Global diversity and geography of soil fungi

3	Leho Tedersoo ¹ *†, Mohammad Bahram ² †, Sergei Põlme ¹ , Urmas Kõljalg ² , Nourou S. Yorou ³ ,
4	Ravi Wijesundera ⁴ , Luis Villarreal Ruiz ⁵ , Aída M. Vasco-Palacios ⁶ , Pham Quang Thu ⁷ , Ave
5	Suija ² , Matthew E. Smith ⁸ , Cathy Sharp ⁹ , Erki Saluveer ² , Alessandro Saitta ¹⁰ , Miguel Rosas ¹¹ ,
6	Taavi Riit ² , David Ratkowsky ¹² , Karin Pritsch ¹³ , Kadri Põldmaa ² , Meike Piepenbring ¹¹ ,
7	Cherdchai Phosri ¹⁴ , Marko Peterson ² , Kaarin Parts ² , Kadri Pärtel ² , Eveli Otsing ² , Eduardo
8	Nouhra ¹⁵ , André L. Njouonkou ¹⁶ , R. Henrik Nilsson ¹⁷ , Luis N. Morgado ¹⁸ , Jordan Mayor ¹⁹ ,
9	Tom W. May ²⁰ , Luiza Majuakim ²¹ , D. Jean Lodge ²² , Su See Lee ²³ , Karl-Henrik Larsson ²⁴ , Petr
10	Kohout ² , Kentaro Hosaka ²⁵ , Indrek Hiiesalu ² , Terry W. Henkel ²⁶ , Helery Harend ² , Liang-dong
11	Guo ²⁷ , Alina Greslebin ²⁸ , Gwen Grelet ²⁹ , Jozsef Geml ¹⁸ , Genevieve Gates ¹² , William
12	Dunstan ³⁰ , Chris Dunk ¹⁹ , Rein Drenkhan ³¹ , John Dearnaley ³² , André De Kesel ³³ , Tan Dang ⁷ ,
13	Xin Chen ³⁴ , Franz Buegger ¹³ , Francis Q. Brearley ³⁵ , Gregory Bonito ²⁰ , Sten Anslan ² , Sandra
14	Abell ³⁶ , Kessy Abarenkov ²
15	
16	¹ Natural History Museum, University of Tartu, Tartu, Estonia.

- ²Institute of Ecology and Earth Sciences, University of Tartu, Tartu, Estonia.
- 18 ³Faculté d'Agronomie, Université de Parakou, Parakou, Benin.
- ⁴Department of Plant Sciences, University of Colombo, Colombo 3, Sri Lanka.
- 20 ⁵Postgrado en Recursos Genéticos y Productividad-Genética, LARGEMBIO, Colegio de
- 21 Postgraduados-LPI 6, México City, Mexico.
- ⁶The Fungal Biodiversity Centre, CBS-KNAW, Utrecht, The Netherlands.
- 23 ⁷Vietnamese Academy of Forest Sciences, Hanoi, Vietnam.

- ⁸Department of Plant Pathology, University of Florida, Gainesville, Florida, USA.
- ⁹Natural History Museum, Bulawayo, Zimbabwe.
- ¹⁰Department of Agricultural and Forest Sciences, Università di Palermo, Palermo, Italy.
- 27 ¹¹Department of Mycology, Goethe University Frankfurt, Frankfurt am Main, Germany.
- 28 ¹²Tasmanian Institute of Agriculture, Hobart, Tasmania, Australia.
- 29 ¹³Institute of Soil Ecology, Helmholtz Zentrum München, Neuherberg, Germany.
- 30 ¹⁴Department of Biology, Nakhon Phanom University, Nakhon Phanom, Thailand.
- 31 ¹⁵Instituto Multidisciplinario de Biología Vegetal, Córdoba, Argentina.
- 32 ¹⁶Department of Biological Sciences, University of Bamenda, Bambili, Cameroon.
- 33 ¹⁷Department of Biological and Environmental Sciences, University of Gothenburg, Göteborg,

34 Sweden.

- 35 ¹⁸Naturalis Biodiversity Center, Leiden, The Netherlands.
- ¹⁹Department of Forest Ecology and Management, Swedish University of Agricultural
- 37 Sciences, Umeå, Sweden.
- 38 ²⁰Royal Botanic Gardens Melbourne, Melbourne, Victoria, Australia.
- 39 ²¹Institute for Tropical Biology and Conservation, University Malaysia Sabah, Sabah,

40 Malaysia.

- 41 ²²Center for Forest Mycology Research, USDA-Forest Service, Luquillo, Puerto Rico.
- 42 ²³Forest Research Institute Malaysia, Kepong, Selangor, Malaysia.
- 43 ²⁴Natural History Museum, University of Oslo, Oslo, Norway.
- 44 ²⁵Department of Botany, National Museum of Nature and Science, Tsukuba, Japan.
- 45 ²⁶Department of Biological Sciences, Humboldt State University, Arcata, California, USA.

- 46 ²⁷State Key Laboratory of Mycology, Institute of Microbiology, Chinese Academy of Sciences,
- 47 Beijing, China.
- 48 ²⁸CONICET Facultad de Cs. Naturales, Universidad Nacional de la Patagonia SJB, Esquel,
- 49 Chubut, Argentina.
- ²⁹Ecosystems and Global Change team, Landcare Research, Lincoln, New Zealand.
- ³⁰School of Veterinary & Life Sciences, Murdoch University, Western Australia, Australia.
- ³¹Institute of Forestry and Rural Engineering, Estonian University of Life Sciences, Tartu,
- 53 Estonia.
- ³²Faculty of Health, Engineering and Sciences, University of Southern Queensland,
- 55 Toowoomba, Queensland, Australia.
- ³³Botanic Garden Meise, Meise, Belgium.
- ³⁴College of Life Sciences, Zhejiag University, Hangzhou 310058, China.
- ³⁵School of Science and the Environment, Manchester Metropolitan University, Manchester,
- 59 United Kingdom.
- 60 ³⁶School of Marine and Tropical Biology, James Cook University, Cairns, Queensland,
- 61 Australia.
- 62
- 63 †Equal contribution
- 64 *Corresponding author. E-mail: leho.tedersoo@ut.ee

66 Abstract

68	Fungi play major roles in ecosystem processes, but the determinants of fungal diversity and
69	biogeographic patterns remain poorly understood. By using DNA metabarcoding data from
70	hundreds of globally distributed soil samples, we demonstrate that fungal richness is decoupled
71	from plant diversity. The plant-to-fungus richness ratio declines exponentially towards the
72	poles, indicating strong biases in previous fungal diversity estimates. Climatic factors,
73	followed by edaphic and spatial variables, constitute the best predictors of fungal richness and
74	community composition at the global scale. Fungi follow general biogeographic patterns
75	related to latitudinal diversity gradients but with several notable exceptions. These findings
76	significantly advance our understanding of fungal diversity patterns at the global scale and
77	permit integration of fungi into a general macro-ecological framework.
78	
79	
80	One-sentence summary
81	
82	A massive, global-scale metagenomic study detects hotspots of fungal diversity and
83	macroecological patterns, and indicates that plant and fungal diversity are uncoupled.
84	
85	

INTRODUCTION: The kingdom Fungi is one of the most diverse groups of organisms on
Earth and they are integral ecosystem agents that govern soil carbon cycling, plant nutrition,
and pathology. Fungi are widely distributed in all terrestrial ecosystems, but the distribution of
species, phyla, and functional groups has been poorly documented. Based on 365 global soil
samples from natural ecosystems, we determined the main drivers and biogeographic patterns
of fungal diversity and community composition.

