from Leutner and Horning (2016).
Figure 3. Flow chart of methodological framework. VVI: Visible Vegetation Index; PCA: Principle Components Analysis; RFE: Recursive Feature Elimination.

4.1 Creation of Reference Data

To create training data the six aerial imagery subsets were classified into woody/non woody masks. We opted for aerial image classification to enable methods to be transferable to other locations, due to the generally satisfactory availability of aerial imagery at appropriate scales (Staben et al., 2016). Firstly, a principal components analysis (PCA) was applied to the three RGB layers and the first two components were extracted. Secondly, we calculated the visible vegetation index (Joseph and Devadas 2015) which uses visible light spectra to estimate photosynthetic activity and is defined as:

\[
VVI = \left[ (1 - \frac{R - R_0}{R + R_0})(1 - \frac{G - G_0}{G + G_0})(1 - \frac{B - B_0}{B + B_0}) \right]
\]
where VVI is the visible vegetation index, $R$, $G$ and $B$ are the red, green, and blue intensities in the image, $R_0$, $G_0$ and $B_0$ are values of red, green, blue used to reference green colour (30, 50, and 1, respectively), determined by the image bit rate (Joseph and Devadas 2015).

A Random Forest classifier was used to create the binary woody-non woody layers from the original RGB layers, principle components, and VVI. Individual models were generated for each image using 400 manually selected points per image (75/25% training-validation split). The mean classification accuracy was 85%. Full accuracy statistics are given in the Appendix. An example classified mask is shown in Fig 4.

![Example of the RGB woody classification. (a) raw RGB image; (b) classified woody cover shown in red, and (c) 30 m grid for fractional cover sampling.](image)

To generate training and validation data for the satellite imagery, Landsat pixel-sized squares (i.e. 30×30 m) were extracted from the woody/non-woody masks and the percentage woody coverage calculated. From each image, $7000/\alpha$ samples were extracted, where $\alpha$ takes the values of 1, 2, 3 or 4, depending on the aggregation level used to test the effect of scale in the accuracy of the woody cover estimates (Fig 3). For example, for a pixel size of 30m, $\alpha=1$ and the samples extracted from each image are 7000, whereas for an aggregation level of $\alpha=2$ or a pixel size of 60 m, the number of samples extracted are $7000/2=3500$. These samples were merged and split into equal training and validation subsets with equal probability distributions of woody cover (Fig 5). The spatial aggregation process may incur central tendency in training values, with the reduction in high or low samples making the
subsequent regression task easier. To quantify this, we tested for any significant difference between
the sample distributions using Pairwise-Wilcoxon tests. These highlighted a significant ($p < 0.05$)
difference between the data at 30 m and all other scales which can also be visualised in the relatively
reduced number of high (>75%) and low (<10%) values in the respective aggregated pixel histograms
(Figure 5).

Figure 5: Density histograms of model training values at the four scales tested

4.2 Random Forest Regression

Predictive models were generated using the Random Forest algorithm. Random Forest is an
ensemble machine learning procedure that combines bootstrapping and aggregation (bagging) with
decision trees (Breiman 2001). All models were individually tuned using 10 repeats of 10-fold cross
validation to identify the ideal parameter specification. This covered the number of variables considered at each tree node and the number of trees constructed.

4.3 Variable Selection

To identify optimum predictive models, we incorporated all potential variables into a variable selection process. According to statistical learning theory, a model containing fewer predictors that is comparably accurate is preferential to a more complex model (James et al., 2013, Kuhn and Johnson 2013). Backwards selection methods are effective in identifying the ideal number of variables for prediction, allowing the selection of the most parsimonious model that offers comparable accuracy (Guyon et al., 2002, James et al., 2013, Kuhn and Johnson 2013). The combination of Landsat metrics and PALSAR data resulted in 92 predictors (90 Landsat metrics + 2 PALSAR backscatter), a number of which are correlated. To identify the most important predictors, we implemented the backwards selection method of recursive feature elimination (RFE). RFE is a parameter selection process that incorporates the estimation of test (validation) errors and variable importance (Guyon et al., 2002). Firstly, a model is constructed using all available predictors ($M_p$). The test error of this model (i.e. adjusted $R^2$ and RMSE) is then estimated using 10-fold cross validation, and variable importance scores are calculated. A second model is then established which excludes the lowest contributing variable from $M_p$, and test error and variable importance are recalculated. This process is repeated until a one-variable model remains. A full iteration of this procedure is repeated 10 times to account for variations in the cross validation sampling, providing a robust estimate of test errors. An ideal model that offers the best performance whilst using the least variables can then be selected.
5 Results

