Laboratory Analysis of Elevated Atmospheric Carbon Dioxide on Carbon Fluxes for

Cyanobacterial Biological Soil Crusts

Richard W Lane^{1*}, Manoj Menon^{1*}, James B McQuaid^{§1}, David G Adams², Andrew

D Thomas^{3*}, Steve R Hoon³ and Andrew J Dougill¹

¹School of Earth and Environment, University of Leeds, Leeds, LS2 9JT, UK

²Institute of Integrative and Comparative Biology, Faculty of Biological Sciences,

University of Leeds, Leeds, LS2 9JT, UK

³School of Science and the Environment, Manchester Metropolitan University,

Manchester, M1 5GD, UK

§Corresponding author; Address: ¹School of Earth and Environment (Dr. Jim

McQuaid), University of Leeds, Leeds, LS2 9JT, UK; Email:

J.B.McQuaid@leeds.ac.uk

Phone: +44 (0) 113 343 6724; Fax: +44 (0) 113 343 5259

*Current address

Manoj Menon: Kroto Research Institute, Department of Civil and Structural

Engineering, University of Sheffield (North Campus), Sheffield, S37HQ, UK;

Andrew D Thomas: Institute of Geography and Earth Sciences, Aberystwyth

University, Penglais Campus, Aberystwyth, SY23 3DB, UK

Keywords: biological soil crust, cyanobacteria, photosynthesis, respiration, drylands, Kalahari, soil carbon sequestration, heterotrophic, autotrophic, CO2 flux double CO2,

Abstract

Metabolic activity of Biological Soil Crusts (BSCs) is principally dependent on moisture availability, but also on temperature and light conditions. Less understood is how BSCs respond to elevated atmospheric CO₂. This paper reports laboratory experimental results of elevated atmospheric CO₂ on carbon fluxes for cyanobacterial BSCs. The study uses newly designed dynamic gas exchange chambers in which the internal atmosphere was controlled. CO₂ flux was monitored during controlled experiments in two phases under simulated rainfall events (2 & 5 mm plus control with no wetting) each lasting 3 days with a dry period in between. Phase 1 subjected crusts to 392 ppm CO₂ (representing ambient level) in dry air; in phase 2, the CO₂ concentration was 801 ppm. Both phases exhibited significant efflux (respiration) of CO₂ immediately after wetting, followed by substantial influx (sequestration) of CO₂. Samples subject to 2 mm wetting sequestered an order of magnitude more C under elevated CO₂ than at ambient CO₂; for samples subject to 5 mm wetting, this increase was threefold. The findings highlight the role of BSCs in future carbon budgets by enabling greater sequestration into dryland soils even under enhanced atmospheric CO₂ concentrations, following both light and heavy rainfall events.

1. Introduction

Dryland soils cover approximately 41% of the Earth's terrestrial surface and support more than one-third of the global population (Reynolds et al., 2007). Recent findings suggest that sub-Saharan Africa has acted as a likely net carbon sink within the last decade (Ciais et al., 2009) with a net balance of approximately 0.97 Pg C y⁻¹ uptake (Bombelli et al., 2009). However, the role of soils and their Biological Soil Crust (BSC) cover remains only sparsely assessed (see reviews by Maestre et al., 2012 a,b; Stringer et al., 2012) but with BSCs identified as a major contributor to overall CO₂ efflux in drylands (e.g. Castillo-Monroy et al., 2011).

BSCs are made of cyanobacteria and other bacterial species, algae, lichens, mosses and microfungi. In general, BSCs cover up to 70 % (Belnap and Lange, 2003) of the soil surface of many drylands and in some undisturbed areas such as in Wildlife Management Areas of the Kalahari BSCs cover more than 90% of the soil surface (Dougill and Thomas, 2004). Globally, cyanobacterial biomass in dryland areas is estimated to be 56 Pg of carbon (Garcia-Pichel et al., 2003), therefore constituting a significant part of the estimated 241 Pg of soil organic carbon (SOC) contained within drylands (Lal, 2004). The estimated total global Carbon net uptake of BSCs has been approximated as 3.9 Pg yr⁻¹ (corresponding to approximately 7% of global net primary production uptake) (Elbert et al., 2012) but the controls on rates of CO₂ efflex remain poorly understood (Frey et al., 2013).

Substantial amounts of carbon are also being lost from BSCs via respiration (Huxman et al., 2004; Thomas et al., 2008; Thomas and Hoon, 2010; Thomas et al., 2011) particularly after rainfall events; due to activity and substrate availability increases within microbial populations (Borken and Matzner, 2009). During hours when moisture, temperature and light availability are optimum gain of carbon

(photosynthesis) occurs; prior to these optimum conditions (particularly in the absence of light) an initial efflux is observed (Evans and Johansen, 1999; Thomas et al., 2011; Thomas, 2012).

Free Air Carbon Enrichment (FACE) facilities have been set up in several dryland study sites to investigate the effects of increasing atmospheric CO₂ on vascular vegetation and soils (e.g. de Soyza et al., 2005; Steven et al., 2012). For example, ten years of FACE field studies in Nevada suggest that enhanced CO₂ treatments cyanobacteria will show a generalized decline with elevated CO₂ though the direct impacts on CO₂ efflux rates remains unknown (Steven et al., 2012). A similar experiment in the Mojave desert conveys the importance of BSCs in the uptake of CO₂; whereby in post rain conditions the net ecosystem CO₂ exchange (uptake) was unexpectedly high and was likely due to autotrophic mechanisms as vascular plant photosynthesis was low at this time (Jasoni et al., 2005).

The aim of this experimental study was to quantify the short-term temporal changes in carbon exchange (respiration and photosynthesis) of BSCs when the atmospheric concentration of CO₂ was doubled (~800 ppm). The study was conducted using specially designed dynamic gas exchange chambers (DGECs) and cyanobacteria-dominated BSCs sampled from Kalahari Sand soils of southern Africa, a large dryland biome which covers 2.5 million hectares (Wang et al., 2007). In addition to the doubled CO₂ atmospheric level in laboratory studies, two wetting treatments (2 and 5 mm) were incorporated. This also added an extra dimension and additional data on BSCs responses to key conditions of both soil moisture and atmospheric CO₂ concentration.

2. Methods

2.1 Culturing BSCs in Controlled Conditions

Both crusts and subsoil samples were collected from a lightly grazed commerciallyowned farm near Tsabong, Botswana (25°56'51"S 22°25'40"E) during a field campaign in summer 2008. The subsoil (97% fine sand, pH 5.9 \pm 0.4, bulk density 1.85 ± 0.03 g cm³ and porosity 0.34 ± 0.01 v/v) (Thomas and Hoon, 2010) was sieved (853 µm), sterilized (autoclaved at 121 °C for 15 min) and then filled (178 g each) in nine polyethylene terephthalate (PET) containers (diameter = 93 mm and depth = 25 mmmm). Soil depth was approximately 20 mm, a depth such that a 5 mm wetting would result in near saturation. On the soil surface, biological soil crust samples were gently broken into similarly sized pieces and distributed evenly and equally (3.32 g per container) in all PET containers. Samples were weighed again and then wetted using 10 ml of pure water (purity > 99%) applied using a calibrated water sprayer. This initial wetting consolidated the soil and the samples were left in an environmental cabinet (the conditions of which are outlined below) to settle into a circadian rhythm. The samples were allowed to establish for two months in a growth cabinet and the crust surfaces were gently moistened several times a week (1-5 mm each), a total of 53.2 mm (361 ml) of water per container. Included in several of these treatments was the addition of standard nutrient medium, BG11 (Rippka et al., 1979), which was diluted with water, to encourage growth of the cyanobacteria.