92 RATIONALE: We identified soil-inhabiting fungi using 454 pyrosequencing and comparison 93 against taxonomically and functionally annotated sequence databases. Multiple regression 94 models were used to disentangle the roles of climatic, spatial, edaphic, and floristic parameters 95 on fungal diversity and community composition. Structural equation models were used to 96 determine the direct and indirect effects of climate on fungal diversity, soil chemistry and 97 vegetation. We also examined if fungal biogeographic patterns matched paradigms derived 98 from plants and animals — namely, that species' latitudinal ranges increase towards the poles 99 (Rapoport's rule) and diversity increases towards the equator. Finally, we sought group-100 specific global biogeographic links among major biogeographic regions and biomes using a 101 network approach and area-based clustering.

RESULTS: Metabarcoding analysis of global soils revealed fungal richness estimates
approaching the number of species recorded to date. Distance from equator and mean annual
precipitation had the strongest effects on richness of fungi including most fungal taxonomic
and functional groups. Diversity of most fungal groups peaked in tropical ecosystems, but
ectomycorrhizal fungi and several fungal classes were most diverse in temperate or boreal
ecosystems and many fungal groups exhibited distinct preferences for specific edaphic
conditions (e.g. pH, calcium, phosphorus). Consistent with Rapoport's rule, the geographic

range of fungal taxa increased toward the poles. Fungal endemicity was particularly strong intropical regions, but multiple fungal taxa had cosmopolitan distribution.

111 CONCLUSIONS: Climatic factors, followed by edaphic and spatial patterning, are the best112 predictors of soil fungal richness and community composition at the global scale. Richness of

all fungi and functional groups is causally unrelated to plant diversity with the exception of

114 ectomycorrhizal root symbionts, suggesting that plant-soil feedbacks do not influence the

115 diversity of soil fungi at the global scale. The plant-to-fungi richness ratio declined

116 exponentially towards the poles, indicating that current predictions assuming globally constant

117 ratios overestimate fungal richness by 1.5-2.5-fold. Fungi follow similar biogeographic

118 patterns as plants and animals with the exception of several major taxonomic and functional

119 groups that run counter to overall patterns. Strong biogeographic links among distant

120 continents reflect relatively efficient long-distance dispersal compared with macro-organisms.

121

126

122 Figure caption

123 Direct and indirect effects of climatic and edaphic variables on plant and fungal richness.

124 Line thickness corresponds to relative path coefficients. Dashed lines indicate negative

125 relationships. Abbreviations: MAP, mean annual precipitation; Fire, time since last fire.



127 Introduction

128

129 Fungi are eukaryotic microorganisms that play fundamental ecological roles as decomposers, 130 mutualists, or pathogens of plants and animals; they drive carbon cycling in forest soils, 131 mediate mineral nutrition of plants, and alleviate carbon limitations of other soil organisms. 132 Fungi comprise some 100,000 described species (accounting for synonyms), but the actual 133 extent of global fungal diversity is estimated at 0.8 to 5.1 million species (1). 134 Globally, the biomass and relative proportions of microbial groups, including fungi, co-135 vary with the concentration of growth-limiting nutrients in soils and plant tissues. Such 136 patterns suggest that the distribution of microbes reflects latitudinal variation in ecosystem 137 nutrient dynamics (2-4). Richness of nearly all terrestrial and marine macro-organisms is 138 negatively related to increasing latitude (5) — a pattern attributed to the combined effects of 139 climate, niche conservatism, and rates of evolutionary radiation and extinction (6). Although 140 morphological species of unicellular microbes are usually cosmopolitan (7), there is growing 141 evidence that the distribution of micro-organisms is shaped by macro-ecological and 142 community assembly processes (8). Only a few of these biogeographic processes have been 143 demonstrated for fungi at the local scale (9). Despite their enormous diversity and importance 144 in ecosystem function, little is known about general patterns of fungal diversity or functional 145 roles over large geographic scales. Here we use a global dataset to disentangle the roles of 146 climatic, edaphic, floristic, and spatial variables governing global-scale patterns of soil fungal 147 diversity. We also address macro-ecological phenomena and show that fungi largely exhibit 148 strong biogeographic patterns that appear to be driven by dispersal limitation and climate.

150 Materials and Methods

151

152 Sample preparation

153 We collected 40 soil cores from natural communities in each of 365 sites across the world using a uniform sampling protocol (Fig. 1A; Data S1). Most plots (2500 m²) were circular, but 154 155 in steep mountain regions and densely forested areas, some plots were oblong. We randomly 156 selected twenty trees located at least 8 m apart. In two opposite directions, 1-1.5 m from each 157 tree trunk, loose debris was removed from the forest floor. PVC tubes (5 cm diam.) were 158 hammered into the soil down to 5 cm depth. These soil cores almost always included fine roots 159 and comprised both the organic layer and top mineral soil. Although deep soil may contain 160 some unique organisms adapted to anoxic conditions or low nutrient levels, our sampling was 161 limited to topsoil for the following reasons. First, in the vast majority of soil types, >50% of 162 microbial biomass and biological activity occur in the topmost organic soil layer. Second, 163 deeper sampling was impossible in shallow, rocky soils or those with high clay concentrations 164 and hardpans. Third, differences among soil horizons may be masked by other variables across 165 large geographic scales (10). The 40 soil cores taken in each site were pooled, coarse roots and 166 stones removed, and a subset of the soil was air-dried at <35 °C. Dried soil was stored in zip-167 lock plastic bags with silica gel to minimize humidity during transit. In the laboratory, dried 168 soil was ground into fine powder using bead beating.

DNA was extracted from 2.0 g of soil using the PowerMax Soil DNA Isolation kit
(MoBio, Carlsbad, CA USA) following manufacturer's instructions. PCR was performed using
a mixture of six forward primers (in equimolar concentration) analogous to ITS3 and a

172 degenerate reverse primer analogous to ITS4 (hereafter referred to as ITS4ngs). Forward and 173 reverse primers were shortened and modified to completely match >99.5% of all fungi (except 174 ca. 60% of Tulasnellaceae that exhibit highly divergent 5.8S rDNA and Microsporidia that 175 exhibit re-arrangements in ribosomal DNA; Table S1). The ITS4ngs primer was tagged with 176 one of 110 identifiers (MIDs, 10-12 bases) that were modified from those recommended by 177 Roche to differ by >3 bases, start only with adenosine, and consist of between 30-70% 178 adenosine and thymidine in order to optimize the adapter ligation step. The PCR cocktail 179 consisted of 0.6 µl DNA extract, 0.5 µl each of the primers (20 pmol), 5 µl 5xHOT FIREPol 180 Blend Master Mix (Solis Biodyne, Tartu, Estonia), and 13.4 µl double-distilled water. PCR 181 was carried out in four replicates using the following thermocycling conditions: an initial 15 182 min at 95 °C, followed by 30 cycles at 95 °C for 30 s, 55 °C for 30 s, 72 °C for 1 min, and a 183 final cycle of 10 min at 72 °C. PCR products were pooled and their relative quantity was 184 estimated by running 5 µl amplicon DNA on 1% agarose gel for 15 min. DNA samples vielding no visible band were re-amplified using 35 cycles in an effort to obtain sufficient PCR 185 186 product, whereas samples with a very strong band were re-amplified with only 25 cycles. It is 187 important to use as few cycles as possible to minimize chimera formation and to be able to 188 interpret sequence abundance in a semiquantitative manner (11). We used negative (for DNA 189 extraction and PCR) and positive controls throughout the experiment. Amplicons were purified 190 with Exonuclease I and FastAP thermosensitive alkaline phosphatase enzymes (Thermo 191 Scientific, Pittsburgh, PA USA). Purified amplicons were subjected to quantity normalization 192 with a SequalPrep Normalization Plate Kit (Invitrogen, Carlsbad, CA, USA) following 193 manufacturer's instructions. Normalized amplicons were divided into five pools that were subjected to 454 adaptor ligation, emulsion PCR, and 454 pyrosequencing using the GS-FLX+ 194

technology and Titanium chemistry as implemented by Beckman Coulter Genomics (Danvers,MA, USA).