5.1 Woody Cover Mapping

A fractional woody cover map derived from the most accurate model tested, is shown in Fig 6. Subsets comparing the mapped woody cover estimates from a number of models and the NGI aerial imagery are shown in Fig 7.

Figure 6. Fractional woody cover results for the Limpopo Province based on the Recursive Feature Elimination model at the 120 m pixel scale. Black squares A and B are the locations of the subsets in Fig 7.
Figure 7. Spatial patterns of woody cover for subsets A and B of Fig 6 at 30 and 120 m pixel scales.

Five model predictions and the respective reference aerial imagery from the NGI are shown.

Aerial imagery acquisition dates: A: 19 April 2009, B: 30 April 2009.
5.2 Seasonal Landsat models

The performance of Landsat-based models is shown in Fig 8 and Table 2. When using metrics derived from a single season, the highest accuracies were obtained by using the dry season metrics, followed by the full annual cycle, with the wet season performing the worst. This pattern was consistent across all scales (Table 2). Using a combination of metrics derived from two seasons, the highest accuracies came from models incorporating both dry and wet season data, followed by dry and annual, and finally wet and annual (Table 2). Reducing the pixel resolutions (i.e. increasing the aggregation factor), consistently raised the model performances, with the largest improvement occurring in the initial aggregation from 30 m to 60 m.

Figure 8. Model accuracy results for Landsat metrics-based models RMSE units are percentage woody cover (0-100%)

17
Table 2. Model accuracy results for the Landsat metrics-based models, RMSE units are percentage woody cover (0-100%)

<table>
<thead>
<tr>
<th>Model</th>
<th>120 m n-3,848</th>
<th>90 m n-6,826</th>
<th>60 m n-10,499</th>
<th>30 m n-21,000</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>R²</td>
<td>RMSE</td>
<td>R²</td>
<td>RMSE</td>
</tr>
<tr>
<td>Landsat Dry and Wet</td>
<td>0.77</td>
<td>9.4</td>
<td>0.762</td>
<td>10</td>
</tr>
<tr>
<td>Landsat Dry and Annual</td>
<td>0.763</td>
<td>9.5</td>
<td>0.76</td>
<td>10</td>
</tr>
<tr>
<td>Landsat Wet and Annual</td>
<td>0.76</td>
<td>9.6</td>
<td>0.756</td>
<td>10.1</td>
</tr>
<tr>
<td>Landsat Dry</td>
<td>0.741</td>
<td>9.9</td>
<td>0.732</td>
<td>10.5</td>
</tr>
<tr>
<td>Landsat Annual</td>
<td>0.717</td>
<td>10.4</td>
<td>0.723</td>
<td>10.7</td>
</tr>
<tr>
<td>Landsat Wet</td>
<td>0.714</td>
<td>10.5</td>
<td>0.702</td>
<td>11.1</td>
</tr>
</tbody>
</table>

5.3 Fused models

Accuracy statistics from models combining Landsat metrics with ALOS PALSAR backscatter are shown in Fig 9 and Table 3. Overall, the same ranking of seasonal performance as Landsat-only models was observed. For a single season, accuracy decreased from dry to annual to wet, whilst multi-season models were ranked: dry and wet, dry and annual, and wet and annual. The only exception to this order was at the 90 m pixel scale, where the single season annual metrics and dry-annual multi-season models performed best (Table 4).