The growth cabinet was set to provide 16 hours of light (intensity = 76 ± 1 µmol m⁻² s⁻¹, measured as an average at the central position of each cabinet shelf) and a temperature of 28 ± 0.3 °C; the remaining 8 hours with no light at a temperature of 10 ± 0.3 °C. The light intensity in the cabinet represented early morning hours during austral winter in the Kalahari where maximum microbial activity occurs (Evans and Johansen, 1999; Thomas and Hoon, 2010). Four USB loggers (USB-500

Measurement Computing Corporation) were placed in various positions within the cabinet, in order to record internal temperature, which ranged from 7.5-30.5°C in phase 1 and 7.5-29.0°C in phase 2. Relative humidity was not recorded because the sensors were located within the environmental cabinet and not in the enclosed and therefore isolated (to external relative humidity) design of the DGECs.

2.2 Dynamic Gas Exchange Chambers (DGECs)

The design of the chambers used was developed from that proposed and used successfully in field conditions by Hoon et al. (2009). Chambers were designed to house the PET containers containing crust samples. Chambers were made using transparent Perspex® material (3 mm thickness) with a height of 103 mm and 100 mm outer diameter, allowing 92% transmission of PAR (400 nm-700 nm), transmission of thermal infrared (IR) and filtered ultraviolet radiation (UV) < 300 nm; the top of the chamber was fitted with a removable cap made of Teflon® film that allowed 96% transmission of incoming radiation (Figure 1). The Teflon® film was able to accommodate pressure changes within the chambers by acting as a diaphragm. Each chamber was made air-tight and had ports (Suba-Seals®) from which samples were collected at regular intervals. Chambers were attached (via the gas inlets) to the main vessel of 32 l volume, containing the CO₂-air mix. Open-cell foam inserts were placed into the gas inlets and outlets in order to prevent diffusion of atmospheric gases into or out of the chamber, while permitting the flushing of the chamber and equilibration with the external atmospheric pressure, should it become over pressurized. A Teflon® bag was attached to the end of the flush-gas feed line in order to confirm flow of the flush gas. Final checks were made for any leaks. After this, PET containers with crust samples were transferred into the DGECs and placed back

inside the growth cabinet. Care was taken to minimize the disturbance to the crust samples. The final set up is given in Figure 2.

2.3 Treatments and Sampling Regime

Experiments were conducted in two phases (3 days each, as during the preliminary run, mentioned below, the behavior had stabilized by the end of day three) to study the impacts of elevated CO₂ along with different wetting treatments. Phase 1 was carried out under ambient CO₂ level (392 ppm in air, with a standard deviation of 4.4 ppm, blended in-laboratory from pure dry air and 1003 ppm CO₂-air mix blended by Air Liquide) with three single wetting treatments (0, 2 and 5 mm) with three replicates. After phase 1, covers were removed and water added to the 0 and 2 mm treatments (5 and 3 mm water, respectively) in order to equalize the moisture content with the 5 mm wetting treatments. Covers were left off and the crusts allowed to dry out again. They reached the initial moisture content 4-5 days after this and were then moistened once a week before commencement of phase 2 (phase 2 was undertaken three months after phase 1). Phase two was undertaken using the same method as phase 1, on the same crust samples, this time with elevated CO₂ (801 ppm in air, blended by Air Liquide).

Approximately 12 hours prior to experiments, the CO₂-air mix was used to flush the chambers. This was in order to allow diffusion through the air spaces in the soil, such that the concentration gradient between subsoil and atmosphere was reduced and to acclimatize the crust to the new environment. The first gas sample was taken (0815 GMT) the following morning, two hours before lights on (1015 GMT). Samples were then weighed and the Teflon® lids opened for wetting (simulating a single rainfall event of 2 or 5 mm) using a Jencons Powerpette for even application.

The chambers were immediately weighed and closed; then placed back into the growth cabinet and re-attached to the gas line. All chambers were again flushed with the CO₂-air mix. Chambers were then left for approximately 40 min before gas samples were taken, followed by flushing; with the exception of the last flush of each day which was left overnight and the gas sample taken immediately the following morning. Each flushing was 12 min (this was determined during initial DGEC design testing in the laboratory). Chamber venting time was close to optimum (40 min), which has been determined by a sensitivity analysis of static gas chambers on soils (Ohlsson et al., 2005), in which it was concluded that it is possible to reduce the uncertainty in the CO₂ flux value by allowing extended CO₂ accumulation/reduction times. Nine gas samples (10 ml each) were collected per chamber at regular intervals during the day (0815-1715 GMT with at least two samples taken before lights on) for three days and a total of 243 samples were collected from each phase of this experiment. A preliminary run (with 410 ppm CO₂-air mix) was carried out prior to the experiment reported here to check the system.

Sample weights were taken only at the beginning and end of each phase in order to quantify moisture loss. Moisture loss during the experiment was found to be negligible in the DGECs, due to the chambers being closed.

2.4 Analysis of CO₂ Using Gas Chromatography with Helium Ionization Detection

Temporal changes in CO_2 concentration within the DGECs reflect the activity (photosynthesis and respiration) of biological soil crusts. To measure the changes in CO_2 levels, 2.5 μ l gas samples were taken at approximately 40 minute intervals throughout the experiment and were separated using a Porous Layer Open Tubular

(PLOT) column (PoraPLOT Q, 25 m x 0.53 mm, 20 μ m i.d. Chrompack, Netherlands). A helium ionization detector (VICI, Houston, USA) was used to quantify the eluate. The dimensionless integrals of the CO_2 peaks were converted to ppm and then to carbon flux in mg C m⁻² hr⁻¹ using Equation 1;

$$C \; Flux \; Rate = \left\{ \left(\left[CO_2 \right]_{diff} \times n \right) / \left(A \times t \right) \right\} \times V_s \times M_C \times 3,600 \; mg \; C \; m^{\text{--}2} \; hr^{\text{--}1} \; \; (1)$$

where $[CO_2]_{diff}$ = difference between measured CO_2 concentration and flush CO_2 concentration in ppm; n = number of moles of gas present in loop at temperature T (°C);

A = surface area of soil sample $(6.793\times10^{-3}~\text{m}^2)$; t = time between end of last flush and taking of gas sample (s); V_s = volumetric scaling factor = 222,072; M_C = molar mass of carbon (12.0107 g mol⁻¹) and 3,600 is used to represent flux over one hour. Net carbon balances were based on the area under each carbon flux time series and were calculated separately for each chamber for each day using the integral function in EasyPlotTM software. In order to characterize significance of treatments (two levels of CO_2 and three levels of wetting), analysis of variance (ANOVA) was undertaken using statistical package SYSTAT 13 which determines the p-value for CO_2 , wetting and also whether or not these values are statistically significant.