197

198 Bioinformatics

199 Pyrosequencing on five half-plates resulted in 2,512,068 reads with a median length of 409

200 bases. The sequences were re-assigned to samples in mothur 1.32.2 (www.mothur.org) based

201 on the barcodes and then trimmed (parameters: minlength=300; maxambigs=1;

202 maxhomop=12; qwindowaverage=35; qwindowsize=50; bdiffs=1) to exclude short and low-

203 quality sequences, resulting in 2,231,188 high quality sequences. We used ITSx 1.0.7

204 (http://microbiology.se/software/itsx) to remove the flanking 5.8S and 28S rRNA genes for

205 optimal resolution of ITS2 clustering and removal of compromised and non-target sequences.

As a filter to remove most of the partial sequences we retained only sequences >99 bp in

207 length. Chimera control was exercised through UCHIME 4.2 (www.drive5.com/uchime/).

After these filtering steps, 1,397,679 sequences were retained and further clustered at 90.0%

and 95.0-99.0% sequence similarity thresholds (12) as implemented in CD-Hit 4.6.1 (www.cd-

210 hit.org). Clustering revealed 37,387, 59,556, 66,785, 77,448, 94,255, and 157,956 taxa based

on 90.0%, 95.0%, 96.0%, 97.0%, 98.0%, and 99.0% sequence similarity thresholds,

212 respectively. The longest sequence of each Operational Taxonomic Unit (OTU), based on

clustering at 98.0% sequence similarity, was selected as the representative for BLASTn

searches (word size=7; penalties: gap=-1; gap extension=-2; match=1) against the International

215 Nucleotide Sequence Databases Collaboration (INSDC: www.insdc.org) and UNITE

216 (unite.ut.ee) databases. In addition, we ran BLASTn searches against established reference

217 sequences of all fungi in 99.0% similarity clusters that include third-party taxonomic and

218	metadata updates (12) as implemented in the PlutoF workbench (13) . For each query, we
219	considered the 10 best-matching references to annotate our global sequences as accurately as
220	possible. If no reliable taxon name was available, we ran manual BLASTn searches against
221	INSDC with 500 best matching sequences as output. We typically relied on 90%, 85%, 80%,
222	and 75% sequence identity as a criterion for assigning OTUs with names of a genus, family,
223	order, or class, respectively. Sequence identity levels were raised in subsets of
224	Sordariomycetes, Leotiomycetes, and Eurotiomycetes, because these taxa contain multiple
225	genera and families that have unusually conserved ITS sequences. As a rule, we considered e-
226	values of BLASTn search results $\leq e^{-50}$ reliable to assign sequences to the fungal kingdom,
227	whereas those $\geq e^{-20}$ were considered 'unknown'. E-values between e^{-20} and e^{-50} were manually
228	checked against the 10 best matches for accurate assignment. We followed INSDC for higher-
229	level taxonomy of eukaryotes (14) and the Index Fungorum (www.indexfungorum.org) for
230	species through class-level taxonomy of fungi. Our group of taxonomic experts assigned each
231	fungal genus, family, or order to functional categories (Data S2). If different functional
232	categories were present within a specific genus, we chose the dominant group (>75% of
233	species assigned to a specific category) or considered its ecology unknown (<75% of species
234	assignable to a single category). All Glomeromycota were considered to be arbuscular
235	mycorrhizal (AM). Taxa were considered to be ectomycorrhizal (EcM) if they best matched
236	any sequences of known EcM lineages (15) and exhibited sequence length / BLASTn scores
237	above lineage-specific thresholds. For several taxonomic groups, we constructed phylogenetic
238	trees to assess the performance of clustering, sequence quality of singletons, accuracy of OTU
239	separation, and taxonomic assignments (Fig. S1). In the course of this project, we provided

10,232 third-party taxonomic re-annotations to INSDC sequences to improve subsequent
identification of fungal sequences and made these available through the UNITE database.

243 Statistical analyses

244 Estimates of the mean annual temperature (MAT), mean annual precipitation (MAP), soil 245 moisture, and soil carbon at 30 arc second resolution were obtained from the WorldClim 246 database (www.worldclim.org). Estimates of potential evapotranspiration (PET) and net 247 primary productivity (NPP) at 30 arc minute resolution were obtained from the Atlas of the 248 Biosphere (www.sage.wisc.edu/atlas/maps.php). Variation coefficients for MAT and MAP 249 were computed based on the average monthly values to represent seasonality of temperature 250 and precipitation. We also calculated the difference of MAP to PET to evaluate the effect of 251 rainfall surplus or deficit. Based on vegetation type and geographical distribution, sites were 252 categorized into biogeographic regions and biomes following the classification of the World 253 Wildlife Foundation (http://worldwildlife.org) with a few exceptions: i) temperate deciduous 254 forests in the Northern and Southern hemispheres were treated separately; ii) tropical montane 255 forests (>1500 m elevation) were separated from the tropical lowland moist forests; and, iii) 256 grasslands and shrublands of all geographic origins were pooled. At each site, we also 257 determined the age of vegetation, time since the last fire, and EcM plant species along with 258 their relative contribution to stand basal area. EcM plants are usually conspicuous trees or 259 prominent shrubs that are relatively easy to identify and their mycorrhizal status is verifiable in 260 the field using root excavation and microscopy. Complete lists of tree species were available 261 for <10% of the sites, so we did not directly include plant community composition parameters 262 in our analyses (but see below).

Concentrations of N, C, ¹³C/¹²C, and ¹⁵N/¹⁴N were determined from 1-20 mg of soil
using GC-combustion coupled to isotope-ratio mass spectrometry (*16*). Concentrations of soil
calcium, potassium, magnesium, and phosphorus were determined as in Tedersoo et al. (*16*).
Soil pH was measured in 1 N KCl solution.

267 For analyses of fungal richness, we calculated residuals of OUT richness in relation to 268 the square root of the number of obtained sequences to account for differences in sequencing 269 depth. This method outperformed the commonly used rarefaction to the lowest number of 270 sequences method, which removes most of the data (17). We also calculated the richness of 271 major class-level taxonomic and functional groups (comprising >100 OTUs). We excluded 272 outlying samples dominated by a few OTUs of molds, which are indicative of poor sample 273 preservation (relative abundance of sequences belonging to Trichocomaceae >5%, 274 Mortierellaceae $\geq 20\%$, or Mucoraceae $\geq 20\%$, that exceeded three times the mean + standard 275 deviation). Although these samples were fairly homogeneously distributed across the world,

they had conspicuously lower fungal richness. We also excluded samples that yielded less than1200 sequences per sample.