The fusion of PALSAR backscatter with Landsat metrics had contrasting impacts on model accuracy (Table 4). At the 120 m scale, all models were improved. Conversely, at the 30 m scale, performances were negatively affected. At mid-range scales (60 and 90 m), the single season annual models were improved, as did the 90 m ‘wet’ model. All other mid-scale models responded negatively to the SAR fusion or were unaffected. At the 120 m scale, the fusion was generally more effective for single season models over multi-temporal combinations. Finally, at all scales, the annual models performed better when used together with the SAR data.
Figure 9: Model accuracies ($R^2$ and RMSE) for Landsat-PALSAR fusion models

Table 3. Accuracy metrics for Landsat-PALSAR fusion models

<table>
<thead>
<tr>
<th>Model</th>
<th>$R^2$ (120 m)</th>
<th>RMSE (120 m)</th>
<th>$R^2$ (90 m)</th>
<th>RMSE (90 m)</th>
<th>$R^2$ (60 m)</th>
<th>RMSE (60 m)</th>
<th>$R^2$ (30 m)</th>
<th>RMSE (30 m)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Landsat Dry and Wet</td>
<td>0.773</td>
<td>9.3</td>
<td>0.757</td>
<td>10</td>
<td>0.729</td>
<td>11.2</td>
<td>0.648</td>
<td>14.3</td>
</tr>
<tr>
<td>Landsat Dry and Annual</td>
<td>0.769</td>
<td>9.4</td>
<td>0.758</td>
<td>10</td>
<td>0.729</td>
<td>11.2</td>
<td>0.642</td>
<td>14.4</td>
</tr>
<tr>
<td>Landsat Wet and Annual</td>
<td>0.763</td>
<td>9.5</td>
<td>0.753</td>
<td>10.1</td>
<td>0.721</td>
<td>11.3</td>
<td>0.641</td>
<td>14.5</td>
</tr>
<tr>
<td>Landsat Dry</td>
<td>0.755</td>
<td>9.6</td>
<td>0.731</td>
<td>10.5</td>
<td>0.703</td>
<td>11.7</td>
<td>0.614</td>
<td>15</td>
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<tr>
<td>Landsat Annual</td>
<td>0.742</td>
<td>9.9</td>
<td>0.737</td>
<td>10.4</td>
<td>0.698</td>
<td>11.8</td>
<td>0.611</td>
<td>15</td>
</tr>
<tr>
<td>Landsat Wet</td>
<td>0.723</td>
<td>10.3</td>
<td>0.706</td>
<td>11</td>
<td>0.668</td>
<td>12.3</td>
<td>0.601</td>
<td>15.2</td>
</tr>
<tr>
<td>PALSAR Only</td>
<td>0.37</td>
<td>15.5</td>
<td>0.313</td>
<td>16.9</td>
<td>0.25</td>
<td>18.7</td>
<td>0.180</td>
<td>22.2</td>
</tr>
</tbody>
</table>
Table 4. Difference between the Landsat only and Landsat-PALSAR fusion models. Green numbers indicate improvement from the fusion while red the opposite.