3. Results

3.1 Carbon fluxes in absence of liquid water (controls)

Temporal changes in CO_2 concentrations in the chambers were measured to understand the short-term responses to elevated CO_2 levels along with pulse wetting events. The positive values of flux in Figures 3 – 5 indicate net respiration and negative values indicate net photosynthesis. Net carbon balances for each individual chamber, on each day of both phases can be seen in Tables 1 & 2.

Figures 3a and 3b show temporal changes in the fluxes in the dry chambers (controls) for phase 1 (392 ppm CO_2) and phase 2 (801 ppm CO_2), respectively. The activity of the crust was very low due to there being no liquid moisture available. For phase 1, the carbon loss (net respiration) dominated over the gain of carbon (net photosynthesis) whereas for phase 2, the opposite was found. Tables 1 & 2 show that, interestingly, many of the replicates exhibited appreciable carbon fixation even when no liquid moisture was available (with the exception of air moisture). However, the net carbon budgets over three days showed that, 0.02 ± 0.01 mg C was lost in phase 1 and 0.39 ± 0.04 mg C was sequestered for phase 2.

3.2 Carbon fluxes under 2 mm wetting treatment

Figures 4a & 4b display the carbon release during the first day from the samples soon after the wetting treatment for both phases. The samples continued respiration for the majority of the first day and settled into a photosynthetic regime by the end of the day. The second and third days show the samples to be predominantly photosynthetic in activity. This photosynthetic activity peaked toward the end of each day. The quantity of liquid moisture available for the samples for each of the three days, remained largely unchanged. Tables 1 & 2 show the net carbon balances (in mg) of all chambers for each day of each phase. After a 2 mm wetting treatment and subject to the conditions of phase 1 the crust sequestered a total of 0.14 ± 0.14 mg C and during

phase 2 this was 1.3 ± 0.5 mg C. All cell group averages displayed net sequestration for both phases under 2 mm wetting, with phase 2 showing the highest quantity of carbon sequestered.

3.3 Carbon fluxes under 5 mm wetting treatment

When the 5 mm wetting treatment was applied the wetting front reached the bottom of each container, therefore saturating the soil. Moisture remained available for the samples for each of the three days, and the quantity remained largely unchanged. Figures 5a & 5b display the carbon fluxes of the chambers subject to a 5 mm wetting treatment for each phase. It can be seen that throughout each experiment, a pulse of carbon was released at the beginning of each day, particularly during phase 1, indicating respiration. The samples exhibited very similar behavior to those subject to 2 mm wetting; the main difference being a moderate carbon pulse at the beginning of each day and overall larger fluxes. For phase 2, the carbon losses were found to be similar to phase 1; however the peak influx (sequestration) was 2-3 times larger than phase 1. Tables 1 & 2 show the net carbon balances (in mg) of all chambers for each day of each phase. After a 5 mm wetting treatment and subject to the conditions of phase 1, 0.62 ± 0.05 mg C was sequestered and during phase 2, 2 ± 0.2 mg C was sequestered. Phase 2 showed the highest quantity of carbon sequestered.

ANOVA significance testing of daily carbon sequestration (i.e. all negative values from Tables 1 & 2) was undertaken. Tables 1 & 2 show the highly significant effects of CO_2 (P value = 0.005) and wetting (P value = 0.001), however the interaction between these factors was not found to be significant.

4. Discussion

All carbon balances, with the exception of the dry samples at 392 ppm CO₂, were negative. This suggests that the overall, the dominant process was photosynthesis. This shows that isolation of the autotrophs has been achieved, although it should be noted that BSCs colonies are an amalgamation of heterotrophs and autotrophs, and as such there is potential for both photosynthesis and respiration to occur simultaneously. This is further confirmed when considering the pulse CO₂ efflux effects observed in those samples subject to wetting treatments, although interestingly initial pulses were lower under 801 ppm CO₂ than under 392 ppm CO₂ perhaps suggesting that the CO₂ concentrating mechanism (CCM) (Badger and Price, 1992) activity was enhanced under the increased CO₂ atmosphere. Respiration was observed early each day in the samples subjected to wetting treatments due to the absence of light (and in some cases lasted for a short duration after the lights were on); again these effluxes were lower under higher ambient CO₂ levels.

Overall, each of the wetting treatments yielded higher influx of carbon when subject to an 801 ppm CO₂-air mix than when subject to the 392 ppm CO₂-air mix. Of these the most significant were the samples subjected to a 2 mm wetting treatment; carbon influx increased by an order of magnitude suggesting that photosynthetic activity in the BSCs samples responded best to 'light' wetting treatments/events. Increases in carbon influx of three fold were observed in those samples with a 5 mm wetting treatment, still representing a significant relative increase in carbon influx under increased atmospheric CO₂. Samples which had no additional water added also exhibited increased carbon influx under enhanced atmospheric CO₂; this is very likely due to moisture formation in the form of condensation on the soil surface during the

'night-time' part of the cycle creating a source of liquid moisture for the BSCs, facilitating photosynthesis.

The finding that the relative increase in carbon influx was highest in the 2 mm wetting treatment is complementary to the field-based findings of Thomas et al. (2008) whereby not only do the autotrophs respond best to 'light' wetting treatments, but the light wetting also limits the depth of soil which receives moisture, hence reducing the effect that subsoil heterotrophs have on the net carbon balance. Light rainfall events are common across the Kalahari, particularly at the beginning of the wet season and have an important effect upon the annual soil carbon balance (Thomas et al., 2011). The results presented here are promising for the continuing role of BSCs as a carbon store as atmospheric CO₂ increases, provided rainfall patterns remain at higher frequencies of light events as currently seen (Simelton et al., 2013) rather than moving to more extreme intense rainfall events as predicted by some for dryland Africa more widely (e.g. Twomlow et al., 2008)).

Although the mechanisms of increased carbon sequestration were not investigated in this study, Badger and Price (1992) describe the CO₂ concentrating mechanisms of cyanobacteria, which are capable of concentrating CO₂ up to 1000-fold. The cyanobacterial CO₂ concentrating mechanism is possibly the most effective of any photosynthetic organism (Badger and Price, 1992), and therefore it would be hoped that this effect would be amplified if elevated levels of CO₂ were made available (up until saturation). Cyanobacterial species are among the earliest forms of life, having evolved under, and been exposed to, the high CO₂/low O₂ atmosphere of the early (~3,000 Ma) Earth (Bowes, 1991); therefore it could be postulated that cyanobacterial photosynthetic activity is optimum when exposed to the conditions under which the organisms evolved, and hence a positive correlation between

photosynthetic activity and increasing CO_2 levels may exist. The findings of this study appear consistent with this hypothesis; however the exact mechanisms require further investigation. Further evidence is the ability of CO_2 concentrating mechanism to modulate activity under the influence of environmental factors; cyanobacteria (and other CO_2 concentrating mechanisms) can acclimate to a wide range of CO_2 concentrations (as also show in FACE studies of Steven et al., 2012).