278 To determine the relationship between plant and fungal richness, we relied on cokriging values from the global vascular plant species richness dataset (18), which covered 279 280 96.7% of our sites. These scale-free values of plant richness were then regressed with residuals 281 from the best fit models for fungal richness and fungal functional groups. We further calculated 282 the ratio of relative plant richness to fungal richness and fitted this ratio with latitude using 283 polynomial functions to test the assumed uniformity of plant-to-fungal richness ratios at the global scale (1, 19, 20). To account for potential latitudinal biases in plant-to-fungal diversity 284 285 estimates, we took into account the non-uniform distribution of land surfaces by calculating an

Inverse Distance Weighting (IDW) spatial interpolation of standardized ratios of plant-toresidual fungal diversity using the *gstat* package in R (*21*). We then used IDW to interpolate
total fungal diversity beyond sampling sites, by accounting for MAP as based on the bestfitting multiple regression model.

290 Distance from the equator, altitude, age of vegetation, time since last fire, climatic 291 variables, and concentrations of nutrients were log-transformed prior to analyses to improve 292 the distribution of residuals and reduce non-linearity. To account for potential autocorrelation 293 effects, we calculated spatial eigenvectors using SAM ver. 4 (22). To determine the best 294 predictors of global fungal diversity, we included edaphic, climatic, floristic, and spatial 295 variables in multiple regression models. Due to the large number of predictors, we pre-selected 296 16 candidate predictors that were revealed by exploratory multiple linear and polynomial 297 regression analyses, based on coefficients of determination and forward selection criteria. The 298 most parsimonious models were determined based on the corrected Akaike information criterion (AICc), which penalizes over-fitting. Finally, components of the best models were 299 300 forward-selected to determine their relative importance as implemented in the *packfor* package 301 in R.

To test the direct effects of climatic variables on richness of fungi and their functional groups, and indirect climatic effects (via soil nutrients and vegetation), we used Structural Equation Modeling (SEM) in Amos ver. 22 (SPSS Software, Chicago, IL, USA). Model fits were explored based on both chi-square test and Root Mean Square Error of Approximation (RMSEA). First, we included all potentially important variables (inferred from both the multiple regression models and correlations for individual response variables to construct separate SEM models. We tested all direct and indirect relations between exogenous and

endogenous variables including their error terms. Then, we used backward elimination to
remove non-significant links to maximise whole model fit. Finally, we combined the obtained
SEM models in a unified path model, following the same elimination procedure.

In addition to full models, we specifically tested the relationships between OTU richness and distance from the equator and soil pH, because these or closely related variables were usually among the most important predictors. For these analyses, we calculated residuals of richness that accounted for other significant variables of the best models. To address nonlinear relationships, we fitted up to fifth order polynomial functions and selected best fit models based on AICc values.

318 The relative effects of climatic, edaphic, spatial, and floristic variables on the total 319 fungal community composition and on particular functional groups were determined using 320 Hellinger dissimilarity (calculated if >90% sites were represented by >1 shared OTUs), 321 exclusion of all OTUs that occurred once, and a multi-stage model selection procedure as 322 implemented in the DISTLM function of Permanova+ (www.primer-e.com/permanova.htm). 323 Considering computational requirements, 15 candidate variables were pre-selected based on 324 unifactorial (marginal test based on largest F_{pseudo} values) and multifactorial (forward selection) 325 models. Spatial eigenvectors were not included in these analyses, because they were typically 326 of minor importance in variation partitioning analyses (see below), and to avoid making the 327 models computationally prohibitive. Optimal models were selected based on the AICc. To obtain coefficients of determination (cumulative $R^2_{adjusted}$) and statistics (F_{pseudo} and P-values) 328 329 for each variable, components of the best models were forward selected. In parallel, we 330 prepared Global Nonmetric Multidimensional Scaling (GNMDS) graphs using the same 331 options. Significant variables were fitted into the GNMDS ordination space using the *envfit*

function in the *vegan* package of R. We also grouped all climatic, edaphic, spatial, and floristic variables into a variation partitioning analysis by integrating procedures in the *vegan* and *packfor* packages of R. Besides group effects, variation partitioning estimates the proportion of shared variation among these groups of predictors.

336 For global biogeographic analyses, we excluded OTUs from the order Hypocreales and 337 family Trichocomaceae (both Ascomycota), because the ITS region provides insufficient 338 taxonomic resolution and known biological species are grouped together within the same OTU 339 (23). We tested the differences among fungal taxonomic and functional groups for the 340 occurrence frequency (number of sites detected) and latitudinal range of OTUs using a non-341 parametric Kruskal-Wallis test and Bonferroni-adjusted multiple comparisons among mean 342 ranks. To test the validity of Rapoport's rule in soil fungi, we calculated the average latitudinal 343 range of OTUs for each site (24). The average latitudinal range was regressed with the latitude 344 of study sites by polynomial model selection based on the AICc criterion. This analysis was 345 run with and without OTUs only detected at a single site (range=0). Because the results were 346 qualitatively similar, we report results including all OTUs. To construct biogeographic 347 relationships among major regions and biomes, we generated cross-region and cross-biome 348 networks based on the number of shared OTUs. We excluded occurrences represented by a 349 single sequence per site. Ward clustering of biogeographic regions and biomes were 350 constructed using the Morisita-Horn index of similarity, which is insensitive to differences in 351 samples size, by use of the *pvclust* package of R. In this procedure, P-values are inferred for 352 non-terminal branches based on multiscale bootstrap resampling with 1,000 replicates. 353

354 Results and Discussion

355

356 Taxonomic and functional diversity

357 Pyrosequencing analysis of global soil samples revealed 1,019,514 quality-filtered sequences 358 that were separated into 94,255 species-level OTUs (see supplementary information). 359 Altogether 963,458 (94.5%) sequences and 80,486 (85.4%) OTUs were classified as Fungi. 360 Most other taxa belonged to animals (Metazoa, 3.3%), plants (Viridiplantae, 3.1%), alveolates 361 (Alveolata, 2.8%), and amoebae (mostly Rhizaria, 1.3%). Kingdom-level assignment of 3.8% 362 OTUs remained elusive. The fungal subset included 35,923 (44.6%) OTUs that were 363 represented by a single sequence; these were removed from further analyses to avoid 364 overestimating richness based on these potentially erroneous sequences (25). The remaining 365 44,563 non-singleton fungal OTUs in our data set numerically correspond to approximately 366 half of the described fungal species on Earth (1). For comparison, there are currently 52,481 367 OTUs based on 98.0% similarity clustering of all fungal ITS sequences in publicly available 368 databases (12). Global soil sampling revealed representatives of all major phyla and classes of 369 Fungi. Of fungal taxa, Basidiomycota (55.7%), Ascomycota (31.3%), Mortierellomycotina 370 (6.3%) and Mucoromycotina (4.4%) encompassed the largest proportion of sequences (Fig. 2), 371 whereas the most OTU-rich phyla were the Ascomycota (48.7%), Basidiomycota (41.8%), 372 Chytridiomycota (2.3%), and Cryptomycota (syn. Rozellida; 2.1%) (Fig. S2; Data S1). Except 373 for the recently described phylum Cryptomycota (26), the relative proportions of major phyla 374 correspond to the proportional distribution of taxa described and sequenced to date (12, 12)375 www.indexfungorum.org). Below the phylum level, approximately 6% of all fungal OTUs 376 could not be assigned to any known class of fungi. Further clustering of unidentified fungal

377 sequences at 70% sequence similarity revealed 14 distinct taxonomic groups comprising >7
378 OTUs, suggesting that there are several deeply divergent class-level fungal lineages that have
379 not yet been described or previously sequenced.