<table>
<thead>
<tr>
<th></th>
<th>120 m R²</th>
<th>120 m RMSE</th>
<th>90 m R²</th>
<th>90 m RMSE</th>
<th>60 m R²</th>
<th>60 m RMSE</th>
<th>30 m R²</th>
<th>30 m RMSE</th>
</tr>
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<tbody>
<tr>
<td>Landsat Dry and Wet</td>
<td>0.003</td>
<td>-0.1</td>
<td>-0.5</td>
<td>0</td>
<td>-0.9</td>
<td>0.2</td>
<td>-0.005</td>
<td>0.1</td>
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<tr>
<td>Landsat Dry and Annual</td>
<td>0.006</td>
<td>-0.1</td>
<td>-0.2</td>
<td>0</td>
<td>-0.4</td>
<td>0.1</td>
<td>-0.004</td>
<td>0</td>
</tr>
<tr>
<td>Landsat Wet and Annual</td>
<td>0.003</td>
<td>-0.1</td>
<td>-0.3</td>
<td>0</td>
<td>-0.7</td>
<td>0.1</td>
<td>-0.002</td>
<td>0.1</td>
</tr>
<tr>
<td>Landsat Dry</td>
<td>0.014</td>
<td>-0.3</td>
<td>-0.1</td>
<td>0</td>
<td>-0.2</td>
<td>0.1</td>
<td>-0.004</td>
<td>0.1</td>
</tr>
<tr>
<td>Landsat Annual</td>
<td>0.025</td>
<td>-0.5</td>
<td>1.4</td>
<td>-0.3</td>
<td>0.9</td>
<td>-0.2</td>
<td>0.001</td>
<td>-0.1</td>
</tr>
<tr>
<td>Landsat Wet</td>
<td>0.009</td>
<td>-0.2</td>
<td>0.4</td>
<td>-0.1</td>
<td>-0.7</td>
<td>0</td>
<td>-0.003</td>
<td>0</td>
</tr>
</tbody>
</table>

5.4 Recursive Feature Elimination (RFE)

The accuracy results from the RFE automated variable selection approach is shown in Fig 10. At all scales, model accuracies were higher when more than 25 variables were included in the model and performance declined rapidly when fewer than that were considered. The optimal number of variables to balance predictive accuracy and model simplicity was established as 57 for the 120 m-pixel scale, 54 for the 90 m, 70 for the 60 m, and 85 for the 30 m, the top five variable for each model are shown in Table 6. Applying a threshold of two standard errors, based on the cross validations samples for the best model, allows similarly performing models to be compared (James et al., 2013). These models ranged from the one that includes all 92 layers to a minimum of 14 variables for the 120 m scale, 20 for the 90 m scale, 29 for the 60 m, and 39 for the 30 m scale. At all scales, the model constructed by the RFE was the best preforming (Fig 11), providing an improvement in the achieved $R^2$ of at least 0.012 (Table 5). The 120m scale RFE model was the overall most accurate (Fig 6). To compare the within model variation in accuracy, Figure 12 shows class accuracy statistics for 10% intervals of woody cover.
Figure 10: Cross-validated $R^2$ and RMSE results from the recursive feature elimination (RFE) process.

Table 5. Accuracy metrics for the model produced by the recursive feature elimination (RFE), all 92 variables, and the best Landsat-only and Landsat-SAR fused combinations.

<table>
<thead>
<tr>
<th></th>
<th>120 m</th>
<th></th>
<th>90 m</th>
<th></th>
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</thead>
<tbody>
<tr>
<td></td>
<td>$R^2$</td>
<td>RMSE</td>
<td>$R^2$</td>
<td>RMSE</td>
<td>$R^2$</td>
<td>RMSE</td>
<td>$R^2$</td>
<td>RMSE</td>
</tr>
<tr>
<td>Recursive Feature Elimination</td>
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<td>8.9</td>
<td>0.777</td>
<td>9.7</td>
<td>0.75</td>
<td>11.0</td>
<td>0.661</td>
<td>14.2</td>
</tr>
<tr>
<td>All 92 Variables</td>
<td>0.778</td>
<td>9.2</td>
<td>0.767</td>
<td>9.8</td>
<td>0.741</td>
<td>11.0</td>
<td>0.655</td>
<td>14.2</td>
</tr>
<tr>
<td>Landsat Dry and Wet</td>
<td>0.77</td>
<td>9.4</td>
<td>0.762</td>
<td>10.0</td>
<td>0.738</td>
<td>11.0</td>
<td>0.653</td>
<td>14.2</td>
</tr>
<tr>
<td>Landsat Dry and Wet + SAR</td>
<td>0.773</td>
<td>9.3</td>
<td>0.757</td>
<td>10.0</td>
<td>0.729</td>
<td>11.2</td>
<td>0.648</td>
<td>14.3</td>
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</tbody>
</table>
Figure 11 Density scatter plot of the Recursive Feature Elimination models at the four resolutions
Figure 12. Balanced accuracy figures for the different 120 and 30 m-scale models and woody cover density classes, the original continuous woody cover values were binned into 10% intervals. RFE - Recursive Feature Elimination model.
### Table 6: Top five variables from the Recursive Feature Elimination model, at each scale (30m, 60m, 90m and 120m).