This study demonstrated the carbon sequestration potential of BSCs under elevated CO₂ levels under two moisture regimes. The response was similar to many crops (Cure and Acock, 1986) and dryland vegetation (Naumburg et al., 2003) under elevated levels of CO₂ in the atmosphere. BSCs have the potential to fix carbon under limited soil moisture availability and nutrient poor soils (typical of drylands) and is especially applicable to BSCs dominated by nitrogen fixing cyanobacteria. Therefore undisturbed BSC-covered drylands could be enhanced carbon sinks, and play an increasingly significant role in global carbon budgets in years to come. With this in mind, protection of BSCs is of increasing importance, particularly considering land use pressures on these areas (Dougill and Thomas, 2004), where implementation of community-based ecosystem service schemes such as those described by Dougill et al. (2012) could be greatly beneficial to dryland areas. More long-term flux monitoring studies are required with varying environmental conditions, particularly concerning BSCS responses to increased temperatures (e.g. Frey et al., 2013) and incoming solar radiation with varying wetting treatments.

Acknowledgements

The author would like to extend much gratitude to Mr. Lance Penketh of the University of Leeds during the course of incubation and experimentation using the growth cabinets. The author is grateful for the funding provided by the Natural Environment Research Council (NERC) Taught Postgraduate Fund, with which this research work was carried out.

References

Badger, M., Price, G., 1992. CO₂ concentrating mechanisms in cyanobacteria: molecular components, their diversity and evolution. J. Exp. Botany. 54, 609-622.

Belnap, J., Kaltenecker, J., Rosentreter, R., Williams, J., Leonard, S., Eldridge, D., 2001. Introduction to Biological Soil Crusts. Biological Soil Crusts: Ecology and Management, Technical Reference. Denver: United States Department of the Interior, pp. 1-7.

Belnap, J., Lange, O., 2003. Preface. In: Belnap J & Lange OL (Eds.), Biological Soil Crusts: Structure, Function, and Management. Berlin: Springer-Verlag, pp. V-IX.

Beymer, R., Klopatek, J., 1991. Potential contribution of carbon by microphytic crusts in pinyon-juniper woodlands. Arid Soil Research and Rehabilitation. 5, 187-198.

Bombelli, A., Henry, M., Castaldi, S., Adu-Bredu, S., Arneth, A., de Grandcourt, A., Grieco, E., Kutsch, W., Lehsten, V., Rasile, A., Reichstein, M., Tansey, K., Weber, U., Valentini, R., 2009. An outlook on the Sub-Saharan Africa carbon balance. Biogeosci. 6, 2193-2205.

Borken, W., Matzner, E., 2009. Reappraisal of drying and wetting effects on C and N mineralisation and fluxes in soils. Global Change Biol. 15, 808-824.

Bowes, G., 1991. Growth at elevated CO₂: photosynthetic responses mediated through Rubisco. Plant, Cell and Env. 14, 795-806.

Brostoff, W., Sharifi, M., Rundel, P., 2001, Photosynthesis of cryptobiotic crusts in a seasonally inundated system of pans and dunes at Edwards Air Force Base, western Mojave Desert, California: Laboratory studies, Flora, 197, 143-151.

Büdel, B., Darienko, T., Deutschewitz, K., Dojani, S., Friedl, T., Kohr, K., Salisch, M., Reisser, W., Weber, B., 2009. Southern African biological soil crusts are ubiquitous and highly diverse in drylands, being restricted by rainfall frequency. Microb. Ecol. 57, 229-247.

Cable, J., Huxman, T., 2004. Precipitation pulse size effects on Sonoran Desert soil microbial crusts. Oecologia. 141, 317-324.

Castillo-Monroy A.P, Maestre F.T, Rey A., Soliveres S., García-Palacios P. (2011) Biological soil crust microsites are the main contributor to soil respiration in a semiarid ecosystem. Ecosystems, 14, 835-847.

Ciais, P., Piao, S.-L., Cadule, P., Friedlingstein, P., Chéduin, A., 2009. Variability and recent trends in the African terrestrial carbon balance. Biogeosci. 6, 1935-1948.

Cure, J., Acock, B., 1986. Crop Responses to Carbon Dioxide Doubling: A Literature Survey. Agric. and Forest Meteo. 38, 127-145.

de Soyza, A., Nowak, R., Knight, E., Babcock, D., Smith-Longozo, V., 2005. Effects of elevated atmospheric CO₂ on soil respiration in a Mojave Desert ecosystem.

Dougill, A.J., Stringer, L.C., Leventon, J., Riddell, M., Rueff, H., Spracklen, D., Butt, E., 2012. Lessons from community-based payment for ecosystem service schemes: from forests to rangelands. Phil. Trans. Royal Soc. Biol. Sci. 367, 3178-3190.

Dougill, A.J., Thomas, A.D., 2004. Kalahari Sand Soils: Spatial Heterogeneity, Biological Soil Crusts and Land Degradation. Land Deg. & Devel. 15, 233-242.

Elbert, W., Weber, B., Burrows, S., Steinkamp, J., Büdel, B., Andreae, M., Pöschl, U., 2012. Contribution of cryptogamic covers to the global cycles of carbon and nitrogen. Nat. Geosci. 5, 459-462.

Evans, R., Johansen, J., 1999. Microbiotic Crusts and Ecosystem Processes. Crit. Rev. Plant Sci. 18, 183-225.

Frey, S.D., Lee, J., Melillo, J.M., Six, J., 2013. The temperature response of soil microbial efficiency and its feedback to climate. Nature Climate Change. 3, 395-398.

Garcia-Pichel, F., Belnap, J., Neuer, S., Schanz, F., 2003. Estimates of global cyanobacterial biomass and its distribution. Algological Stud. 109, 213-227.

Hoon, S., Thomas, A.D., Linton, P., 2009. The Design and Development of a Closed Chamber for the in-situ Quantification of Dryland Soil Carbon Dioxide Fluxes. Geog. Res. 47, 71-82.

Huxman, T., Snyder, K., Tissue, D., Leffler, A., Ogle, K., Pockman, W., Sandquist, D., Potts, D., Scwinning, S., 2004. Precipitation pulses and carbon fluxes in semiarid and arid ecosystems. Oecologia. 141, 254-268.

Jasoni, R., Smith, S. Arnone III, J., 2005. Net ecosystem CO₂ exchange in Mojave Desert shrublands during the eight year of exposure to elevated CO₂. Global Change Biol. 11, 749-756.