380 Our classification revealed that 10,801 (24.2%) fungal OTUs exhibited >98% sequence 381 similarity, and 33.8% exhibited >97% similarity, to pre-existing ITS sequences in public 382 databases. This is consistent with Taylor et al. (19), reporting 48% of OTUs amplified from 383 Alaskan soils with >97% similarity to any database sequences. In our study, only 4353 fungal 384 OTUs (9.8%) were matched to sequences from herbarium specimens or fully described culture 385 collections at >98.0% sequence similarity. Although many type collections are yet to be 386 sequenced, the paucity of matches to database entries indicates that a majority of soil-387 inhabiting fungal taxa remain undescribed (19-20). These results highlight the current lack of data from understudied tropical and subtropical ecosystems. The phenomenon of high cryptic 388 389 diversity and low success in naming OTUs at the genus or species level have been found in 390 other groups of soil microbes and invertebrates, emphasizing our poor overall knowledge of 391 global soil biodiversity (27-28).

The main fungal phylogenetic and functional groups were present in all ecosystems, but their relative proportions varied several-fold across biomes (Figs. 2, S2-S4). The ratio of Ascomycota to Basidiomycota OTUs was highest in grasslands and shrublands (1.86) and tropical dry forests (1.64) but lowest in the temperate deciduous forests (0.88). Chytridiomycota, Cryptomycota, and Glomeromycota were relatively more diverse in the

grasslands and shrublands, accounting for 4.6%, 3.6%, and 1.4% of OTU richness,

398 respectively. The relative OTU richness of Mortierellomycotina and Mucoromycotina

397

399 (including most fast-growing molds but also some plant symbionts) peaked in the tundra biome

400 (4.8% and 2.7%, respectively), but their abundance was lowest in tropical dry forests (1.0%401 and 0.6%, respectively). Archaeorhizomycetes, a recently described class of Ascomycetes from 402 a boreal forest (29), was most diverse in tropical moist and montane forests, particularly in 403 northern South America and New Guinea. 404 Among all fungal taxa, OTUs assigned to saprotrophs, EcM mutualists, and plant 405 pathogens comprised 19,540 (43.8%), 10,334 (23.2%), and 1770 (4.0%), respectively (Fig. 406 S4). Other trophic categories were contained <1% of remaining OTUs. EcM fungi contributed 407 34.1% of all taxa in the northern temperate deciduous forests, but accounted for a relatively 408 low proportion (11.9%) in grasslands and shrublands, reflecting the paucity of host plants in 409 these ecosystems. Similarly, the proportion of EcM fungal taxa was lowest in northern South 410 America (8.0%), where AM trees often dominate. Plant pathogens were relatively more 411 abundant and diverse in lowland tropical moist (6.2%) and dry (6.3%) forests.

412

413 Predictors of global richness

414 Structural equation models revealed that climate has both a strong direct effect on plant and 415 fungal richness and functional groups, but it also indirectly affects these metrics by altering 416 edaphic conditions (Main text; Fig. S5). Both SEM and regression models suggest that the best 417 predictors of diversity differed among phylogenetic and functional groups of fungi. Positive 418 effects of mean annual precipitation (MAP) and soil Ca concentration were the strongest 419 predictors of total fungal diversity, explaining 7.2% and 8.9% of residual richness, respectively 420 (Table S2). Richness of EcM fungi responded positively to the relative proportion and species 421 richness of EcM plants (explaining 18.3% and 8.5% of variance, respectively), as well as soil 422 pH (13.0%). EcM host species richness (5.9%) and soil pH (20.4%) remained the strongest

423	predictors in the best model for sites with EcM vegetation accounting for $>60\%$ of basal area, a
424	critical point above which the proportion of EcM plants had no further effect on EcM fungal
425	richness. MAP had a strong positive effect (14.8%) on richness of saprotrophs. Diversity of
426	plant pathogens declined with increasing distance from the equator (17.8%) and soil C/N ratio
427	(11.6%). Animal parasites responded positively to MAP (20.3%), whereas monthly variation of
428	precipitation (MAP CV) had a negative impact on richness of mycoparasites (fungus-parasitic
429	fungi; 8.2%). Richness of the AM Glomeromycota was negatively related to the age of
430	vegetation (7.3%) but positively related to potential evapotranspiration (PET, 3.5%) and soil
431	pH (4.3%). Of the major taxonomic groups, the richness of Ascomycota in general (18.5%)
432	and that of Archaeorhizomycetes (21.7%) were negatively related to distance from the equator
433	in best-fit models. Climatic variables were the best predictors for richness of
434	Mortierellomycotina (MAT: negative effect, 26.1%) and the ascomycete classes
435	Dothideomycetes (MAT: positive effect, 20.9%), Lecanoromycetes (MAT: negative effect,
436	26.7%), Leotiomycetes (MAT: negative effect, 30.1%), Orbiliomycetes (MAT: positive effect,
437	12.8%), and Sordariomycetes (MAP: positive effect, 33.4%). The richness of Chytridiomycota
438	and the ascomycete class Pezizomycetes was best explained by a positive response to soil pH
439	(8.6% and 40.5%, respectively). Concentration of soil nutrients or their ratio to other nutrients
440	were the strongest predictors for OTU richness of Cryptomycota (N concentration: positive
441	effect, 10.1%), Geoglossomycetes (N/P ratio: positive effect, 3.7%), Mucoromycotina (C/N
442	ratio: positive effect, 19.0%), and Wallemiomycetes (P concentration: negative effect, 14.9%).
443	The richness of Basidiomycota and its class Agaricomycetes were best explained by a positive
444	response to soil Ca concentration (13.5% and 12.8%, respectively).

445	Although geographical distance per se had negligible effects on richness (Moran's
446	I=0.267), spatial predictors were included in the best richness models of nearly all functional
447	and phylogenetic groups (except Glomeromycota), indicating regional- or continental-scale
448	differences in OTU richness (Fig. 1B). Compared to other tropical regions, richness of fungi
449	was conspicuously lower in Africa, independent of biome type. These results might reflect the
450	relatively lower MAP in much of Africa compared with other tropical continents.
451	Alternatively, lower fungal richness could be related to the disproportionately strong shifts in
452	biomes during the Pleistocene, which impoverished the African flora (18).
453	Among edaphic variables, soil pH and Ca concentration were typically the most
454	important predictors of fungal OTU richness. These variables positively correlated with fungal
455	richness at the global scale (F _{1,335} =290.7; R _{Pearson} =0.682; P<0.001). The strong positive
456	influence of soil Ca concentration on richness of fungi, in particular Basidiomycota, is
457	congruent with a similar positive relationship found for Ca and EcM fungal richness associated
458	with Northern Hemisphere Alnus spp. (30). Exchangeable Ca is important for many
459	physiological processes in plants and microorganisms and it influences the turnover rate of soil
460	organic matter (31). In soil geochemical processes, pH and Ca concentration affect each other
461	and thus may have both direct and indirect effects on soil biota. Fungal functional groups were
462	differentially affected by pH. Richness of EcM fungi was greatest in slightly acidic to neutral
463	soils (Fig. S6), whereas saprotrophs, especially white rot decomposers, were more diverse in
464	moderately to strongly acidic soils. Richness of Pezizomycetes peaked distinctly in neutral
465	soils.
466	