<table>
<thead>
<tr>
<th></th>
<th>30 m</th>
<th>60 m</th>
<th>90 m</th>
<th>120 m</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>HH</td>
<td>HH</td>
<td>HV</td>
<td>HV</td>
</tr>
<tr>
<td>2</td>
<td>B1 SD Dry</td>
<td>B1 SD Dry</td>
<td>HH</td>
<td>HH</td>
</tr>
<tr>
<td>3</td>
<td>B2 SD Dry</td>
<td>B4 SD Dry</td>
<td>B4 Median</td>
<td>B3 Median</td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td>Annual</td>
<td>Annual</td>
</tr>
<tr>
<td>4</td>
<td>B4 Median Wet</td>
<td>HV</td>
<td>B3 Median Dry</td>
<td>B5 Median Dry</td>
</tr>
<tr>
<td>5</td>
<td>B4 SD Dry</td>
<td>B1 Min Wet</td>
<td>B3 Median</td>
<td>B3 Median</td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td>Annual</td>
<td>Dry</td>
</tr>
</tbody>
</table>

SD: standard deviation.

## 6 Discussion

### 6.1 Landsat metrics, seasonality and scale

The accuracies obtained from the Landsat-derived woody cover maps varied according to the temporal window for which metrics were calculated. For single season data, the dry period metrics were the most useful. This result was anticipated due to the persistence of green shrubs into the dry season, compared with the grass layer (Fig 2; Naidoo et al., 2016). This makes woody cover easier to discriminate, compared to other periods where differences are less pronounced (Brandt et al., 2016). This can also explain the overestimation of wet season models in Fig 7, as the grass and wood layers are more difficult to separate and identify.

The distribution of errors also varied with seasonality. Dry season metrics performed better in areas of sparse woody cover (0-30% cover), whereas wet metrics offered marginal improvements in the 30-40% and 50-60% percentiles (Fig 12). This can be attributed to the dry season metrics having relatively a greater discriminatory power at sparse coverage where woody canopies are more distinct.
Furthermore, some areas of moderate woody coverage were under-predicted by dry season only metrics. This can be attributed to the fact that some woody species are less persistent in dry conditions (Subset B in Fig 7).

The best result from the multi-seasonal Landsat comparisons was the combination of dry and wet season metrics. Although wet season metrics were the least effective mono-temporal models, when combined with the contrasting dry season, the information covering the peak biomass period was beneficial. This improvement was mainly limited to coverage between 10 and 70% where each percentile produced greater class accuracies than either single-season case, at both fine (30 m) and coarse (120 m scales). In general, the multi-seasonal combinations improved prediction across the full range of woody cover densities, with the 10-40% percentiles, at 120 m resolution, achieving the highest-class accuracies. The ability to extract multiple sets of metrics from a time-series of images is noteworthy, reducing to a certain extent the drawback of a temporally limited Landsat archive in many savannah regions.

As fractional woody cover approaches the highest values (>70%), all models perform poorly with no model achieving a percentile class accuracy of more than 56% (Fig 12). This is partly due to the rare occurrence of this class, which affects the regression analysis. The poor accuracy for dense woody savannahs has been noted by numerous other studies (e.g. Bucini et al., 2010, Naidoo et al., 2016), and should be a priority for future studies.