Jeffries, D., Link, S., Klopatek, J., 1993. CO₂ fluxes of cryptogamic crusts. New Phytol. 125, 163-173.

Kaplan, A., Reinhold, L., 1999. CO₂ Concentrating Mechanisms in Photosynthetic Microorganisms. Ann. Rev. Plant Phys. and Plant Mol. Biol. 50, 539-570.

Lal, R., 2004. Carbon Sequestration in Dryland Ecosystems. Env. Man. 33, 528-544.

Lange, O., 2003. Photosynthesis of Soil-Crust Biota as Dependent on Environmental Factors. In: Belnap, J. & Lange, O. (Eds.), Biological Soil Crusts: Structure, Function, and Management. Berlin: Springer-Verlag, pp. 218-240.

Naumburg, E., Housman, D., Huxman, T., Charlet, T., Loik, M., Smith, S., 2003. Photosynthetic responses of Mojave Desert shrubs to Free Air CO₂ Enrichment are greatest during wet years. Glob. Change Biol. 9, 276-285.

Ohlsson, K., Singh, B., Holm, S., Nordgren, A., Lövdahl, L., Högberg, P., 2005. Uncertainties in static closed chamber measurements of the carbon isotopic ratio of soil respired CO₂. Soil Biol. and Biochem. 37, 2273-2276.

Maestre, F.T. Quero, J.L. et al., Gotelli, N.J., Escudero, A., Ochoa, V., et al., 2012a. Plant Species Richness and Ecosystem Multifunctionality in Global Drylands. Science, 335, 214-218.

Maestre, F.T., Salguero-Gomez, R., Quero, J.L., 2012b. It is getting hotter in here: determining and projecting the impacts of global environmental change on drylands. Phil. Trans. R. Soc. B, 367, 3062-3075.

Reynolds, J., Smtih, D., Lambin, E., Turner II, B., Mortimore, M., Batterbury, S., Downing, T., Dowlatabadi, H., Fernández, R., Herrick, J., Huber-Sannwald, E., Jiang, H., Leemans, R., Lynam, T., Maestre, F., Ayarza, M., Walker, B., 2007. Global Desertification: Building a Science for Dryland Development. Science. 316, 847-851.

Rippka, R., Deruelles, J., Waterbury, J., Herdman, M., Stanier, R., 1979. Generic assignments, strain histories and properties of pure cultures of cyanobacteria. J. General Microbiol. 111, 1-61.

Simelton, E., Quinn, C.H., Batisani, N., Dougill, A.J., Dyer, J.C., Fraser, E.D.G., Mkwambisi, D.D. Sallu, S.M., Stringer, L.C. 2013. Is rainfall really changing? Farmers' perceptions, meteorological data and policy implications. Climate and Development, 5(2), 123-138.

Steven, B., Gallego-Graves, L., Yeager, C., Belnap, J., Evans, R.D., Kuske, C., 2012. Dryland biological soil crust cyanobacteria show unexpected decreases in abundance under long-term elevated CO₂. Environmental Microbiology, 14(12), 3247-3258.

Stringer, L.C., Dougill, A.J., Thomas, A.D., Spracklen, D.V., Chesterman, C., Ifejika-Speranza, C., Rueff, H., Riddell, M., Williams, M., Beedy, T., Abson, D., Klintenberg, P., Syampungani, S., Powell, P., Palmer, A.R., Seely, M., Mkwambisi, D.D., Falcao, M., Sitoe, A., Ross, S., Kopolo, G. 2012. Challenges and opportunities in linking carbon sequestration, dryland livelihoods and ecosystem service provision. Environmental Science and Policy, 19-20,121-135.

Thomas, A.D., 2012. Impact of grazing intensity on seasonal variations of soil organic carbon and soil CO₂ efflux in two semi-arid grasslands in southern Botswana. Phil. Trans. Royal Soc. Biol. Sci. 367, 3076-3086.

Thomas, A.D., Hoon, S., 2010. Carbon dioxide fluxes from biologically-crusted Kalahari Sands after simulated wetting. J. Arid Env. 74, 131-139.

Thomas, A.D., Hoon, S., Dougill, A.J., 2011. Soil respiration at five sites along the Kalahari Transect: Effects of temperature, precipitation pulses and biological soil crust cover. Geoderma. 167-168, 284-294.

Thomas, A.D., Hoon, S., Linton, P., 2008. Carbon dioxide fluxes from cyanobacteria crusted soils in the Kalahari. Appl. Soil Ecol. 39, 254-263.

Twomlow, S., Mugabe, F.T., Mwale, M., Delve, R., Nanja, D., Carberry, P., & Howden, M., 2008. Building adaptive capacity to cope with increasing vulnerability

due to climatic change in Africa – A new approach. Physics and Chemistry of the Earth, 33, 780–787.

Wang, L., D'Odorico, P., Ringrose, S., Coetzee, S., Macko, S., 2007. Biogeochemistry of Kalahari sands. J. Arid Env. 71, 259-279.

FIGURE CAPTIONS

Figure 1. Photograph of the dynamic gas exchange chamber. A Teflon® film was attached to the top of the chamber using strong adhesive tape, and the gas inlet/gas outlet had open-cell foam inserts in order to allow pressure equilibration, but to prevent gas diffusion. Gas inlets/outlets were attached using grommets which were sealed using Instant Gasket. Gas samples were taken from the Suba-Seal® sampling port (to the rear of the chamber in this image). A hole was drilled in the bottom of the chamber in order that the samples could be inserted into the cell in a controlled manner; during experiments this hole was sealed using high strength adhesive tape.

Figure 2. Layout of DGECs within controlled environment cabinet. The USB data loggers (not shown) were placed at various points within the cabinet (one on top and bottom shelf, two on middle shelf). A flow rate meter (not shown) was attached to the gas feed line close to the point of exit of the main 32 l vessel.

Figure 3a & 3b. Gas sample data from phases 1 & 2, samples with no additional liquid water, data presented as average values of three replicates for each treatment with error bars representing standard error. Gas sample data from phases 1 (Figure 3a) & 2 (Figure 3b) (392 ppm & 801 ppm CO_2 , respectively, in air flush-gas), measuring carbon fluxes in chambers subject to no additional liquid water. Night-time (lights off) was 8 hours before 1 hour, between 17-25 hours and between 41-49 hours, after wetting. The left y-axis (Y1) corresponds to the solid dots / line (carbon flux of crust in chamber averaged over three chambers) and the right y-axis (Y2) corresponds to the dashed curve (internal temperature of cabinet). The Y1=0 line represents the point at which net carbon flux is zero. Y1=+ve \rightarrow net respiration, Y1=-ve \rightarrow net photosynthesis. The point at which water was added to the crusts is given by the line X=0.