467 Macroecological patterns

468	In general agreement with biogeographic patterns of plants, animals, and foliar endophytic
469	fungi (5,32), the overall richness of soil fungi increased towards the equator (Fig. 3A).
470	However, major functional and taxonomic groups showed dramatic departures from the
471	general latitudinal richness patterns (Figs. 3, S7). Namely, diversity of saprotrophic fungi,
472	parasites, and pathogens increased at low latitudes, whereas richness of EcM fungi peaked at
473	mid-latitudes, especially in temperate forests and Mediterranean biomes of the Northern
474	Hemisphere (40-60 °N; Fig. S8). By contrast, saprotrophic fungi had a broad richness peak
475	spanning from ca. 45 °S to 25 °N. Richness of Ascomycota, in particular that of
476	Archaeorhizomycetes, Dothideomycetes, Eurotiomycetes, Orbiliomycetes, and
477	Sordariomycetes, peaked in tropical ecosystems (Fig. S7). Conversely, the ascomycete classes
478	Lecanoromycetes and Leotiomycetes as well as Microbotryomycetes (basidiomycete yeasts),
479	Mortierellomycotina, and Mucoromycotina increased in diversity towards the poles, with no
480	noticeable decline in boreal forests and tundra biomes. Agaricomycetes, Pezizomycetes, and
481	Tremellomycetes exhibited distinct richness peaks at mid-latitudes. Richness of
482	Agaricomycetes was greater in the Northern Hemisphere, whereas that of Microbotryomycetes
483	Tremellomycetes, and Wallemiomycetes peaked in the Southern Hemisphere temperate
484	ecosystems (Fig. S8).
485	All of these phylogenetic groups originated >150 million years ago on the

486 supercontinent Pangaea (*33*) and have had sufficient time for long-distance dispersal. However, 487 our data suggest that particular regional biotic or abiotic conditions (e.g., soil pH and favorable 488 climatic conditions) have likely stimulated evolutionary radiations in certain geographic areas 489 and not in others. Adaptation to cold climate in younger fungal phyla has been suggested to 490 explain differential latitudinal preferences among fungal groups (*34*). However, our global

analysis provided no support for this hypothesis (Fig. S9). Instead, it revealed that ancient

492 lineages are relatively more common in non-wooded ecosystems.

493

494 Relation of plant and fungal richness

Plant and fungal richness were positively correlated (Fig. S10), but plant richness explained no residual richness of fungi based on the best regression model ($R^2_{adj} < 0.01$; P>0.05). These results and SEM path diagrams suggest that correlations between plant and fungal richness are best explained by their similar response to climatic and edaphic variables (i.e., covariance) rather than by direct effects of plants on fungi. However, when separating functional categories, trophic groups of fungi exhibited differential response to plant diversity and relative proportion of potential hosts.

502 Plant pathogens usually attack a phylogenetically limited set of host plants (35), 503 suggesting that that plant pathogens have at least partly co-evolved with their hosts and may 504 have radiated more intensively in the tropics where high plant diversification and richness 505 permit greater diversification. Strong phylogenetic signals in soil feedbacks, adaptive radiation, 506 and negative density dependence (the Janzen-Connell hypothesis) have probably contributed to 507 the pronounced richness of both plants and their pathogens at low latitudes (36, 37). However, 508 our analyses revealed no significant effects of plant richness per se on residual richness of 509 pathogens in soil. Similarly to pathogens, richness of AM fungi was unrelated to the proportion 510 of AM host trees or interpolated host richness, which may result from non-specific associations 511 with tree and understory species. Hence both AM and soil pathogen richness were unaffected 512 by plant richness. By contrast, host richness explained 6% of variation in EcM fungal richness, 513 indicating either niche differentiation of fungi in forests of mixed hosts or sampling effects

514 (i.e., forests with higher host diversity are more likely to include plant species that harbor high 515 fungal diversity). With a few notable exceptions, most studies have found low levels of host 516 preference or host specificity among EcM fungi (38). We found that relative EcM host density 517 had a strong influence on EcM fungal richness, suggesting that greater availability of 518 colonizable roots in soil provides more carbon for EcM fungi and thereby yields greater 519 species density and local-scale richness regardless of latitude. The peak of EcM fungal 520 taxonomic and phylogenetic richness in northern temperate biomes coincides with the 521 geographical distribution and dominance of Pinaceae, which is the oldest extant EcM plant 522 family (15, 39).

523 The ratio of plant-to-fungal richness decreased exponentially with increasing latitude, 524 because plant diversity dropped precipitously toward the poles relative to fungal diversity (Fig. 525 4). This finding calls into question present global fungal richness estimates. These estimates 526 assume similar spatial turnover of plant and fungal species and a constant plant-to-fungus ratio, 527 and have been formulated based mostly on data from temperate and boreal ecosystems (1, 19, 528 20). Yet local-scale beta diversity of both plants and fungi differ among temperate and tropical 529 sites (40, 41) and there are profound differences in plant species turnover depending on 530 propagule size (42). Natural distribution of very few vascular plant species encompass several 531 continents, but there are multiple fungal species with circumpolar or cosmopolitan distribution 532 (43, 44; see Biogeography section below). While we cannot directly compare plant and fungal 533 beta diversity, spatial turnover of plant species is inarguably greater (42). Based on the 534 function of fungi-to-plant richness ratio to latitude and latitudinal distribution of land, we 535 calculated that fungal richness is overestimated by 1.5- and 2.5-fold based on constant 536 temperate (45° latitude) and boreal (65° latitude) richness ratios, respectively.

Since richness estimates are calculated based on the frequency of the rarest species, the
reliability of singleton data call into question biologically meaningful extrapolations (11). In
metabarcoding studies such as ours, sequencing errors tend to give rise to singleton sequences,
and the number of rare artificial taxa grows rapidly with increasing sequencing depth (25).
Therefore, despite the size of our dataset, it cannot readily be used to produce reliable
taxonomic richness extrapolations.

543

544 Community ecology

545 Variation partitioning analysis revealed that climatic, edaphic, and floristic variables (and their 546 shared effects) are the strongest predictors for community composition of all fungi and most of 547 their functional groups (Fig. S11). However, the saprotroph community composition was most 548 strongly explained by purely spatial variables. More specifically, PET and soil pH explained 549 2.4% and 1.5%, respectively, of the variation in total fungal community composition (Table 550 S3; Fig. S12). PET contributed 3.8%, 2.8%, and 11.7% to community structure of saprotrophs, 551 plant pathogens, and yeasts, respectively. Distance from the equator (1.3%) and soil pH (0.7%)552 were the strongest predictors of EcM fungal community composition, whereas mean annual temperature (4.0%) was the strongest predictor for animal parasites, and distance from the 553 554 equator (3.5%) was the best predictor for mycoparasites (Table S3; Fig. S12).

These results indicate that both environmental and spatial predictors generally have a minor influence on species-level composition of fungi at the global scale. Nonetheless, the significant global-scale pH effect in several groups of fungi is consistent with the substantial influence of pH on the phylogenetic structure of soil fungal and bacterial communities in both local and continental scales (*27, 45*). The relatively stronger climatic and edaphic drivers of

richness at the class and phylum level suggest that phylogenetic niche conservatism in fungal
lineages is similar to cross-biome distribution patterns in vascular plants (*46*) and protists (*47*).