We tested models at four scales: 30, 60, 90, and 120 m pixels. As pixel size increased, model accuracies consistently improved (Figs 8 and 9). The largest improvement occurred with the initial aggregation from 30 to 60 m, with a mean $R^2$ increase of 13.09%±0.9, across the 13 models tested. However, this change must be considered with the distribution of the input training values. At 30 m, there is a relatively larger spread of values and a higher proportion of dense and sparse woody coverage (Fig 5). Accordingly, this distribution is a more complicated endeavour for the regression analysis, as indicated by the low class accuracies for high cover percentiles (Fig 12). Concurrently, the
greater proportion, and pixel purity, of sparsely (0-10%) wooded areas at 30 m result in comparably high class accuracies for the first percentile class (Fig 12). Resampling to a coarser resolution reduces the occurrence of dense woody coverage, due to central tendency, making the regression exercise easier. This simplification is restricted to the 30 to 60 m aggregation with no visual or statistical evidence that additional resampling improves the outcome of the regression. Further reductions in the pixel resolution result in more modest but consistent improvements of 4.20%±0.74 and 1.30%±0.99 in the $R^2$ when re-scaling from 60 to 90 m and from 90 to 120 m, respectively. At coarser scales, artefacts from the Landsat processing are likely to be smoothed, as errors resulting from the Scan Line Corrector failure and undetected clouds are minimised (Fig 7). Furthermore, despite the high georeferencing accuracy of the datasets, errors from potential misalignment of the training imagery and Landsat data may be more prevalent at 30 m and averaged at coarser scales. For many regional-scale applications, land cover maps at 90 or 120 m may be sufficient, and an accuracy vs. precision trade-off might be appropriate. Maps at 120 m may be more accurate, but have less fidelity for detecting the clumps and canopies of dryland vegetation. This trade off may become more pertinent with the availability of 10-20 m imagery from Sentinel-2 (Bastin et al., 2017).

Overall, the accuracies achieved by the Landsat-based models are comparable to those of radar-based studies at similar scales. Urbazaev et al., (2015) achieved $R^2$ values of 0.71 and 0.66 using multiple and single season PALSAR images at 50 m resolution, respectively, whilst Naidoo et al., (2016) obtained $R^2$ of 0.8 and 0.81 using single-season PALSAR data at 105 m. Given our considerably larger study area, our results are promising for regional-scale analysis, as the spatial breadth, temporal depth, and rapid processing potential of the Landsat archive is unmatched by any radar system (Kennedy et al., 2014, Roy et al., 2014). Our metrics-based approach outperforms the various single date Landsat scenarios across multiple seasons achieved by Naidoo et al. (2016) who reported $R^2$ values of 0.32-0.65 at 105 m resolution. There are clear benefits to quantifying seasonal variability using metrics, as demonstrated by the high ranking of standard deviation layers (Table 6). Furthermore, multi-seasonal metrics further improved results over multi-seasonal image
combinations. We attributed this refinement to two factors: firstly, metrics are more resistant to bias incurred by rainfall and moisture variation; secondly, metrics such as standard deviation can represent the temporal profile of the land cover, imitating time-series approaches. This is in agreement with Müller et al. (2015), who found that annual metrics outperform best available pixel composites for tropical savannahs in Brazil.

Large-area mapping of savannah systems remains a challenge due to high heterogeneity and subjective biome classifications (Herold et al., 2008, Hüttich et al., 2011). Current approaches for regional-scale mapping generally focus on best-available pixel composites for classification (Griffths et al., 2013, White et al., 2014, Frantz et al., 2017). Due to the high temporal variation in savannahs, this method is particularly vulnerable to bias effects caused by pixels being selected in different phenological stages (Hüttich et al., 2011, Müller et al., 2015). We demonstrate that Landsat-based spectral variability metrics offer a robust alternative for land cover mapping at large spatial scales, applicable to epochal or annual analyses. South Africa possesses good availability of Landsat imagery in the USGS archive, owing to the successful transfer of data from the Johannesburg receiving station, active since 1980 (Wulder et al., 2015). However, in many savannah regions, such as the Sahel and east Africa, the historical Landsat archive is sparse. By combining multiple years of observations, wall-to-wall mapping should be possible even with low annual image availability. Furthermore, segmenting a time-series into multiple temporal windows allows additional value to be extracted from a single series of observations, potentially compensating for a relatively limited archive. The high image acquisition rate of Landsat 8 relative to the historic Landsat archive, combined with comparable imagery from the Sentinel-2 satellites, will result in improved temporal resolution for optical imagery (Drusch et al., 2012, Roy et al., 2014). Increased observations should enable our multi-seasonal metrics approach to be expanded by using more or smaller temporal windows, for example the beginning or ending of the dry season. Evidence from MODIS-based studies suggests that this refinement may allow increased discrimination of subtle land covers, such as densely wooded savannahs, which are currently poorly mapped (Hüttich et al., 2009).
6.2 Landsat-PALSAR fusion