Figure 4a & 4b. Gas sample data from phases 1 & 2, samples subject to 2 mm wetting treatment, data presented as average values of three replicates for each treatment with error bars representing standard error. Gas sample data from phases 1 (Figure 4a) & 2 (Figure 4b) (392 ppm & 801 ppm CO₂, respectively, in air flush-gas), measuring carbon fluxes in chambers subject to a 2 mm wetting treatment. Night-time (lights off) was 8 hours before 1 hour, between 17-25 hours and between 41-49 hours, after wetting. The left y-axis (Y1) corresponds to the solid dots / line (carbon flux of crust in chamber – averaged over three chambers) and the right y-axis (Y2) corresponds to the dashed curve (internal temperature of cabinet). The Y1=0 line represents the point at which net carbon flux is zero. Y1=+ve → net respiration, Y1=-ve → net photosynthesis. The point at which water was added to the crusts is given by the line X=0.

Figure 5a & 5b. Gas sample data from phases 1 & 2, samples subject to 5 mm wetting treatment, data presented as average values of three replicates for each treatment with error bars representing standard error. Gas sample data from phases 1 (Figure 5a) & 2 (Figure 5b) (392 ppm & 801 ppm CO_2 , respectively, in air flushgas), measuring carbon fluxes in chambers subject to a 5 mm wetting treatment. Night-time (lights off) was 8 hours before 1 hour, between 17-25 hours and between 41-49 hours, after wetting. The left y-axis (Y1) corresponds to the solid dots / line (carbon flux of crust in chamber – averaged over three chambers) and the right y-axis (Y2) corresponds to the dashed curve (internal temperature of cabinet). The Y1=0 line represents the point at which net carbon flux is zero. Y1=+ve \rightarrow net respiration, Y1=-ve \rightarrow net photosynthesis. The point at which water was added to the crusts is given by the line X=0.

*Highlights (for review)

HIGHLIGHTS

- Response to elevated atmospheric CO₂ of cyanobacterial samples was investigated
- Three wetting treatments were applied under fixed temperature and light conditions
- Newly designed dynamic gas exchange chambers were developed and used
- Samples subjected to 2 mm wetting sequestrated 10 times more C when CO₂ was doubled
- Samples subjected to 5 mm wetting sequestrated 3 times more C when CO₂ was doubled