563 Global biogeography

564 Consistent with Rapoport's rule formulated for macro-organisms (*24*) and later applied to 565 marine bacteria (*48*), the mean latitudinal range of fungi strongly increased towards the poles 566 (Fig. S13). These results also suggest that a greater proportion of fungi are endemic within 567 tropical rather than extra-tropical ecosystems.

568 Major taxonomic and functional groups of fungi differed markedly in their distribution range (Figs. S14, S15). Animal parasites were more widely distributed compared with all other 569 570 groups, suggesting that there are many generalist OTUs with global distribution. Saprotrophs 571 and plant pathogens had broader distribution ranges than EcM and AM root symbionts. Taxa 572 belonging to Mortierellomycotina, Mucoromycotina, Tremellomycetes, and Wallemiomycetes 573 - groups that include a large proportion of saprotrophs and parasites that produce exceptionally 574 large quantities of aerially dispersed mitospores – were generally most widely distributed. 575 Besides the AM Glomeromycota, OTUs belonging to the ascomycete classes 576 Archaeorhizomycetes, Geoglossomycetes, and Orbiliomycetes were detected from the fewest 577 sites. 578 The northernmost biogeographic regions (Europe, West Asia, East Asia, and North

579 America) had the most similar fungal communities as revealed by shared fungal OTUs (Fig. 5).

580 Based on the Morisita-Horn similarity index, the northern and southern temperate regions

581 clustered together with marginally non-significant support (P=0.064; Fig. 6A). In spite of the

582 large geographical distance separating them, paleo- and neotropical biogeographic regions

583 clustered together (P=0.059). However, biogeographic clustering of regions deviated markedly 584 in certain functional groups of fungi (Fig. 6). For instance, EcM fungi in the southern 585 temperate and tropical regions had greater similarity compared with northern temperate 586 ecosystems (P=0.001). Among biomes, boreal forests, temperate coniferous forests, and 587 temperate deciduous forests shared the largest numbers of fungal OTUs (Fig. S16). Fungal 588 OTUs in temperate deciduous forests were highly similar to Mediterranean and tropical 589 montane forests, whereas fungal OTUs in tropical montane forests were linked to tropical 590 moist forests, which in turn exhibited substantial connections with tropical dry forests and 591 savannas. As a result, cluster analysis supported separation of tropical and non-tropical biomes 592 (Fig. 6B). Consistent with biogeographic region-level analysis, lowland tropical biomes, arctic 593 tundra and boreal forests biomes, and temperate biomes formed three well-supported clusters. 594 Tropical montane forests and grasslands and shrublands were clustered with temperate biomes 595 based on distribution of all fungi and most functional groups. However in EcM fungi, taxa 596 from southern temperate forests, tropical montane forests, and grass/shrublands clustered with 597 tropical lowland and Mediterranean biomes. A relatively large proportion of EcM fungal taxa 598 were shared across various biomes in Australia and New Guinea, which explains these 599 deviating patterns. In contrast, plant pathogens from tropical montane forests clustered with 600 tropical lowland biomes rather than with temperate biomes.

601 Our biogeographic analyses complement the community-level results suggesting that 602 both climate and biogeographic history shape macro-ecological patterns of fungi. Co-migration 603 with hosts over Pleistocene land bridges (e.g., Beringia, Wallacea, Panamanian) and long-604 distance dispersal by spores appear to have played important roles in shaping current fungal 605 distribution patterns (*30, 43*). The relative influence of climate and biotrophic associations with

606 host plants of varying extant distributions probably contribute to differences in the range and 607 biogeographic relationships among fungal functional groups (49). In addition, taxon-specific 608 constraints for dispersal, such as shape and size of propagules and sensitivity to UV light, may 609 differentially affect long-distance dispersal among taxa (7). For instance, Glomeromycota 610 OTUs, which form relatively large non-wind dispersed asexual spores, had the lowest average 611 geographical range. In general, region-based distribution patterns of fungi are somewhat 612 conflicting with clustering of plants and animals, where Holarctic lineages are deeply nested 613 within larger tropical groups (50). Consistent with macro-organisms, fungi from the Southern 614 Hemisphere temperate landmasses cluster together. Differences observed in macro-ecological 615 patterns among fungi, plants, and animals may originate from the relative strength of dispersal 616 limitation and phylogeographic history, but exaggeration by methodological differences among 617 studies cannot be discounted. The use of homogenous sampling and analytical methods, as 618 done in this study, are necessary to confidently compare macro-ecological patterns amongst 619 distinct life forms and to reliably test degrees of consistency among all kingdoms of life.

620

621 Conclusions and perspectives

622 Climatic variables explained the greatest proportion of richness and community composition in 623 fungal groups by exhibiting both direct and indirect effects through altered soil and floristic 624 variables. The strong driving climatic forces identified here open up concerns regarding the 625 impact of climate change on the spread of disease (*51*) and the functional consequences of 626 altered soil microorganism communities (*52*). The observed abrupt functional differences 627 between fungal communities in forested and treeless ecosystems, despite spatial juxtaposition, 628 suggests that plant life form and mycorrhizal associations determine soil biochemical processes

more than plant species *per se*. Loss of tree cover and shrub encroachment resulting from
drying and warming may thus have a marked impact on ecosystem functioning both aboveand belowground.

632 In addition to natural mechanisms, such as long-distance dispersal and migration over 633 past land bridges, global trade has enhanced the spread of some non-native soil organisms into 634 other ecosystems, where they sometimes become hazardous to native biota, economy, and 635 human health (53). Our results highlight how little insight we still have into natural microbial 636 distribution patterns, and this undermines our ability to appraise the actual role of humans in 637 shaping these biogeographic processes. Even larger-scale sampling campaigns are needed to 638 provide data for establishing natural distributions and building species distribution models 639 (52), which will enable us to predict the spread and habitat suitability of non-native 640 microorganisms.

642 References and Notes

- 643
- 644 1. M. Blackwell, Am. J. Bot. 98, 426 (2011).
- 645 2. N. Fierer *et al.*, *Ecol. Lett.* **12**, 1 (2009).
- 646 3. H. Serna-Chavez, N. Fierer, P. M. van Bodegom, *Glob. Ecol. Biogeogr.* 10, 1162 (2013).
- 647 4. X. Xu, P. Thornton, W. M. Post, *Glob. Ecol. Biogeogr.* 22, 737 (2013).
- 648 5. H. Hillebrand, Am. Nat. 163, 192 (2004).
- 649 6. G. G. Mittelbach *et al.*, *Ecol. Lett.* **10**, 315 (2007).
- 650 7. B. J. Finlay, *Science* **296**, 1061 (2002).
- 651 8. D. R. Nemergut et al., Microbiol. Mol. Biol. Rev. 77, 342 (2013).
- 652 9. K. G. Peay, M. I. Bidartondo, A. E. Arnold, New Phytol. 185, 878 (2010).
- 653 10. J. M. Talbot et al., Proc. Natl. Acad. Sci. USA, 111, 6341 (2014)
- 654 11. B. D. Lindahl *et al.*, *New Phytol.* **199**, 288 (2013).
- 655 12. U. Kõljalg *et al.*, *Mol. Ecol.* **22**, 5271 (2013).
- 656 13. K. Abarenkov *et al., Evol. Bioinform.* 6, 189 (2010).
- 657 14. S. M. Adl et al., J. Eukaryot. Microbiol., 59, 527 (2012).
- 658 15. L. Tedersoo, M. E. Smith, Fung. Biol. Rev. 27, 83 (2013).
- 659 16. L. Tedersoo *et al.*, New Phytol. **195**, 832 (2012).
- 660 17. I. Hiiesalu *et al.*, New Phytol. **203**, 233 (2014).
- 661 18. H. Kreft, W. Jetz, Proc. Natl. Acad. Sci. USA 104, 5925 (2007).
- 662 19. D. L. Taylor *et al.*, *Ecol. Monogr.* **84**, 3 (2014).
- 663 20. H. E. O'Brien *et al.*, *Appl. Environ. Microbiol.* **71**, 5544 (2005).