Integrating the Landsat metrics with L-band PALSAR backscatter had divergent impacts. Finer-scale maps (30 to 90 m) were negatively affected by the inclusion of radar, with the Landsat-only models outperforming their fused counterparts (Table 4). Comparably, the PALSAR-only models performed poorly, especially at fine-scales (Table 3). We attribute this to the high-level of noise in radar imagery at higher resolutions, as illustrated in Fig 13. Errors caused by factors such as speckle, moisture content and geolocation accuracy are far more prevalent in finer-scale radar data. Therefore, at 30 to 90 m pixel scales, the PALSAR imagery contains a weak signal-to-noise ratio, incurring a negative impact on the regression model. This is further reflected in the increasing ranks of radar variables in the Recursive Feature Elimination (RFE, Table 6). Accordingly, SAR-fusion reduced class accuracy by ~1% for area 20-60% coverage, at 30 m scales (Fig 12). Conversely, the coarse scale models (120 m) were consistently improved by the addition of PALSAR backscatter to the Landsat metrics, with the single-season combinations undergoing the greatest improvement. The lower improvements for the multi-seasonal scenarios indicates that some of the information contained in radar backscatter can be obtained from multi-seasonal metrics. The inclusion of L-band radar had the highest impact on sparse woody cover classes (0-30%). Within these classes, inclusion of the SAR variables increased balanced accuracies by 1-9% and 1-2%, at 120 and 30 m scales, respectively (Fig 12). Visual examination of the prediction subset maps indicates that this improvement is due to the SAR fusion correcting for overestimations when there is 0 - 20 % woody cover (e.g. the central pivot irrigation fields in Subset A of Fig 7).
Multi-sensor fusion approaches are becoming more popular, due to an increase in the number of operational sensors and the open-access data policies. The improvements at coarse scales are in line with those found in other studies employing SAR and Landsat data together (Bucini et al., 2010, Naidoo et al., 2016). However, this study is the first to quantify the effect and mechanism of this fusion across multiple seasons and scales. The accuracies of the PALSAR-only models generated here are lower than other South African studies (e.g. Naidoo et al., 2016, Urbazaev et al., 2015). We attribute this to the much larger and heterogeneous study area that we cover, encompassing human modified landscapes where the other two studies were confined within the Kruger National Park. The source of training data could also have affected the accuracy of our PALSAR-based estimates: we employ aerial photographs while Naidoo et al. (2016) and Urbazaev et al. (2015) use more accurate characterisations from the field or from LiDAR sources. It should also be noted that our study used a mosaicked ALOS PALSAR layer produced from images acquired across a three month window (1st July - 3rd October), including two images acquired in the previous year. Seasonal effects, such as canopy density and moisture content, may prevent the mosaicked images from being artefact-free. Alternatively, the global-scale processing undertaken in the creation of the mosaicked PALSAR layer,
such as speckle reduction and topographic normalisation, may reduce the fidelity of backscatter measurements when compared to scene-specific methods applied elsewhere (e.g. Naidoo et al., 2016, Urbazaev et al., 2015). Furthermore, multi-sensor fusion has a potential for image miss-registration errors between the imagery (Lehmann et al., 2015).