Replacement_Figure1(COLOUR)
Click here to download high resolution image



replacement_Figure1(MONOCHROME) Click here to download high resolution image



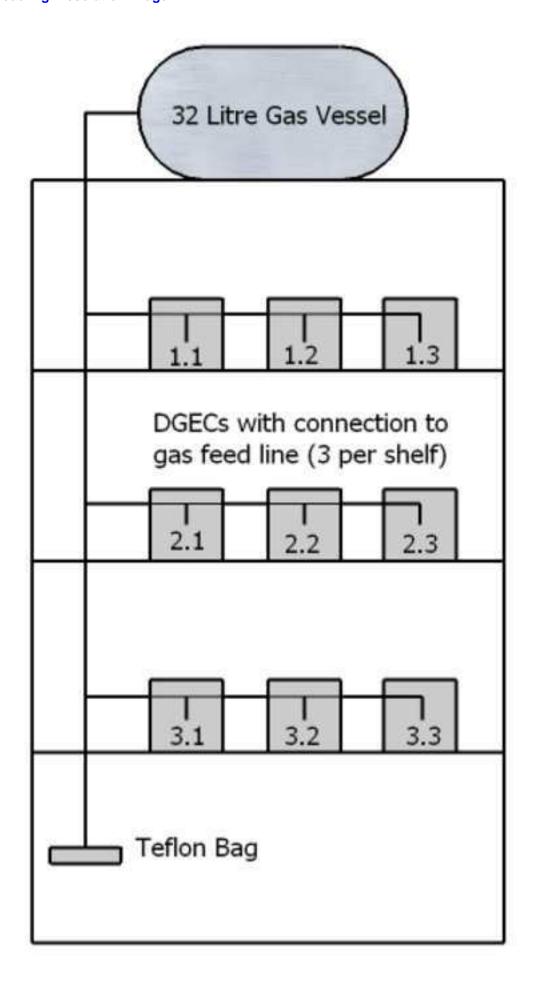


Figure3a Click here to download high resolution image

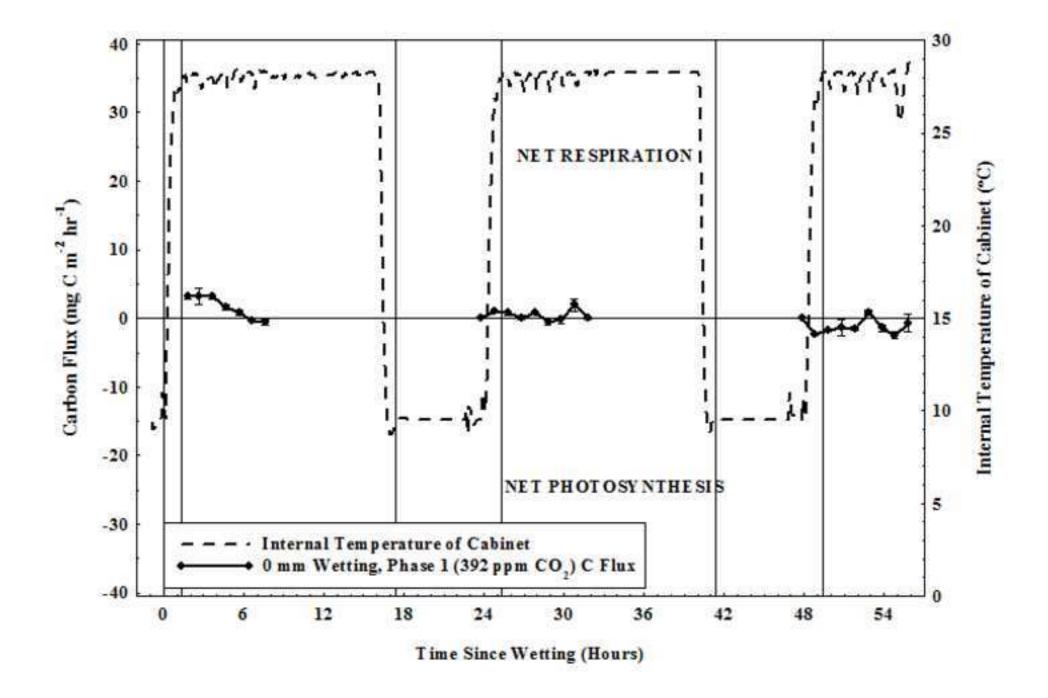


Figure3b Click here to download high resolution image

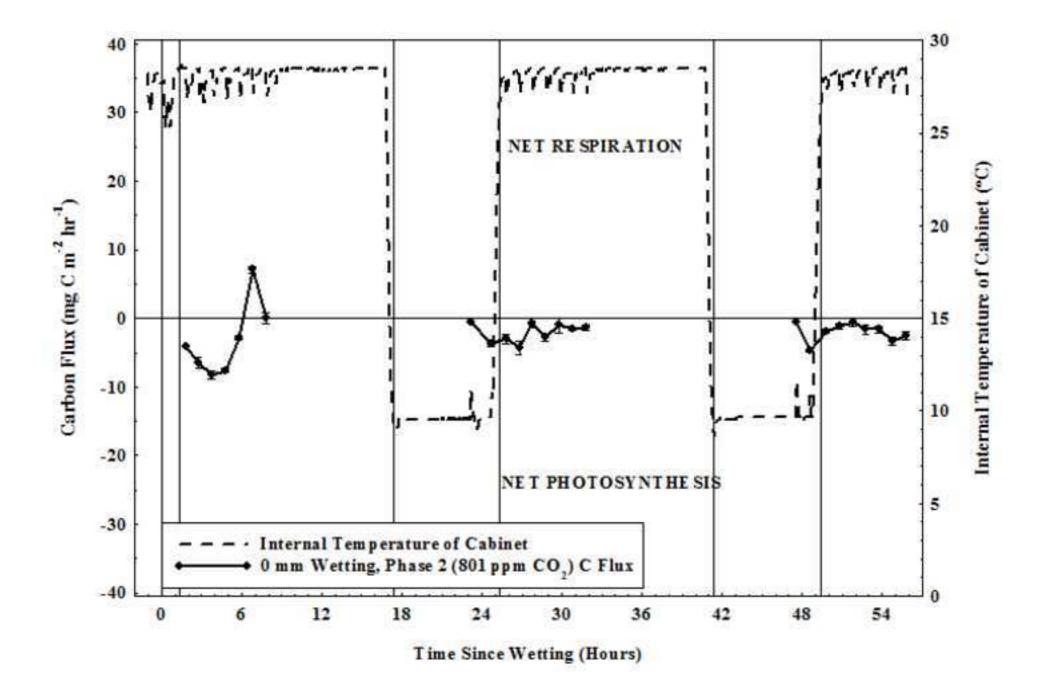


Figure4a Click here to download high resolution image

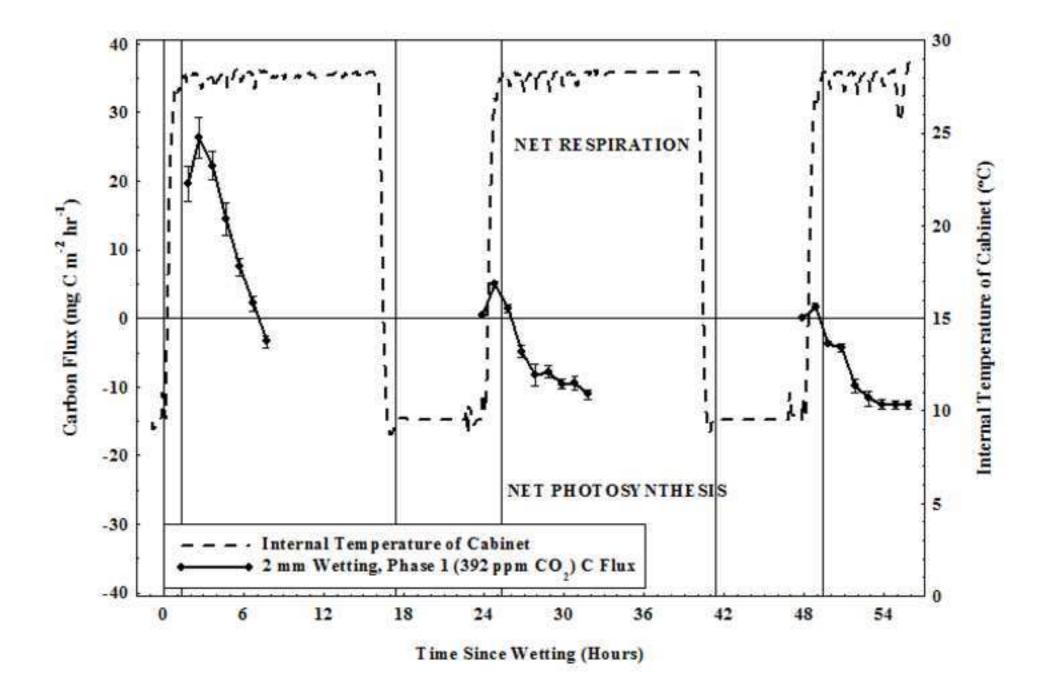


Figure4b Click here to download high resolution image

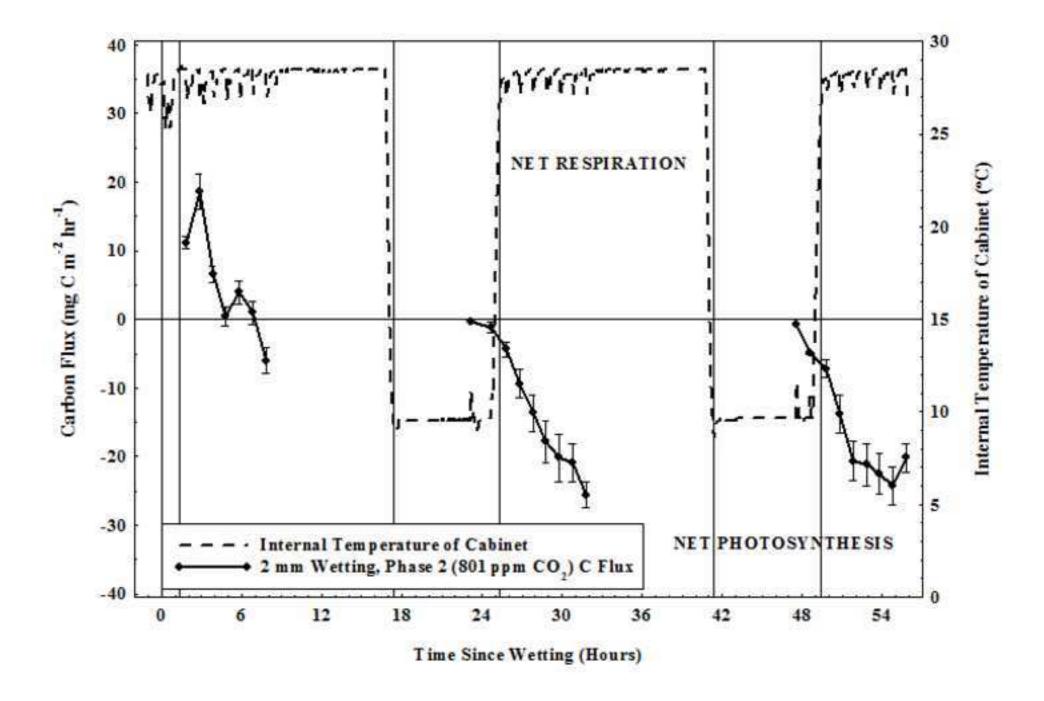


Figure5a
Click here to download high resolution image

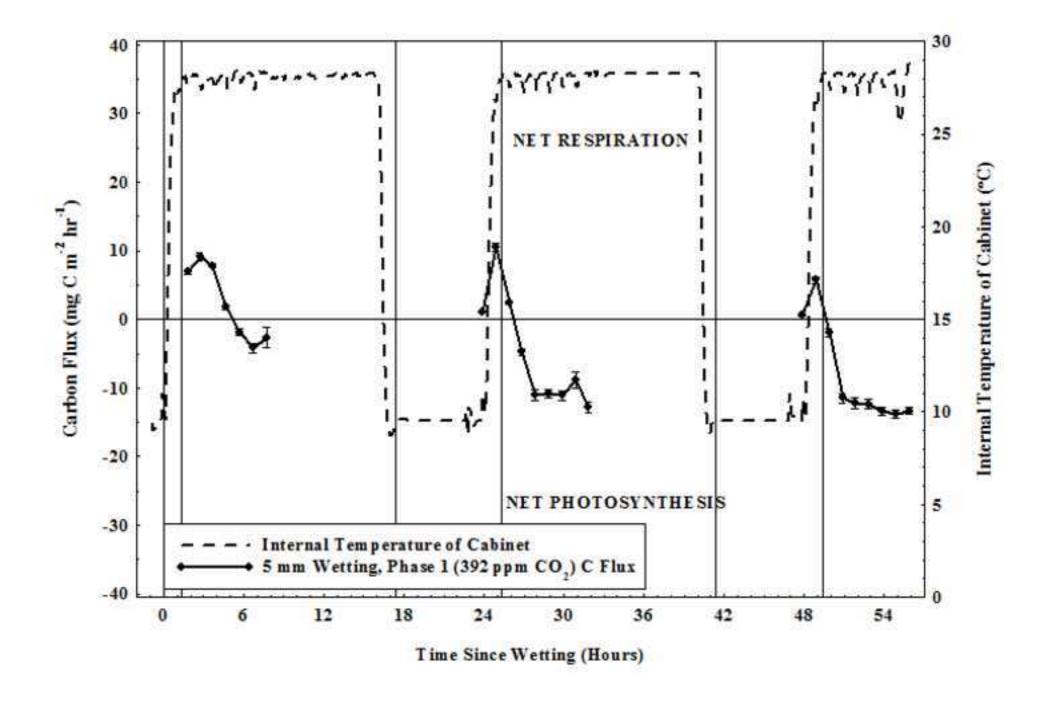
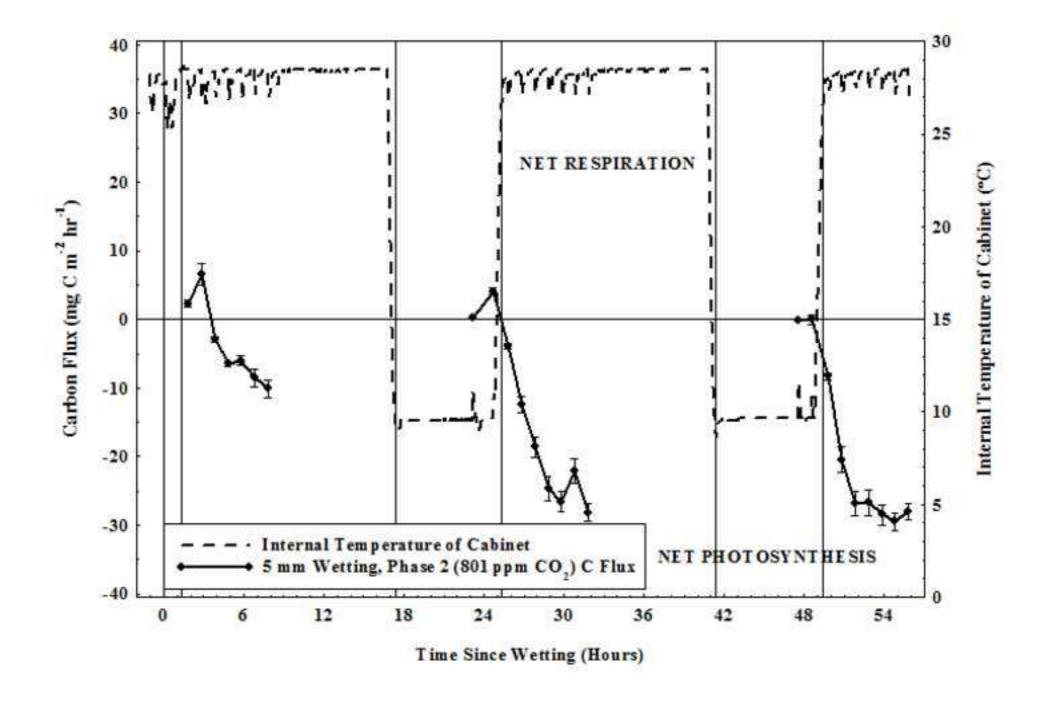


Figure5b Click here to download high resolution image



TABLES

Table 1. Carbon balances for each chamber during Phase 1 (392 ppm CO₂). Chambers 1.2-1.3 had no additional water, chambers 2.1-2.3 each had a 2 mm wetting treatment and chambers 3.1-3.3 each had a 5 mm wetting treatment.

3.5 cuch had a 5 hin wetting treatment.									
	Carbon Balance (mg C)								
	(+ve = efflux, -ve = sequestration)								