- 664 21. R Core Team, R: a language and environment for statistical computing. Vienna: R
- 665 Foundation for Statistical Computing (2014).
- 666 22. T. F. Rangel *et al.*, *Ecography* **33**, 46 (2010).
- 667 23. C. L. Schoch et al., Proc. Natl. Acad. Sci. USA 109, 6241 (2012).
- 668 24. G. C. Stevens, Am. Nat. 133, 240 (1989).
- 669 25. I. A. Dickie, New Phytol. 188, 916 (2010).
- 670 26. M. D. M. Jones et al., Nature 474, 200 (2011).
- 671 27. C. Lauber et al., Appl. Environ. Microbiol. 75, 5111 (2009).
- 672 28. M. S. Robeson et al., Proc. Natl. Acad. Sci. USA 108, 4406 (2011).
- 673 29. A. Rosling et al., Science 333, 876 (2011).
- 674 30. S. Põlme *et al.*, *New Phytol.* **198**, 1239 (2013).
- 675 31. P. B. Reich *et al.*, *Ecol. Lett.* **8**, 811 (2005).
- 676 32. A. E. Arnold, Fung. Biol. Rev. 21, 51 (2007).
- 677 33. M. L. Berbee, J. W. Taylor, Fung. Biol. Rev. 24, 1 (2010).
- 678 34. K. K. Treseder *et al.*, *Ecol. Lett.* 9, 1086 (2014).
- 679 35. G. S. Gilbert, C. O. Webb, Proc. Natl. Acad. Sci. USA 104, 4979 (2007).
- 680 36. X. Liu et al., Ecol. Lett. 15, 111 (2012).
- 681 37. R. Bagchi *et al.*, *Nature* **506**, 85 (2014).
- 682 38. M. Bahram *et al.*, *Fung. Ecol.* 7, 70 (2013).
- 683 39. L. Tedersoo *et al.*, *Mol. Ecol.* **21**, 4160 (2012).
- 684 40. M. Bahram *et al.*, J. Ecol. 101, 1335 (2013).
- 685 41. H. Qian et al., Glob. Ecol. Biogeogr. 22, 659 (2013).
- 686 42. H. Qian, *Glob. Ecol. Biogeogr.* **18**, 327 (2009).

- 687 43. J. Geml et al., J. Biogeogr. 34, 74 (2012).
- 688 44. I. Timling *et al.*, *Mol. Ecol.* 23, 3258 (2014).
- 689 45. J. Rousk *et al.*, *ISME J.* 4, 1340 (2010).
- 690 46. M. D. Crisp et al., Nature 458, 754 (2009).
- 691 47. S. T. Bates *et al.*, *ISME J.* 7, 652 (2013).
- 692 48. W. Jun Sul et al., Proc. Natl. Acad. Sci. USA 110, 2342 (2013).
- 693 49. H. Sato *et al.*, *Mol. Ecol.* **21**, 5599 (2012).
- 694 50. I. Sanmartin, F. Ronquist, Syst. Biol. 53, 216 (2004).
- 695 51. S. Altizer *et al. Science* **341**, 514 (2013).
- 696 52. W. H. van der Putten *et al.*, *Phil. Trans. R. Soc. B* 365, 2025 (2010).
- 697 53. M.-L. Desprez-Loustau et al., Trends Ecol. Evol. 22, 472 (2007).
- 698

699 Acknowledgements

- 700
- 701 The sequence data and metadata are deposited in the Short Read Archive (accession
- 702 SRP043706) and UNITE databases. Data used for analyses are available as supplementary
- online material Data S1 and S2. We thank H. Mann, D. Sveshnikov, F.O.P. Stefani, A. Voitk,
- and Y. Wu for supplying single soil samples; R. Puusepp, M. Haugas, and M. Nõukas for
- sample preparation; H. Kreft for providing interpolated plant diversity data; S. Jüris for
- designing the printed figure; M.I. Bidartondo, K.G. Peay and three anonymous reviewers for
- 707 constructive comments on the manuscript; and relevant institutions of multiple countries for
- issuing permissions for sampling and delivery. The bulk of this project was funded from
- Estonian Science Foundation grants 9286, 171PUT, IUT20-30; EMP265; FIBIR; ERC; and in

- 710 part by numerous funding sources that facilitated co-author efforts in collecting and pre-
- 711 processing samples.

714 Figure legends

Fig. 1. Map of A) global sampling (circles as study sites); B) Interpolated taxonomic richness
of all fungi using Inverse Distance Weighting (IDW) algorithm and accounting for the
relationship with mean annual precipitation (based on the best multiple regression model).
Different colors depict residual Operational Taxonomic Unit (OTU) richness of all fungi
accounting for sequencing depth. Warm colors indicate OTU-rich sites, whereas cold colors
indicate sites with fewer OTUs.





- **Fig. 2.** Relative proportion of fungal sequences assigned to major taxonomic groups in
- 725 different biomes.

729 Fig. 3. Relationships between residual richness of fungal taxonomic or functional groups and

- distance from the equator. A, all fungi; B, ectomycorrhizal (EcM) fungi; C, saprotrophic fungi;
- 731 D, plant pathogens; E, animal parasites; F, mycoparasites; G, white rot decomposers; and H,
- 732 yeasts. Lines indicate best-fitting linear or polynomial functions.



Fig. 4. Relationship between standardized plant richness to fungal richness ratio and distance
from the equator based on (A) interpolated values and (B) polynomial regression. Residuals of
fungal richness are taken from the best linear regression model accounting for other significant
predictors. Warm colors indicate high plant-to-fungal richness ratio, whereas cold colors
indicate low plant-to-fungal richness.



Fig. 5. Connectedness of biogeographic regions by shared Operational Taxonomic Units
(OTUs) of ectomycorrhizal fungi (blue), saprotrophs (black), and plant pathogens (red). The
width of lines and diameter of circles are proportional to the square root of the number of
connections and sample size (number of sites), respectively. Numbers in circles indicate the
number of OTUs found in each region. OTUs with a single sequence per site and OTUs
belonging to Hypocreales and Trichocomaceae (in which the ITS region is too conservative for
species-level discrimination) were excluded.



Fig. 6. Ward clustering of biogeographic regions (left panes) and biomes (right panes) based
on the Morisita-Horn pairwise similarity index in A and B, all fungi; C and D, ectomycorrhizal
fungi; E and F, saprotrophs; G and H, plant pathogens. Numbers above branches indicate Pvalues.



759 Supplementary Materials

760

761 Figs. S1-S16

762

763 Tables S1-S3

764

765 Data S1-S2