Although overall model accuracies are only moderately changed by the inclusion of L-band SAR data, the consistent allocation of improvements at low densities of woody cover may be highly relevant to semi-arid savannah case studies. The process of shrub encroachment into grasslands is a major threat to the livelihoods of many pastoralists in the developing world. For prevention and remediation to be successful, action must be taken as early as possible. The periodic monitoring of sparsely wooded savannahs, which are vulnerable to shrub encroachment, is therefore a pressing requirement. For this purpose, the fusion of PALSAR and Landsat imagery is beneficial, offering a higher likelihood of timely change detection than single-sensor approaches. In the coming years, fusion techniques based C-band radar from Sentinel-1 may offer good promise, owing to the 12 day revisit time.

6.3 Merit of variable reduction methods

To ascertain the value of variable reduction methods we applied a Recursive Feature Elimination (RFE) on out 92 variable dataset. The RFE produced the best preforming model at all scales, compared to all Landsat and Landsat-PALSAR fusion cases (Fig 11). In general, the number of variables used in the RFE models decreased with aggregation: we attribute the requirement of less variables at coarser resolutions to improvements in signal-to-noise ratios as noisier layers are smoothed. Dimension reduction methods are also useful for highlighting the type of variables that contain useful information for the model building. The high ranking of standard deviation - a proxy for seasonal variability highlights the importance of temporal information for woody cover mapping.
As both the number of active sensors and the availability of open data archives increase, remote sensing analyses are using high-dimensional datasets. The utility of variable selection or dimension reduction methods will inevitably increase in order to deal with the increasing data volume. Currently, these tools are primarily used in hyperspectral analyses, but are underutilised in other areas (Pal and Foody 2010). The fact that the RFE was able to automate the process of selecting a superior model highlights the potential of automating model construction using machine learning methods that may currently be underused (Meyer et al., 2016). At large-scales, mapping land cover with fewer variables can drastically reduce processing time, leaving unnecessary variables out can therefore be useful for computing and statistical purposes.

7 Conclusions

We tested the potential of Landsat-derived spectral variability metrics and PALSAR composites for mapping woody coverage, in southern African savannas. We compared the role of seasonal compositing period, and the effect of multi-sensor fusion through the addition of ALOS PALSAR backscatter to the Landsat layers. Furthermore, we investigated the role of pixel scale on map accuracy, and the potential of variable selection methods for automating the model building process.

We draw a number of conclusions from our modelling scenarios. Firstly, Landsat metrics can produce highly accurate maps of fractional cover in savannas, with dry season imagery being the preferred temporal window. Further improvements can be made by combining multi-seasonal metrics, derived from two contrasting seasons. In particular, integrating dry and wet season layers produced good improvements in map accuracy. Secondly, the fusion of Landsat and PALSAR layers is not always beneficial. At fine scales (30-60 m), L-band SAR integration reduced model performance consistently, potentially due to the high level of noise inherent to radar data,
particularly in savannahs. Conversely, at the 120 m scale, the addition of PALSAR was beneficial, particularly for areas with less than 30% coverage, and for some models at 90 m scales as well. Finally, the use of a recursive feature elimination automated variable selection process was very efficient in constructing an accurate parsimonious model, producing the most effective model at every scale examined whilst reducing the number of variables used to of 57 out of 90.

In summary, Landsat metrics offer a suitable option for regional-scale mapping of savannah woody cover, and should allow decadal scale analysis of land cover changes. The use of multi-seasonal composites are particularly promising for accurate fractional woody cover mapping. For contemporary monitoring, the fusion of Landsat metrics with L-band radar is recommended for areas with lower woody cover densities, and particularly for the rapid detection of shrub encroachment into grass-dominated savannahs. Future studies will benefit from automated variable reduction approaches and the increased image acquisition rates from the Sentinel constellation, that feature both radar (C-Band) and optical satellites.

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Appendix

Table A1 Woody cover mask classification accuracies.

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37
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**Positive Class: Woody Cover**