					Average for each chamber group				
Chamber	Day 1	Day 2	Day 3	Net	(±SE)				
1.1	0.119	0.021	-0.098	0.041					
1.2	0.030	0.013	-0.024	0.020					
1.3	0.033	0.046	-0.082	-0.003	0.02 ± 0.01				
2.1	0.431	-0.163	-0.330	-0.062					
2.2	0.775	-0.300	-0.421	0.055					
2.3	0.354	-0.324	-0.450	-0.420	-0.14 ± 0.14				
3.1	0.120	-0.307	-0.420	-0.607					
3.2	0.085	-0.199	-0.428	-0.541					
3.3	0.068	-0.293	-0.487	-0.712	-0.62 ± 0.05				

Table 2. Carbon balances for each chamber during Phase 2 (801 ppm CO₂). Chambers 1.2-1.3 had no additional water, chambers 2.1-2.3 each had a 2 mm wetting treatment and chambers 3.1-3.3 each had a 5 mm wetting treatment.

	Carbon Balance (mg C)							
	(+ve = e	fflux, -ve =						
					Average for each chamber group			
Chamber	Day 1	Day 2	Day 3	Net	(±SE)			
1.1	-0.109	-0.096	-0.163	-0.369				
1.2	-0.174	-0.190	-0.102	-0.466				
1.3	-0.125	-0.114	-0.082	-0.321	-0.39 ± 0.04			
2.1	0.045	-0.703	-0.880	-1.538				
2.2	0.317	-1.067	-1.217	-1.966				
2.3	0.304	-0.284	-0.468	-0.449	-1.3 ± 0.5			
3.1	-0.145	-0.621	-0.845	-1.611				
3.2	-0.076	-0.811	-1.115	-2.001				
3.3	-0.215	-0.946	-1.194	-2.356	-2.0 ± 0.2			